



# Assessment of Environmental Radioactivity Surveillance Results around Korean Nuclear Power Utilization Facilities in 2017

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## ABSTRACT

**Background:** Government conducts environmental radioactivity surveillance for verification purpose around nuclear facilities based on the Nuclear Safety Law and issues a surveillance report every year. This study aims to evaluate the short and the long-term fluctuation of radionuclides detected above MDC and their origins using concentration ratios between these radionuclides.

**Materials and Methods:** Sample media for verification surveillance are air, rainwater, groundwater, soil, and milk for terrestrial samples, and seawater, marine sediment, fish, and seaweed for marine samples. Gamma-emitting radionuclides including  $^{137}\text{Cs}$ ,  $^{90}\text{Sr}$ , Pu,  $^3\text{H}$ , and  $^{14}\text{C}$  are evaluated in these samples.

**Results and Discussion:** According to the result of the environmental radioactivity verification surveillance in the vicinity of nuclear power facilities in 2017, the anthropogenic radionuclides were not detected in most of the environmental samples except for the detection of a trace level of  $^{137}\text{Cs}$ ,  $^{90}\text{Sr}$ , Pu, and  $^{131}\text{I}$  in some samples. Radioactivity concentration ratios between the anthropogenic radionuclides ( $^{137}\text{Cs}/^{90}\text{Sr}$ ,  $^{137}\text{Cs}/^{239+240}\text{Pu}$ ,  $^{90}\text{Sr}/^{239+240}\text{Pu}$ ) were similar to those reported in the environmental samples, which were affected by the global fallout of the past nuclear weapon test, and Pu atomic ratios ( $^{240}\text{Pu}/^{239}\text{Pu}$ ) in the terrestrial sample and marine sample showed significant differences due to the different input pathway and the Pu source. Radioactive iodine ( $^{131}\text{I}$ ) was detected at the range of  $< 5.6\text{--}190 \text{ mBq} \cdot \text{kg-fresh}^{-1}$  in the gulfweed and sea trumpet collected from the area of Kori and Wolsong intake and discharge. A high level of  $^3\text{H}$  was observed in the air (Sangbong:  $0.688 \pm 0.841 \text{ Bq} \cdot \text{m}^{-3}$ ) and the precipitation (Meteorology Post:  $199 \pm 126 \text{ Bq} \cdot \text{L}^{-1}$ ) samples of the Wolsong nuclear power plant (NPP).  $^3\text{H}$  concentration in the precipitation and pine needle samples showed typical variation pattern with the distance and the wind direction from the stack due to the gaseous release of  $^3\text{H}$  in Wolsong NPP.

**Conclusion:** Except for the detection of a trace level of  $^{137}\text{Cs}$ ,  $^{90}\text{Sr}$ , Pu, and  $^{131}\text{I}$  in some samples, anthropogenic radionuclides were below MDC in most of the environmental samples. Overall, no unusual radionuclides and abnormal concentration were detected in the 2017's surveillance result for verification. This research will be available in the assessment of environment around nuclear facilities in the event of radioactive material release.

**Keywords:** Environmental Radioactivity, Anthropogenic Radionuclides, Radioactivity Concentration, Normal Fluctuating Range, Radioactivity Concentration Ratio

## Introduction

The IAEA provides general requirements for the radiation monitoring and design of monitoring programs to protect the environment and human beings based on the

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Safety Standards Series [1]. Further, IAEA Safety Guide No. RS-G-1.8 [2] offers the responsibilities and generic aspects of surveillance programs of the facility operators and regulatory body; IAEA Safety Reports Series No. 64 [3] provides detailed descriptions of the actual design and operation of the environmental monitoring program that comprised monitored materials, radionuclides, and frequencies.

In South Korea, the Nuclear Safety Act, Article 104 (Preservation of Environment) stipulates that the installer or operator of facilities shall conduct the survey of radiation environment and the evaluation of the environmental impact for the surroundings of the facilities; Article 104, Clause 2 stipulates that a regulatory body will perform the verification surveillance to confirm and evaluate the surveillance results conducted by the facility operator. For this, the Nuclear Safety and Security Commission (NSSC) entrusted the Korea Institute of Nuclear Safety (KINS) to perform the surveillance of the surroundings of nuclear power utilization facilities in South Korea. A total of seven nuclear power utilization facilities in South Korea including four nuclear power plant sites in Kori, Wolsong, Hanul, and Hanbit, the nuclear fuel fabrication and nuclear power research facility in Daeduk, and the radioactive waste disposal facility in Gyeongju. The environment media, radionuclides, measurement frequency, and assessment for the verification surveillance are designed based on the NSSC Notification No. 2017-17 (the Regulation on Survey of Radiation Environment and Assessment of Radiological Impact on Environment in Vicinity of Nuclear Power Utilization Facilities) [4].

In the US and the UK, the government does not conduct verification surveillance; instead, it reviews and discloses the results of the assessments conducted by the operator. In Canada, the Radiation Protection Bureau of Health Canada [5] prepares the annual reports on the surveillance results, and in China, the Radiation Monitoring Technical Center prepares the annual reports on the surveillance results conducted by the Department of Environmental Protection in each province where power plants are located [6]. South Korea began the environmental radiation surveillance around the Gori nuclear power plant Unit 1 for the first time in 1977, and since 1983, detailed, government-led surveillance has been performed. Along with the expansion of nuclear power utilization facilities, the surveillance scope targets have gradually increased.

The government's surveillance results are open to the public as annual reports; these reports include the verifica-

tion of the reliability of the operator's surveillance results as compared with two results and the assessment of environmental impact by the facility operation based on the surveillance results. However, these reports were limited in the in-depth interpretation of the data at the research level, such as the long-term accumulation trends of each radionuclide, short-term fluctuation, and the relationships between radionuclide using the radioactivity concentration.

