



Artificial induction and isolation of cadmium-tolerant soil bacteria

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Received: 9 January 2020 / Accepted: 21 March 2020 / Published Online: 30 June 2020
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Abstract Environmental pollution caused by various heavy metals is a serious global problem. To solve this problem, microbial bioremediation of contaminated metals has developed rapidly as an effective strategy when physical and chemical techniques are not suitable. In this study, cadmium (Cd)-tolerant soil bacteria were isolated via artificial induction in laboratory conditions instead of screening bacteria naturally adapted to metal-contaminated soils. Wild-type (WT) bacteria grown in uncontaminated soils were artificially and sequentially adapted to gradually increasing Cd concentrations of up to 15 mM. The resultant cells, named Soil-CdR15, survived at a Cd concentration of 10 mM, whereas WT cells failed to survive with 4 mM Cd on solid media for 2 d. In liquid media containing Cd, the Soil-CdR15 cells grew with 15 mM Cd for 7 d, whereas the WT cells could not grow with 5 mM Cd. Both Soil-CdR15 and WT cells removed approximately 35% of Cd at the same capacity from liquid media containing either 0.5 or 1.0 mM Cd over 2 d. In addition to Cd, the Soil-CdR15 cells showed increased resistance to nickel, zinc, and arsenic compared to WT cells. The Soil-CdR cells were identified as *Burkholderia* sp. by partial sequencing of 16S rRNA. The data presented in this study demonstrate that isolation of heavy metal-tolerant microorganisms via artificial induction in laboratory conditions is possible and may be useful for the application of the microorganisms for the bioremediation of heavy metals.

Keywords Bioremediation · Cadmium · Environment Heavy metal · Soil bacteria

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Introduction

Heavy metals have become one of the most prevalent and concerning environmental pollutants worldwide [1]. Non-essential heavy metals, including cadmium (Cd), lead, mercury, and chromium, are not needed for the growth of organisms and can be toxic at low concentration. They are non-degradable and, therefore, must be separated from living organisms. The typical methods for cleaning up toxic heavy metals from contaminated soils and waters involve physical and chemical methods [2]. These methods are effective when the contaminated area is small and the concentration of metals is high. However, they are expensive, can make the land infertile because of the removal of all essential elements during the clean-up procedure, and ineffective for large areas contaminated with low metal concentrations. Therefore, as an alternative to the typical methods, bioremediation, which uses microorganisms to clean up pollutants from a contaminated site, has been developed [3-7].

There has been rapid progress in techniques for microbial bioremediation of heavy metals [8-10]. Biosorption, as a bioremediation of metal removal in wastewaters, using either live or dead microorganisms has been suggested as effective and economic technique when the contaminated metal concentrations are not high [11]. Bioremediation of Cd was conducted by metal-resistant mutated bacteria isolated from industrial effluent [12]. *Saccharomyces cerevisiae* was used for removing Cd from aqueous waste solutions [9]. Therefore, one of recent research focus is improving the development of organisms that possess both, a higher tolerance to heavy metals and the ability to accumulate them in large amounts. In general, bacteria that are pre-exposed to heavy metals for a long time have a greater tolerance than those that are not pre-exposed because of the adaptation of the bacteria to metal stress [13]. The adaptation includes both physiological and genetic changes. Most of heavy metal-tolerant microorganisms have been isolated from areas polluted with heavy metals, and these organisms have reduced various metal contents in metal-polluted wastewaters [10, 14-18].

Cadmium-resistant mutants of *Aspergillus nidulans* were isolated by exposing to ultraviolet [19]. Copper-tolerant mutants of sake yeast were isolated by treatment of ethyl methanesulfonate [20]. And, the development of metal tolerance in soil communities was conducted in agricultural soils amended with different heavy metals in laboratory conditions [21]. However, there has been a lack of studies on whether heavy metal-tolerant microorganisms can be generated artificially by exposing them to heavy metal stress in laboratory conditions.

