



Extent and persistence of dissolved oxygen enhancement using nanobubbles

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ABSTRACT

In this study, change in water-dissolved oxygen (DO) was analyzed under various synthetic water qualities and nanobubbles (NBs) application conditions, such as gas type, initial DO as well as water dissolved, suspended and organic matters contents. When oxygen, rather than air, was introduced into nitrogen-desorbed ultra-pure water, the stagnation time was significantly increased. It took ten days for DO concentration to drop back to saturation. The higher the initial DO concentration, the longer particles were observed above saturation due to particle stability improvement. The oxygen mass transfer rate of 0.0482 mg/L/min was found to reach a maximum at an electrolytic concentration of 0.75 g/L, beyond which the transfer rate decreased due to adsorption of negative ions of the electrolyte at the interface. High levels of turbidity caused by suspended solids have become a barrier to dissolution of NBs oxygen into the water solution, and thus affected the transfer performance. On the other hand, by applying NBs for just an hour, up to 7.2% degradation of glucose as representative organic matter was achieved. Thus, NBs technology would maintain a high DO extent for an extended duration, and thus can improve water quality provided that water chemistry is closely monitored during its application.

Keywords: Application conditions, Dissolved oxygen, Nanobubbles, Water quality

1. Introduction

Bubbles are gas bodies surrounded by water, and when macro size bubbles rapidly rise and burst at the surface of a water body, they empty a significant amount of the gas content into the atmosphere. However, applications in the areas of surface water remediation and wastewater treatment have shown that smaller bubbles have achieved huge mass transfer effects.

Michioku et al. [1] conducted an in-situ experiment by using very simple microbubble aeration, and the system was a powerful tool for minimizing reservoir eutrophication in an economical way. Srithongouthai et al. [2] improved dissolved oxygen (DO) levels in the net pen water to a level suitable for fish farming using microscopic bubbles during the warm seasons. Li et al. [3] suggested that microbubble-collapse process was valuable as a pre-treatment before the traditional biological processes are carried out to increase the effectiveness and economic feasibility of the overall treatment. Kim et al. [4] also carried out a field investigation on water purification technique in a brackish water lake in an effort to test the performance of microbubble system to increase

DO at the lower layer. In addition, the use of microbubbles in various environmental and industrial processes for solid-liquid separations and in mass transfer is well-established [5].

Recently, lab experiments with a liquid particle counter enabled recognition of the existence of stabilized invisible bubbles finer than the common size range of microbubbles [6]. These submicron bubbles are called nanobubbles (NBs). Due to the many potential applications of NBs in fields such as medicine and the food industry, the methods of generating and measuring bubbles have advanced. However, limited research has been carried out on NBs in water with inert electrolytes, or in the presence of surfactants [7]. Applied research in the area of water treatment has not been growing along with our understanding of the NB generation kinetics, bubble physicochemical properties, bubble dynamics and evaluation techniques [8-13]. The opportunity to conduct deeper studies that will provide additional insights into the practical application of NBs is wide open.

Since NBs cannot be thermodynamically stable, their persistence in solutions has been debated for decades [9, 14-15]. The difference between the internal pressure of bubbles and the atmospheric pressure governs the dissolution of gas into solutions. In this re-



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spect, NBs have such a huge internal pressure that they are supposed to dissolve instantly [12, 16]. But despite this, studies have shown that it is possible to produce NBs that have a stability lasting from several minutes to days, and even months in aqueous solutions [12, 14]. Thus, in order to practically extend the application of NBs in environmental management areas, it is important to reveal additional fundamental properties of the bubbles through the study of design parameters and physico-chemical factors.

In this study, bubbles of 800 to 900 nm in size having a life span of several hours to a few days were generated using a hydrodynamic cavitation method. The effects of operating parameters such as gas type and the water chemistry on the DO improvement extent and persistence of the NBs were investigated. More specifically, the effect of varying the initial oxygen content of the water and gas type on oxygen supersaturation in water was examined. In addition, the impact of dissolved and suspended solids on NBs DO enhancement extent was studied. To accomplish this, separate experiments using synthetic water in which sodium chloride (NaCl) and kaolin were dissolved were conducted.

Before assessing the effect, the correlation of sodium chloride as dissolved solid and the corresponding electrical conductivity (EC), as well as kaolin amount, representing the suspended solids and the resulting turbidity were established. The aim of the correlation was to mimic the recent EC and turbidity of the rivers of Korea, collected from the 'Real time water quality information.' The study also assessed the performance of NBs aeration in reducing the chemical oxygen demand (COD) level of synthetic wastewater made by dissolving glucose in ultra-pure water. The efficiency of the digestion process was analyzed based on the percentage of COD removal and of COD reduction.