Therefore, this research used only detected radionuclides in the environmental radioactivity in the annual report on "The Environmental Radiological Surveillance Assessment around the Nuclear Facilities in 2017" [7] published by KINS, and studied for the short-term and long-term fluctuations of these radionuclides and the comparison with previous year's results, and the origin of each artificial radionuclides using the radioactivity concentration ratio between them. This research will be available in the evaluation of the environmental impact of the surroundings of nuclear power utilization facilities at nuclear or radiation emergency situations.

## Materials and Methods

Table 1 shows the environmental radioactivity monitoring

**Table 1.** Environmental Radioactivity Monitoring Program around Nuclear Power Utilization Facilities

Monitored constituents		Analysis Items	Frequency
Air, deposition	Soil	Gamma RN	Twice a year
		$^{90}\text{Sr}$ , $^{239}\text{Pu}$ , $^{240}\text{Pu}$	Yearly
		$^{234}\text{U}$ , $^{235}\text{U}$ , $^{238}\text{U}$	Yearly
	Sea and stream sediments	Gamma RN	Twice a year
		$^{90}\text{Sr}$ , $^{239}\text{Pu}$ , $^{240}\text{Pu}$	Yearly
		$^{234}\text{U}$ , $^{235}\text{U}$ , $^{238}\text{U}$	Yearly
Air	$^3\text{H}$ , $^{14}\text{C}$	Monthly	
Pine Needle	$^3\text{H}$ , $^{14}\text{C}$	Monthly	
Water bodies	Seawater	Gamma RN, $^3\text{H}$	Quarterly
		$^{90}\text{Sr}$ , $^{239}\text{Pu}$ , $^{240}\text{Pu}$	Twice a year
	Underground water	Gamma RN, $^3\text{H}$ , $^{90}\text{Sr}$ , U, Pu isotopes, $^{99}\text{Tc}$	Twice a year
	Surface water Precipitation	Gamma RN Gamma RN $^3\text{H}$	Quarterly Monthly
Foodstuffs	Milk	Gamma RN	Quarterly
		Gamma RN	Monthly
		$^{90}\text{Sr}$	Twice a year
		$^3\text{H}$ , $^{14}\text{C}$	Monthly
	Leafy Vegetables	Gamma RN	Yearly
	Rice	Gamma RN	Yearly
Marine Samples	Fishes	Gamma RN	Twice a year
	Seaweeds	Gamma RN	Twice a year

program around nuclear power utilization facilities, and the sampling points are described in detail in the annual report published by KINS [7]. The targets are generally divided into land samples and marine samples. Land samples include air, water, soil, and agricultural product, and marine samples include seawater, seabed sediment, fish, and seaweed. The sampling points, preprocessing, and chemical separation methods of samples are described in detail in the reports; this study briefly describes the measuring samples preparation and measurement methods per radionuclide.

- 1) Gamma-radionuclide analysis: Soil, riverbed, and seabed sediments were dried, ground, and filled into the measurement container (D6H4) made by KINS. For seawater and underground water,  $\gamma$ -radionuclides were isolated from them using AMP (ammonium phosphomolybdate) and  $\text{MnO}_2$  coprecipitation method and filled into D6H4. The agricultural product samples were turned to ash in an electronic furnace at  $450^\circ\text{C}$ , and the ash was filled into D6H4. The precise  $\gamma$ -radionuclide analysis was performed using high purity germanium (HPGe)  $\gamma$ -radionuclide spectrometry and a multichannel analyzer.
- 2) Radiostrontium ( $^{90}\text{Sr}$ ) analysis: From the dried inorganic samples and ash sample, metallic elements were leached using strong acid treatment, and chemical and ion-column separation method were used to isolate Sr. For marine samples, the coprecipitation method with carbonates and oxalate was used, and Sr was purified using fuming nitric acid. The purified Sr in the form of strontium nitrate ( $\text{Sr}(\text{NO}_3)_2$ ) crystal that were melted with ultrapure water and stand for 14 days to make equilibrium between  $^{90}\text{Sr}$  and  $^{90}\text{Y}$ , and an ultra low-level liquid scintillation counter (LSC, LKB, Quantulus 1220) was used to measure the Cherenkov radiation from  $^{90}\text{Sr}$ .
- 3) Plutonium (Pu) analysis: Using acid, metallic elements were extracted from the small amount of samples (inorganic sample: 4 g, ash sample: 10 g), and Pu was purified using the TEVA (EiChrom) extraction chromatography. The amount of seawater sample (~20 kg) was reduced via  $\text{Fe}(\text{OH})_3$  coprecipitation and Pu was purified using the same method as the inorganic sample. The isolated Pu was further purified using extraction chromatography to minimize the effect of interference elements. The final Pu solution was measured using the multi-collector inductively coupled plasma mass spectrometry (MC-ICPMS, NEPTUNE, Thermo Scientific, Germany) through the sample introduction system including the desolvation equipment (ARIDUS-II, Cetac Technologies, USA) inside.
- 4) Tritium ( $^3\text{H}$ ) analysis: Tritium in the air was collected for one month with the automatic air sampling device having the molecular sieve (5 Å) column inside. The column was heated for over two hours at  $450^\circ\text{C}$  in the horizontal tube furnace, and the recovered tritium was moved to the distillation process to remove organic materials in the solution. Tritium was then put in the Teflon measurement container, and the scintillator (Ultima gold LLT, Packard, USA) was added and measured with an ultra low-level LSC. For pine needles and milk samples, a freeze dryer was used to extract the tissue free water, and the dried samples were used to extract the organic bound water. Water was directly distilled and measured in the same way as the air sample.
- 5) Radiocarbon ( $^{14}\text{C}$ ) analysis:  $\text{CO}_2$  in the air was collected for one month along with tritium in the same air sampling device. Using the horizontal tube furnace, the  $\text{CO}_2$  collected in the molecular sieve (5 Å) column inside the collection device was heated for over two hours at  $450^\circ\text{C}$ , and the extracted  $\text{CO}_2$  was dissolved into the ammonia solution ( $\text{NH}_4\text{OH}$ ). Calcium chloride ( $\text{CaCl}_2$ ) was added to the ammonia solution to synthesize calcium carbonate ( $\text{CaCO}_3$ ) precipitate. For pine needles and milk samples, free water was completely removed using a freeze dryer, and using a high-pressure bomb,  $\text{CO}_2$  was collected from the dried samples and  $\text{CaCO}_3$  was synthesized using the same method as the air sample. The retrieved  $\text{CaCO}_3$  was mixed very slowly with HCl to evolve  $\text{CO}_2$  from the  $\text{CaCO}_3$ , and  $\text{CO}_2$  was trapped into absorber (CarboSorb E, Packard Bioscience B.V.) and mixed with a scintillator (PermaFluor E+, Packard Bioscience B.V.) for measurement using LSC.