Most heavy metals, including Cd, have mutagenic effects [22, 23]. Thus, the long-exposure of microorganisms to heavy metals in artificial conditions can induce metal tolerance by inducing mutations during the adaptation process. Therefore, this study focused on the artificial generation of bacterial mutants that have a strong tolerance to Cd, a representative non-essential heavy metal, using the sequential adaptation of bacteria to gradually increases in Cd stress.

Materials and Methods

Induction and isolation of Cd-resistant bacteria

A soil sample was collected from a rose garden at Kyungpook National University, Korea. Approximately 5 g soil was mixed with 10 mL Luria-Bertani (LB) medium, and 100 μ L of the supernatant was inoculated in 5 mL LB medium. The cells were incubated for 24 h at 30 °C with shaking at 200 rpm. Then, 100 μ L of the grown cells was inoculated in 5 mL LB medium containing 100 μ M CdCl₂ and incubated until cell growth reached a level close to saturation. If the cells grew well within 48 h, then the experiment was repeated by inoculating 100 μ L of the grown cells in new media in which the Cd concentration was gradually increased in 100 μ M units up to 1000 μ M. After this point, the Cd concentrations were gradually increased in 1 mM units. If the inoculated cells did not grow well in the medium with the higher concentration of Cd within 48 h, the cells grown in the previous Cd concentration were inoculated in medium with a Cd concentration that was either the same or lower than the previous concentration of Cd. The Cd-resistant soil bacteria (named Soil-CdR15) were isolated by growing cells in media containing Cd and gradually increasing the concentration of CdCl₂ to 15 mM, and this procedure took approximately 3 months.

Identification of Soil-CdR15 based on 16S rRNA

The genomic DNA of Soil-CdR15 cells was extracted using a MagListo™ 5M Genomic DNA Extraction Kit (Bioneer, Daejeon, Korea), and the 16S rDNA was amplified via polymerase chain reaction (PCR). The universal primers 27F (5'-AGA GTT TGA TCC TGG CTC AG-3') and 1492R (5'-GGC TAC CTT GTT ACG ACT T-3') were used. The PCR reaction conditions were as follows: an initial denaturation at 95 °C for 5 min; 30 cycles of denaturation at 95 °C for 30 s, primer annealing at 56 °C for 30 s,

and extension at 72 °C for 90 s; a final extension at 72 °C for 10 min; and cooling of the amplified products to 4 °C. The amplified PCR products were analyzed by 0.8% (w/v) agarose gel electrophoresis and purified using a Gel Extraction Kit (Invitrogen, Carlsbad, CA, USA). The purified DNA fragments were sent to Solgent (Daejeon, Korea) for sequencing.

Assay for metal sensitivity

First, 50 μ L of wild-type (WT) or Soil-CdR15 cells was inoculated in 5 mL LB media containing various concentrations of metals (CdCl₂, NiCl₂, ZnCl₂, and Na₂HAsO₄ obtained from Sigma-Aldrich, St. Louis, MO, USA) and incubated for 24 h, or another indicated time, at 30 °C with shaking at 200 rpm. Cell densities were then measured using spectrophotometry at 600 nm. WT and Soil-CdR15 cells (10 μ L) were grown for 48 h at 30 °C on LB agar media containing the indicated Cd concentrations and then imaged.

Analysis of Cd removal

After preparation of 15 mL LB medium in a flask containing either 0.5 or 1.0 mM CdCl₂, 5 mL of the medium was each aliquoted into two flasks. Then, 100 μ L of WT or Soil-CdR cells was inoculated in the media and incubated for 3 d at 30 °C with shaking at 200 rpm. These cells were spun down by centrifuging at 16,800 \times g for 10 min, and this was repeated three times to remove the cells. Total Cd concentrations of the media were determined using inductively coupled plasma atomic emission spectrometry (ICP-AES; Optima 7300DV, PerkinElmer, Waltham, MA, USA) via serial dilution of sample. Multi-element standard solution V (Merck Chemicals, Darmstadt, Germany) was used as reference standard. The ICP-AES was calibrated for Cd by running different concentrations of standard solutions (0.1, 1, and 10 mg/L).