## Results and Discussion

### 1. Cesium-137 ( $^{137}\text{Cs}$ ) results

The 2017 results showed that radioactive cesium ( $^{137}\text{Cs}$ ) and radioactive iodine ( $^{131}\text{I}$ ) among the  $\gamma$ -radionuclides were detected in some samples, and the other artificial  $\gamma$ -radionuclides shown in Appendix 1 of the Notification of NSSC [4] were not detected. The detection percentage of  $^{137}\text{Cs}$  in soil, seabed sediment, fish, and seawater samples was 32%, 13%, 90%, and 98%, respectively, and  $^{131}\text{I}$  was detected in few seaweed

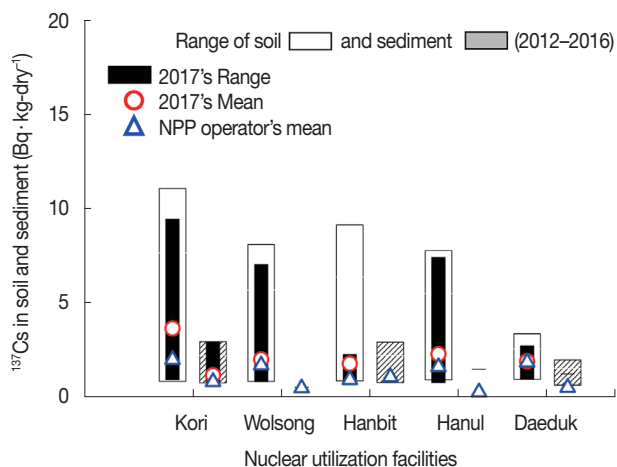


Fig. 1. Comparison of  $^{137}\text{Cs}$  concentration in soil and sediment samples with previous year's data and NPP operator's mean.

samples. The detection percentage in the seabed sediment was lower than that in the soil sample because the seabed sediments collected at the nuclear power plants' (NPP) intake and discharge area, except in Kori, had a higher percentage of sand than in soil and accordingly, the content of the organic materials and clay that contribute to  $^{137}\text{Cs}$  absorption is low. Since the detection of these radionuclides depends on their minimum detectable radioactivity concentration (MDC) that relies on the sample amounts in the analysis and the equipment capability in the measurement, the detection itself does not imply the existence of the radionuclides in the samples. The MDC of soil, seabed sediment, fish, and seawater was  $<0.80$ – $<2.58$   $\text{Bq}\cdot\text{kg}\cdot\text{dry}^{-1}$ ,  $<0.56$ – $<1.91$   $\text{Bq}\cdot\text{kg}\cdot\text{dry}^{-1}$ ,  $<3.72$ – $<65.2$   $\text{mBq}\cdot\text{kg}\cdot\text{fresh}^{-1}$ , and  $<1.36$ – $1.69$   $\text{mBq}\cdot\text{kg}^{-1}$ , respectively.

Shown in Figure 1 is the  $^{137}\text{Cs}$  analysis results in the soil and seabed (riverbed) sediments, and the soil results in 2017 ( $<0.80$ – $9.47$   $\text{Bq}\cdot\text{kg}\cdot\text{dry}^{-1}$ ) were within the normal fluctuation of  $0.89$ – $11.1$   $\text{Bq}\cdot\text{kg}\cdot\text{dry}^{-1}$  for the last five years without any drastic change in the surrounding regions of the five NPPs. This was similar to the concentration range ( $<0.65$ – $20.1$  (Jeju)  $\text{Bq}\cdot\text{kg}\cdot\text{dry}^{-1}$ ) in the soil (top layer) at the 15 regional radioactivity monitoring stations in 2017 [8].  $^{137}\text{Cs}$  in the seabed sediments was detected in only Kori, including the New Kori, and the range of the concentration was very low with  $<0.67$ – $2.95$   $\text{Bq}\cdot\text{kg}\cdot\text{dry}^{-1}$  that was within the concentration variation in the last five years at  $0.92$ – $2.50$   $\text{Bq}\cdot\text{kg}\cdot\text{dry}^{-1}$ .