Statistical analysis

Data (Figs. 3, 4) are presented as the mean \pm standard deviation (SD) of three independent experiments. A one-way analysis of variance (ANOVA) was carried out to measure significant differences among the groups at the significant level of 5%, $p < 0.05$ using a SPSS program with a 23.0 version.

Results and Discussion

Cd-tolerance analysis of Soil-CdR15 bacteria

The isolated Cd-resistant Soil-CdR15 cells showed a higher tolerance to Cd than WT cells (Fig. 1). The growth of WT cells was inhibited significantly on solid media containing 2 mM Cd, and only a few WT colonies appeared when the Cd concentration was 3 mM, which was possibly due to the survival of a few types of soil bacteria that had an inherent tolerance to that Cd concentration. When the Cd concentration was higher than 5 mM,

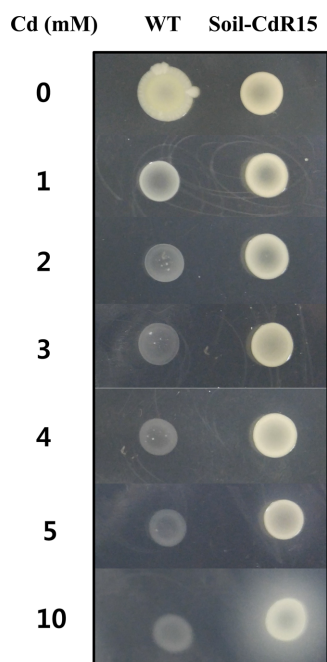


Fig. 1 Cd sensitivities of WT and soil-CdR15 cells. WT and soil-CdR15 cells were grown for 48 h at 30 °C on agar media containing the indicated Cd concentrations, and images were then taken. WT, wild-type soil bacteria; soil-CdR15, soil bacteria adapted to 15 mM Cd stress

the WT cells could not survive. In contrast, the Soil-CdR15 cells showed no significant growth retardation until the Cd concentration reached 5 mM and survived even with 10 mM Cd. Cd-tolerance analysis in liquid media also showed a higher tolerance of Soil-CdR15 cells compared with WT cells (Fig. 2). The growth of WT cells was almost inhibited with 5 mM Cd, whereas the Soil-CdR cells were inhibited with 12.5 mM Cd for 1 d. After a 3 d incubation time, WT cells showed no growth with 5 mM Cd, whereas the Soil-CdR cells grew with 15 mM Cd. After a 7 d incubation time, WT cells did not grow with 5 mM Cd, whereas the Soil-CdR15 cells survived with 15 mM Cd.

Analysis of Cd removal from media

The Cd removal capacity of WT and Soil-CdR15 cells was compared by growing both cells in liquid media containing either 0.5 mM or 1.0 mM Cd for 3 d (Fig. 3). There was no significant difference in Cd removal capacity, and both cell types removed approximately 35% of the Cd present in the liquid media after 3 d. Even though the Soil-CdR15 cells did not exhibit a higher Cd removal capacity compared to WT cells, they can be used for Cd removal at Cd concentrations at which WT cells cannot survive. Moreover, it is difficult to obtain yeast species that are both hyper-accumulative and tolerant to heavy metals [24]. Therefore, this study focused on acquiring heavy metal-tolerant microorganisms which may be useful for application in bioremediation of highly Cd contaminated waste waters.

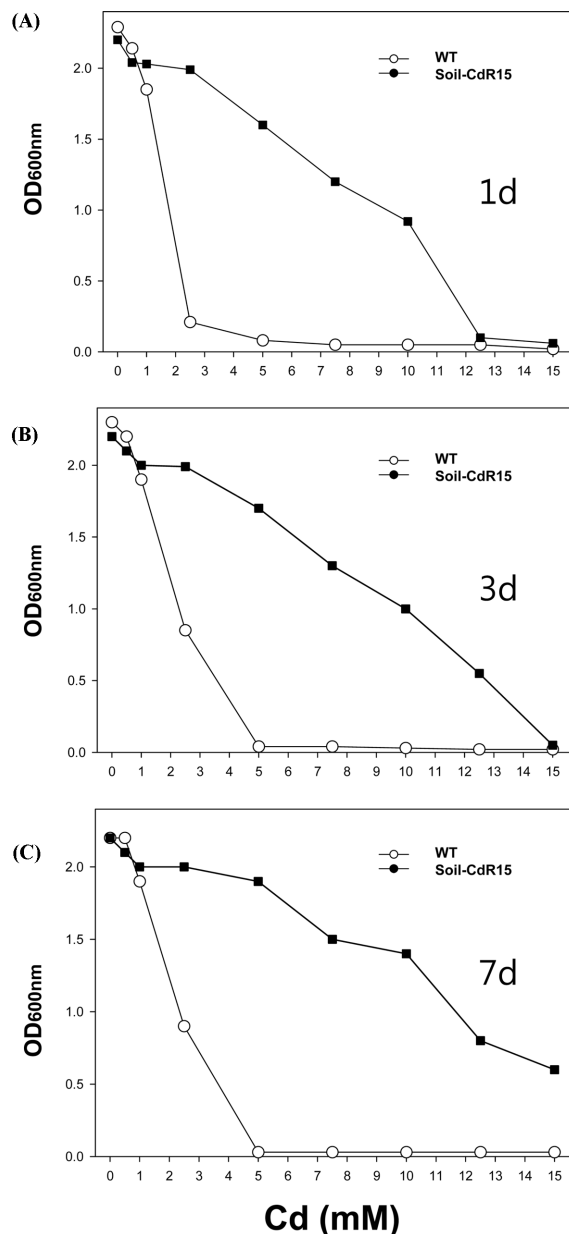


Fig. 2 Survival analysis of WT and soil-CdR15 cells over time to high levels of Cd. WT and soil-CdR15 cells were grown in liquid media containing various concentrations of Cd for the indicated time at 30 °C. Cell densities were then measured at 600 nm after 1 d (top), 3 d (middle), and 7 d (bottom) using a spectrophotometer. WT, wild-type soil bacteria; soil-CdR15, soil bacteria adapted to 15 mM Cd stress

Tolerance analysis of Soil-CdR15 to Ni, As, and Zn

Occasionally, microorganisms that have adapted to one type of metal stress acquire tolerance to other metals [18,25]. Soil bacterial communities became tolerant to metals other than the metals added to soils [14]. Hence, the sensitivity of the Soil-CdR15 cells to other metals was analyzed. Compared to WT cells, the Soil-CdR15 cells showed an increased tolerance to Ni (Fig. 4A), Zn

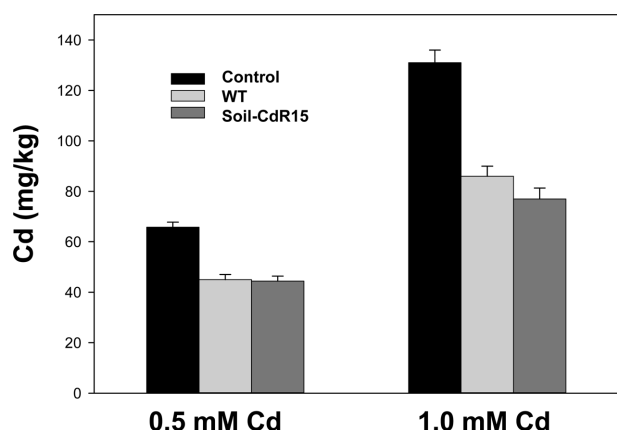


Fig. 3 Cd-removal analysis of WT and soil-CdR15 cells. WT and soil-CdR15 cells were grown in liquid media containing either (A) 0.5 or (B) 1.0 mM Cd for 3 d at 30 °C. Then, the Cd concentrations in the media were measured. Values shown are the mean \pm SD (standard deviation) of three replicates. Control, medium without inoculation; WT, wild-type soil bacteria; soil-CdR15, soil bacteria adapted to 15 mM Cd stress

(Fig. 4B), and As (Fig. 4C). The Soil-CdR15 cells showed ~20% and WT cells showed ~90% inhibition of growth in liquid media containing 2 mM Ni. For Zn stress, WT cells showed ~85% inhibition of growth at 10 mM Zn, whereas the Soil-CdR15 cells did not show significant inhibition of growth until 30 mM Zn. For As stress, the Soil-CdR15 cells showed ~10% and WT cells showed ~55% inhibition of growth at 30 mM As.