Meanwhile,  $^{137}\text{Cs}$  was detected above MDC ( $\sim 1$   $\text{mBq}\cdot\text{kg}^{-1}$ ) in most seawaters and the concentration was within  $1.36$ –

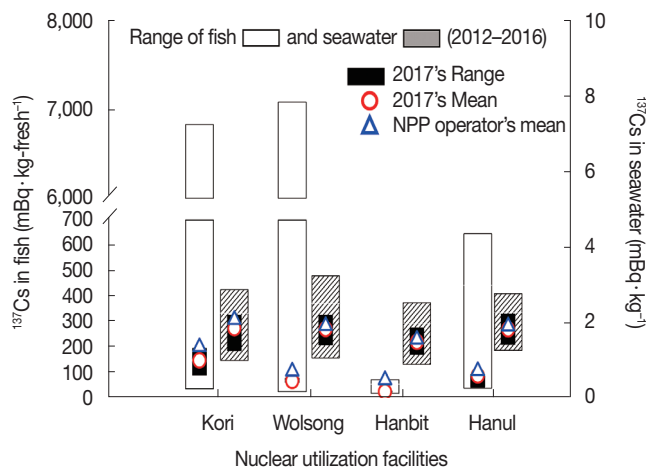


Fig. 2. Comparison of  $^{137}\text{Cs}$  concentration in fish and seawater samples with previous year's data and NPP operator's mean.

$2.23$   $\text{mBq}\cdot\text{kg}^{-1}$ . The low MDC obtained by applying a large number of samples (approximately 60 kg) and the long measurement time (80,000 seconds). Owing to the homogeneity characteristics of the seawater sample, there was little variation in the concentration between the sampling point, and the results in 2017 were similar to those of the previous years. However, the results of the six species of fish (a total of 28 samples) were ranged from  $<22.2$  to  $196$  (Spanish mackerel)  $\text{mBq}\cdot\text{kg}\cdot\text{fresh}^{-1}$  (Figure 2), but that in all seaweeds were below MDA ( $37.2$ – $97.6$   $\text{mBq}\cdot\text{kg}\cdot\text{fresh}^{-1}$ ). The mean enrichment coefficient calculated by the average radioactivity concentration in the seawater and fish collected around NPPs ( $n=23$ ) was 59, and it was ranged between 9 and 109. This result was within the enrichment coefficient (mean = 39; range 6–105;  $n=90$ ) [9] presented in the annual report of the Korean marine environments surveillance for five years (2005–2009) and the reported value (mean = 84; range 5–1,800;  $n=1,812$ ) by the IAEA [10].

## 2. Strontium-90 ( $^{90}\text{Sr}$ ) results

As shown in Figure 3, the  $^{90}\text{Sr}$  concentration range in the soil in 2017 was  $<0.17$ – $0.94$  (mean:  $0.51$ )  $\text{Bq}\cdot\text{kg}\cdot\text{dry}^{-1}$ , which was within the last five years' concentration range of  $0.11$ – $2.19$   $\text{Bq}\cdot\text{kg}\cdot\text{dry}^{-1}$ , and the detection percentage above the MDC ( $0.17$   $\text{Bq}\cdot\text{kg}\cdot\text{dry}^{-1}$ ) was approximately 76%. The concentration range in the seabed sediments was  $<0.10$ – $0.32$  (mean:  $0.23$ )  $\text{Bq}\cdot\text{kg}\cdot\text{dry}^{-1}$ , which was about one-third of that of the soil sample. It is assumed that such a difference in concentration was because of the physio-chemical properties of the soil and seabed sediments as in the  $^{137}\text{Cs}$  results. In

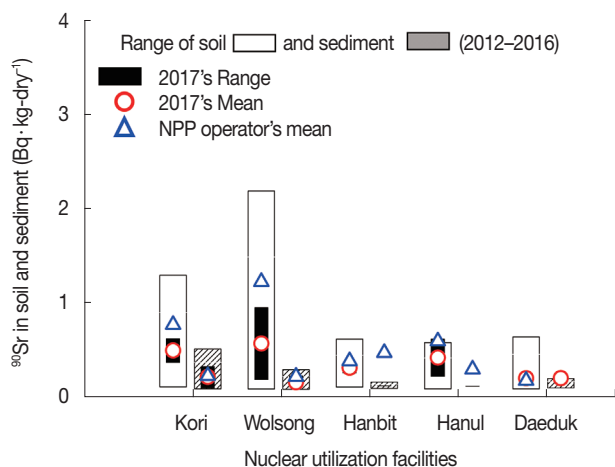


Fig. 3. Comparison of  $^{90}\text{Sr}$  concentration in soil and sediment samples with previous year's data and NPP operator's mean.

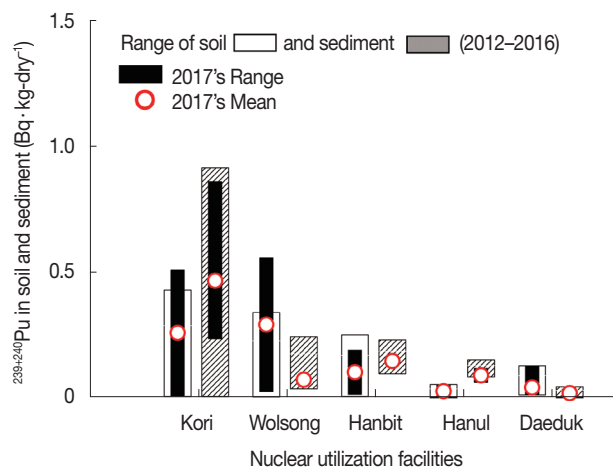


Fig. 4. Comparison of  $^{239+240}\text{Pu}$  concentration in soil and sediment samples with previous year's data.

Table 2.  $^{240}\text{Pu}/^{239}\text{Pu}$  Atom Ratios in the Marine and Terrestrial Samples around Facilities

Site (Sample no.)	Seawater		Sediment (Sea/stream)		Soil	
	Mean	Range	Mean	Range	Mean	Range
Kori (9)	$0.239 \pm 0.007$	0.228–0.253	$0.243 \pm 0.004$	0.235–0.249	$0.181 \pm 0.002$	0.179–0.182
Wolsong (5)	$0.241 \pm 0.008$	0.231–0.255	$0.244 \pm 0.006$	0.234–0.250	$0.174 \pm 0.005$	0.170–0.177
Hanbit (2)	$0.220 \pm 0.017$	0.194–0.243	$0.209 \pm 0.000$	0.209	$0.171 \pm 0.001$	0.170–0.172
Hanul (2)	$0.233 \pm 0.004$	0.226–0.236	$0.240 \pm 0.003$	0.238–0.242	$0.181 \pm 0.003$	0.179–0.183
Daeduck* (2)	-	-	$0.178 \pm 0.001$	0.177–0.179	$0.182 \pm 0.001$	0.181–0.182
Gyeongju† (1)	-	-	$0.234 \pm 0.007$	-	$0.184 \pm 0.020$	-

\*Stream sediment.

†Low and intermediate level radioactive waste disposal facility.

the seawater samples collected from the intake and discharge of the five NPPs,  $^{90}\text{Sr}$  was above MDC ( $\sim 0.4 \text{ mBq} \cdot \text{kg}^{-1}$ ), and the concentration range was  $0.58\text{--}1.32 \text{ mBq} \cdot \text{kg}^{-1}$ , which was similar to the results ( $0.64\text{--}1.37 \text{ mBq} \cdot \text{kg}^{-1}$ ) obtained in the 2017 Korean marine survey [11]. There was no large difference in the concentration between plant sites, and the  $^{90}\text{Sr}$  concentration in 2017 was within the normal fluctuation.

### 3. Plutonium results

Plutonium was detected in all soil, seabed (riverbed) sediment, and seawater samples; this was because of the considerably low MDC owing to the use of MC-ICPMS that offers extremely high measurement sensitivity. MDC in the soil and seabed sediment was  $\sim 5 \text{ mBq} \cdot \text{kg}^{-1}$  and that in the seawater was  $\sim 0.1 \text{ } \mu\text{Bq} \cdot \text{kg}^{-1}$ . In addition, by detecting  $^{239}\text{Pu}$  and  $^{240}\text{Pu}$  using MC-ICPMS, the atom ratio between them could be determined, and thereby, Pu origin could be estimated. In 2017, the average concentration in the soil and seabed (riverbed) sediment samples at the four NPPs was all within

the normal fluctuation level, and except in Kori Power Plant, the concentration in the soil was higher than that in the seabed sediments (Figure 4). The seabed sediments collected from the intake area of Kori 1 and Kori 2 NPP showed a higher value than the soil sample because of the absorption effect caused by the higher clay content than any other region and the sinking effect of Pu in the seawater. The average concentration of the Pu ( $^{239+240}\text{Pu}$ ) from the total 42 seawater samples was  $5.2 \pm 2.3 \text{ } \mu\text{Bq} \cdot \text{kg}^{-1}$ , which was within the fluctuation of the last five years and was similar to the result obtained through the marine environmental radioactivity survey [11].

For soil and riverbed sediments, the  $^{240}\text{Pu}/^{239}\text{Pu}$  atom ratio (Table 2) is similar to the atom ratio at 0.18 to 0.19 because of the global fallout caused by a nuclear test [12]. The Pu in the riverbed sediments in Daeduk originated from the atmosphere fallout, as opposed to the Pu in the marine samples, and therefore it was similar to the soil samples. Meanwhile, Pu atom ratios in the marine samples (seabed sediments and seawater) were over 0.22, except in the seabed sedi-



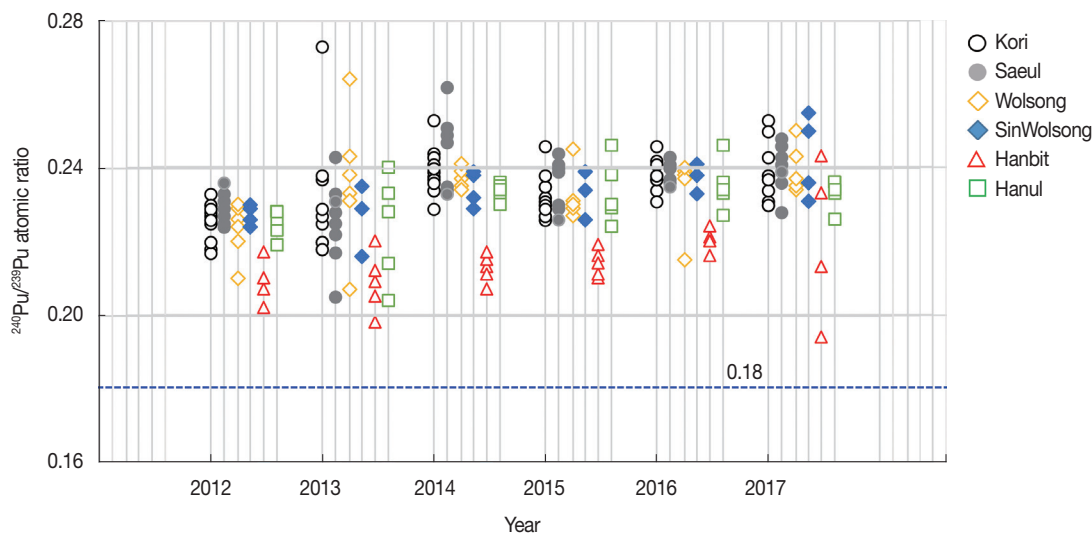


Fig. 5. Pu atom ratios variation with year in seawater samples collected around 6 nuclear facilities.