Identification of Soil-CdR15 by sequencing of 16S rRNA

The Soil-CdR15 cells may be either a single cell type or mixture of several types of cells. To identify the cells, 10 single colonies were prepared by streaking them on solid media. Then, metal tolerance was analyzed by growing them in liquid media as described in Fig. 2 and Fig. 4. The analysis indicated that all colonies looked similar (data not shown). Then, 16S rRNA from all the 10 colonies was sequenced. All sequencing results indicated that the Soil-CdR15 cells are *Burkholderia* sp.

In conclusion, the results of this study suggest that heavy metal-tolerant microorganisms can be artificially developed in laboratory conditions by adapting naturally grown microorganisms to higher concentrations of metal stress. The results presented here will be useful for the development of an effective method for the bioremediation of heavy metals using microorganisms.

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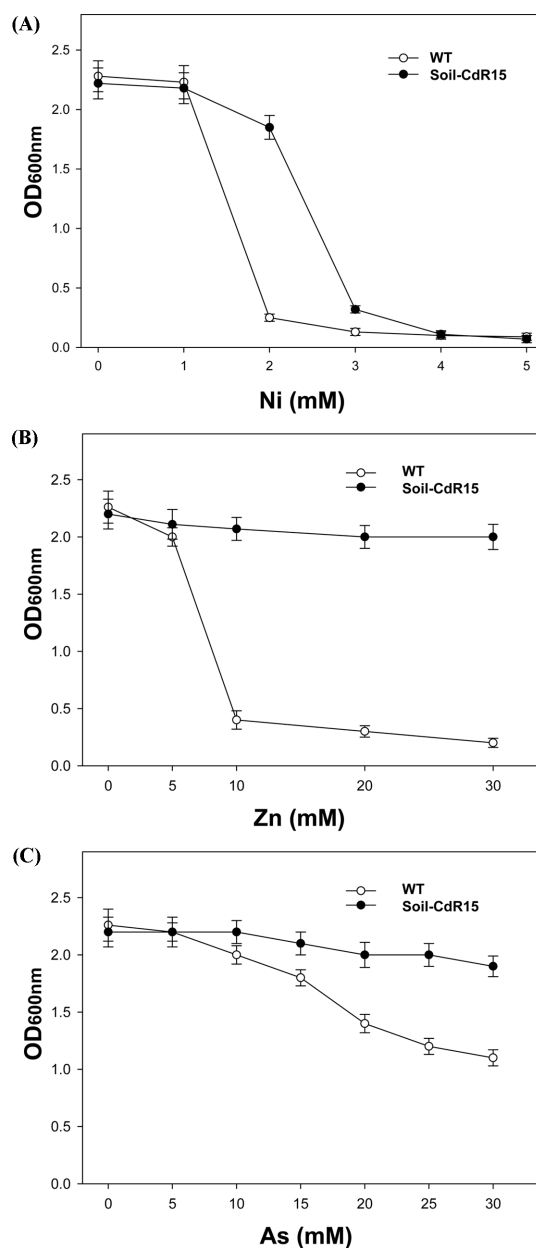


Fig. 4 Sensitivity of WT and soil-CdR15 cells to Ni, Zn, and As. WT and soil-CdR15 cells were grown in liquid media containing various concentrations of (A) Ni, (B) Zn, and (C) As for 24 h at 30 °C. Cell densities were then measured at 600 nm using a spectrophotometer. Values shown are the mean \pm SD (standard deviation) of three replicates. WT, wild-type soil bacteria; soil-CdR15, soil bacteria adapted to 15 mM Cd stress

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