ments in Hanbit NPP, and such a high atom ratio was similar to that in the seawater in the Pacific region and seabed sediments around the Korean Peninsula in the past [13, 14]. It has been reported that the high Pu atom ratio in the marine samples was because of the contribution of close-in fallout Pu from nuclear weapons tests conducted at the Pacific Proving Grounds that had a different fission yield than the inland nuclear weapons tests, directly to the sea, and thus, it caused different ratio from the global fallout in the land [15]. Owing to the effect of the marine environments, the seabed sediments at the intake and discharge area showed a high atom ratio. However, in the case of the seabed sediments from the Hanbit NPP, the average atom ratio was 0.21, meaning that the sediments had high clay content, most of which are mixed with the land environmental substance. Such regional characteristics had also been observed in the atom ratio results (Figure 5) in seawater in 2017 and for the last five years. The seawater collected at the Hanbit intake and discharge had a high content of fine clay particles, and because of the effect of some Pu eluted from it, its atom ratio was lower than the other seawater samples.

#### 4. Iodine-131 ( $^{131}\text{I}$ ) results

The 2017 survey results showed that  $^{131}\text{I}$  was detected in some seaweed samples collected around the power plants' intake and discharge, and MDC was  $<55 - <176 \text{ mBq} \cdot \text{kg-fresh}^{-1}$ . It was detected only in the gulfweed and sea trumpet collected in the sea near Kori and Wolsong, and it was not detected in Hanbit (gulfweed) or Uljin (sea mustard, gulfweed). The re-

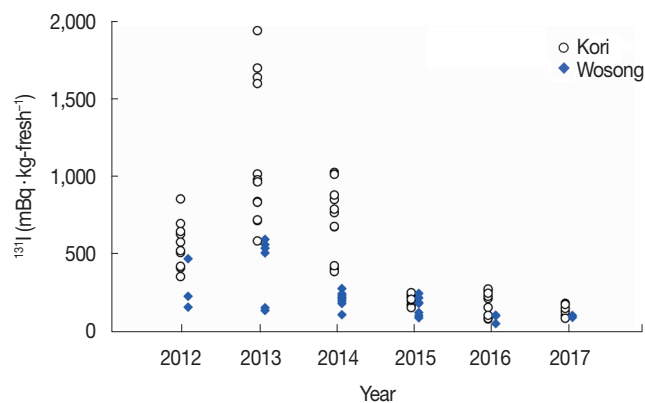


Fig. 6.  $^{131}\text{I}$  concentration variation with year in seaweed collected near Kori and Wolsong NPP.

sults from the last five years (Figure 6) also showed that it had been detected in only Kori and Wolsong, and Kori had always shown a higher concentration of  $^{131}\text{I}$  than Wolsong. It showed the maximum concentration in 2013, and the concentration and the detection rate gradually decreased, and for the last six years, the detection percentage in Kori and Wolsong was 85% and 68%, respectively. In these seaweed samples, however, other gamma radionuclides and  $^{137}\text{Cs}$  were not detected, and  $^{131}\text{I}$  was not detected in the other environmental samples (underground water and seawater). It is assumed that this originated from the  $^{131}\text{I}$  used in medical applications in hospitals, and hence,  $^{131}\text{I}$ , medical-origin, artificial radionuclide, was reported in the intake water and effluent in the river near Kori sewage treatment facility as well as seaweed collected near plant [16]. Therefore, the relatively

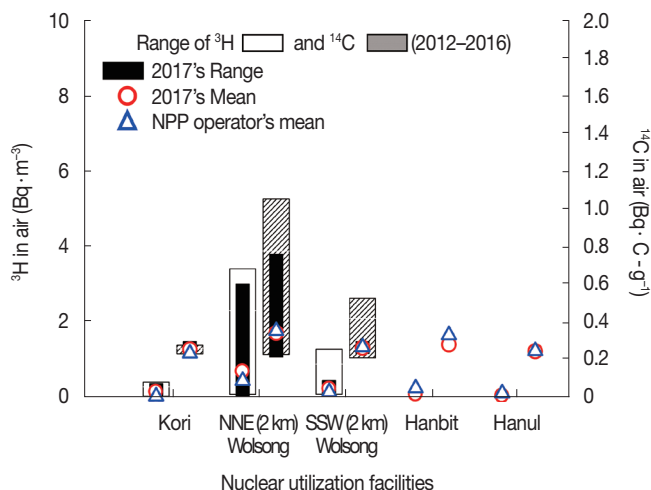


Fig. 7. Comparison of <sup>3</sup>H and <sup>14</sup>C in the air samples with previous year's data and NPP operator's mean.

high detection of <sup>131</sup>I in the seaweed samples near Kori was because of the effect of the <sup>131</sup>I, contained in the effluent from the large sewage treatment facility near the NPP site. Meanwhile, it is evaluated that the continuous reduction of the <sup>131</sup>I concentration and detection rate since 2013 was because of the strict control and improvement of the effluent management of <sup>131</sup>I in hospitals.

5. Tritium (<sup>3</sup>H) and carbon-14 (<sup>14</sup>C) results

Figure 7 demonstrates that the <sup>3</sup>H in the air near Wolsong NPP was higher than the other NPPs both in the 2017 survey result as well as in the last five years' reports; this was because of the operation of its heavy-water reactor, the <sup>3</sup>H emissions were higher than the other plants that used light-water reactors by five to eight times. In the case of <sup>14</sup>C, the emissions from Wolsong NPP were about twice higher than the others, but the difference in concentration was negligible, except for slightly high in Sangbong, Wolsong, in summer. Figure 8, illustrating the concentration variation in the air samples with direction near Wolsong NPP, shows the wind direction dependency (NEE > SSW) and seasonal variations. This trend was remarkable in <sup>3</sup>H having a large emission and high mobility as compared to <sup>14</sup>C. From the survey of the NPP emissions data in the last five years showed that the difference decreased owing to the reduction trend of gaseous <sup>3</sup>H waste recently.

The analysis results of the water samples showed that <sup>3</sup>H concentration in the rainwater sample from the inside of Wolsong NPP was considerably higher than in the water samples of the other NPPs (34.2–482 Bq · L<sup>-1</sup>), which is result-

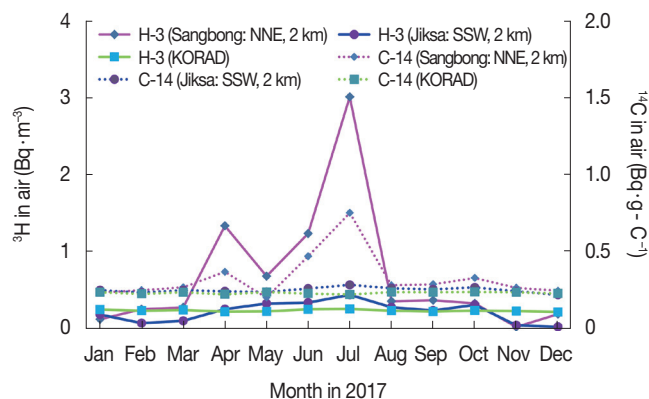


Fig. 8. Monthly variation of <sup>3</sup>H and <sup>14</sup>C in air samples with direction at Wolsong NPP.

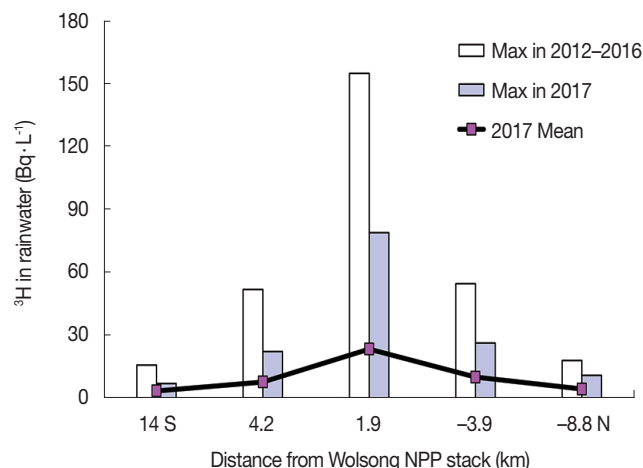


Fig. 9. <sup>3</sup>H concentration in the rainwater along distance from Wolsong NPP stack.

ing from the difference in the gaseous emissions at the plant. Figure 9 illustrates the <sup>3</sup>H results in rainwater samples with distance and direction near Wolsong NPP. <sup>3</sup>H concentration in rainwater exponentially decreased by distance, and compared to the previous data [17], the 2017 results showed the overall decrease in the concentration. The maximum concentration was 79 Bq · L<sup>-1</sup> from the sample collected at the staff housing in February, which was significantly lower than WTO's drinking water guidance level 10,000 Bq · L<sup>-1</sup> [18]. The concentration in underwater was as low as < MDC-3 Bq · L<sup>-1</sup> that is significantly lower than rainwater because of the dilution and natural decay as rainwater flows down underground, and there was little difference between the NPPs.

The <sup>3</sup>H and <sup>14</sup>C results in the pine needles near Wolsong NPP showed that only <sup>3</sup>H collected in May in Sangbong, close to NPP, was 2.5 times higher than the annual average concentration. This is attributed to the influence of a higher

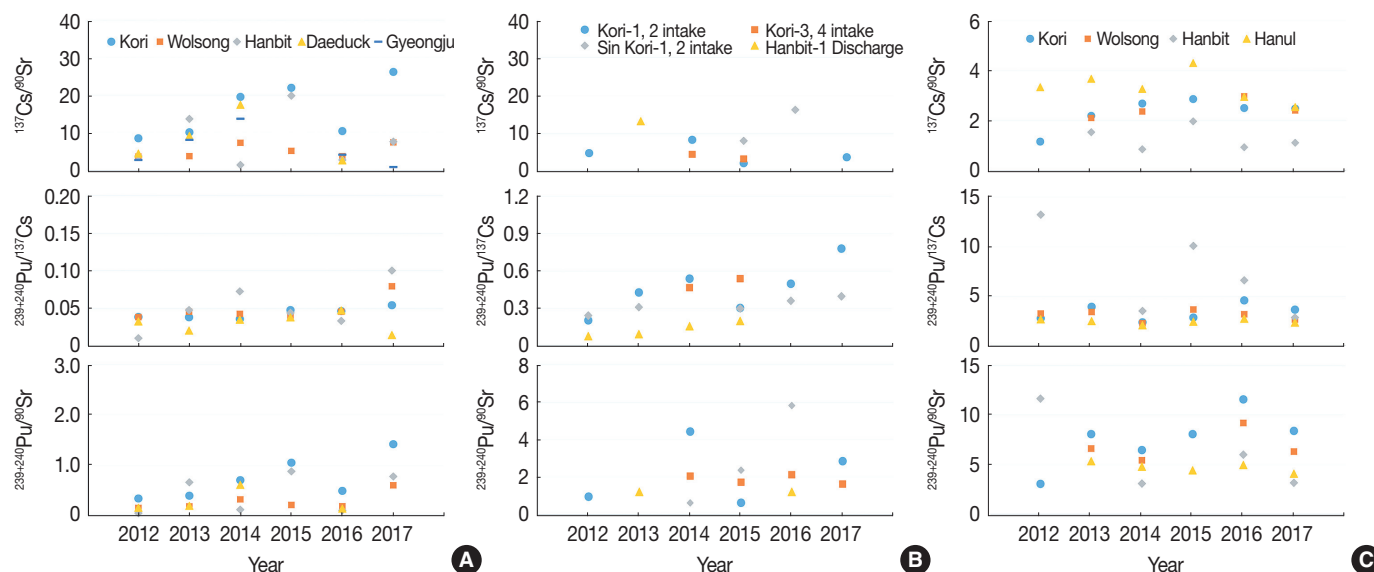


Fig. 10. Concentration ratios in anthropogenic radionuclides in soil (A), sediment (B), and seawater (C) samples in NPP sites.

monthly concentration of  $^3\text{H}$  in the air samples and seasonal factors, such as the seasonal wind direction and rainfall trends. However,  $^{14}\text{C}$  did not show any regional or seasonal differences because of low emissions and diffusions.

## 6. Radioactivity concentration ratio among $^{137}\text{Cs}$ , $^{90}\text{Sr}$ , and Pu

Figure 10 shows the comparison of the concentration ratio between radionuclides in the soil, seabed sediments, and seawater samples for the last five years. The average  $^{137}\text{Cs}/^{90}\text{Sr}$  concentration ratio in soil and seabed sediments for the last six years, including the 2017 data, was  $8.97 \pm 6.90$  and  $7.08 \pm 4.84$ , respectively. Considering the fluctuation of the concentration ratio in the environmental samples, these results had little difference compared to the previous results in soil ( $10.7 \pm 6.62$ ) [19] and the coastal seabed sediments ( $11.2 \pm 2.9$ ) [20] caused by global fallouts, which were surveyed in Korea in the past. The  $^{239+240}\text{Pu}$  radioactivity concentration ratio in relation to  $^{137}\text{Cs}$  and  $^{90}\text{Sr}$  in soil was  $0.04 \pm 0.02$  and  $0.43 \pm 0.36$ , respectively, and that in the seabed sediments was  $0.35 \pm 0.18$  and  $2.1 \pm 1.5$ , respectively; these results were higher than  $^{239+240}\text{Pu}/\text{Cs}$  ( $0.024 \pm 0.004$ ) and  $^{239+240}\text{Pu}/\text{Sr}$  concentration ratios ( $0.24 \pm 0.13$ ) [19] in Korean soil, surveyed in 1994, and it is believed that this is because of the concentration reduction mainly by the natural decay of  $^{137}\text{Cs}$  and  $^{90}\text{Sr}$ . In the case of the coastal seabed sediments, the concentration of Pu was higher than in soil because of the sinking of a particle-reactive Pu in seawater and the input from the land owing to natural erosion. Meanwhile, the  $^{137}\text{Cs}/^{90}\text{Sr}$  concen-

tration ratio in seawater in 2017 was  $2.40 \pm 0.91$ , and the  $^{239+240}\text{Pu}/^{137}\text{Cs}$  and  $^{239+240}\text{Pu}/^{90}\text{Sr}$  concentration ratios were  $3.88 \pm 2.67$  and  $7.08 \pm 4.12$ , respectively. Compared to the survey data in the past, seawater is less affected by the sampling location than soil, and therefore, there was little difference in the deviation between concentration ratio and sampling sites. The concentration ratios in these radionuclides in seawater were similar to the previous results surveyed from the East Sea [21, 22], and thus, it is believed that these radionuclides might have originated from previous nuclear tests or nuclear plant accidents.

## Conclusion

This study is significant in the assessment of the short- and long-term concentration fluctuation and evaluation of the source contribution by concentration ratios analysis between these radionuclides for anthropogenic radionuclides detected above MDC in the result of the environmental radiological surveillance around nuclear power utilization facilities in 2017. Among the  $\gamma$ -radionuclides,  $^{137}\text{Cs}$  was detected from some soil, seabed sediments, and fish samples.  $^{131}\text{I}$  was detected from seaweed samples near Kori and Wolsong NPP, but considering the sampling points and no detection of other artificial radionuclides, it was traced from medical applications.  $^{90}\text{Sr}$  was detected from most of the samples where  $^{137}\text{Cs}$  was also detected, and the radioactivity concentration of most of these radionuclides was low, which was just little over MDC, and all of them were within the normal



fluctuation. Most of  $^{137}\text{Cs}$ ,  $^{90}\text{Sr}$ , and Pu isotopes were detected from seawater for which the considerable amounts of sample were used, and all of them were within the normal fluctuation. The radioactive concentration ratio of these radionuclides in soil, seabed sediments, and seawater was similar to that of the global fallout caused by the previous nuclear tests. The Pu atom ratio ( $^{240}\text{Pu}/^{239}\text{Pu}$ ) from the land samples was different from that in the marine samples, and it is believed that this is because of the close-in fallout Pu from the previous Pacific Proving Grounds nuclear weapons tests, as opposed to global fallout Pu in the land. The  $^3\text{H}$  concentration was relatively high in the samples near Wolsong NPP (air, rainwater), and compared to that of  $^3\text{H}$ , the  $^{14}\text{C}$  concentration did not show much difference between NPPs, and the concentration was similar to the previous year.  $^3\text{H}$  in pine needle showed the concentration difference by distance and wind direction from NPPs due to its gaseous emissions. Most of the 2017 survey results were below MDC, and the concentration of artificial radionuclides detected was within the normal fluctuation and did not show any extraordinary results. It is expected that the outcomes of this study will be available in the verification of the surveillance results conducted by the operators of nuclear power utilization facilities and the assessments of the environmental impact in the vicinity of NPPs in case of the leakage of radioactive materials.

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