2. Materials and Methods

2.1. Effect of Gas Type and Initial DO Concentration

The NBs were generated using ultrapure water, which passed through ion exchange, activated carbon and membrane filtration particle removal processes. The ultrapure water used had conductivity $4.5 \mu\text{S}/\text{cm}$, salinity of $0.0 \text{ mg}/\text{L}$ and surface tension of $0.0725 \text{ N}/\text{m}$, all at 25°C . The gas used for the bubbles generation was natural air or oxygen. The air supply was provided using Hi-POWER master air pump (DAE Kwang, Korea) while pure oxygen was supplied to the NBs generator from an oxygen cylinder (99.99% purity) at $0.025 \text{ L}/\text{min}$ flow rate. Oxygen was nano-bubbled into nitrogen-desorbed ultra-pure water, which itself was prepared by introducing nitrogen (99.99% purity) NBs into the water to adjust the initial DO concentration.

The NBs were discharged into the bottom center of a cylindrical Pyrex glass reactor (120 mm dia. and holding 6 L of water) at a rate of $0.25 \text{ L}/\text{min}$. The water was allowed to circulate between the NB generator and the reactor, while the DO concentration of ultra-pure water kept in the reactor was continuously measured until it became stable. The DO concentration measurement was taken using an A223 RDO portable meter (Thermo Scientific, USA), which automatically compensates variation of temperature from 20°C . The DO meter had measuring range from $0 \text{ mg}/\text{L}$ to $50 \text{ mg}/\text{L}$

with accuracy of $\pm 1\%$ ($0\text{-}20 \text{ mg}/\text{L}$) and $\pm 10\%$ ($20\text{-}50 \text{ mg}/\text{L}$). Water samples were collected for daily measurement after the generation, until the DO value reached saturation equilibrium concentration.

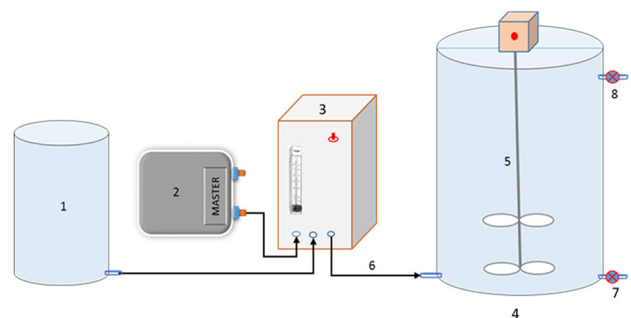
The temperature of the liquid in the reactor was kept constant by allowing a continuous flow of cooling water in the jacket around the reactor. Initial DO of 3, 4, 5, 7, 8 and $9 \text{ mg}/\text{L}$ at 20°C temperature were considered. The widely-applied ASCE standard guideline assumption to determine mass transfer has reduced design variability and allowed a better evaluation of the operation of existing treatment plants [17]. However, Muroyama *et al.* [18] and Sadatomi *et al.* [19] considered initial DO concentration of $2 \text{ mg}/\text{L}$ and $4 \text{ mg}/\text{L}$, respectively, which also deviated from the standard value of zero. The effect of the ratio of the gas-water exchange area to water volume on the time to reach equilibrium DO concentration was also examined.

2.2. Effect of Electrolyte Concentration and Turbidity on DO Nature

The experimental set-up of this section is shown in Fig. 1. The actual EC and turbidity data of the four major river systems of Korea from Jul. 2012 - Jan. 2015 were assessed to keep the EC and turbidity of the synthetic water reasonable.

99.0% extra pure NaCl and extra pure kaolin ($\text{H}_2\text{Al}_2\text{Si}_2\text{O}_8 \cdot \text{H}_2\text{O}$), both sourced from SAMCHUN Pure Chemical, Korea, were used to prepare the synthetic electrolytic and the synthetic turbid water, respectively. Several trial experimental data were collected to determine what turbidity would be produced from a given suspension of kaolin clay, as well as to relate the amount of NaCl added to the EC measured. YSI professional series product family (USA), pro 1030 conductivity meter and the Hach Model 2100AN IS Turbidimeter were used for EC and turbidity measurement, respectively.

In an effort to determine the relationship between concentration of the solute and EC, 0, 0.025, 0.05, 0.15, 0.25, 0.5, 0.75, 1, 1.25 and $1.5 \text{ g}/\text{L}$ NaCl were considered. The turbidity experiment was carried out by adding 0, 25, 50, 100, 200, 400, 600, 800, 1,000 and $1,200 \text{ mg}/\text{L}$ of kaolin clay to the 20 L ultra-pure water in the reactor, after establishing the correlation between turbidity and the kaolin amount causing it. The agitator was set to run continuously to avoid sedimentation, while pH and DO measurements



- | | |
|--------------------------|----------------------------|
| 1. Ultra-pure water tank | 2. Air Compressor |
| 3. Nanobubbles Generator | 4. Synthetic water tank |
| 5. Agitator | 6. Nanobubbles supply line |
| 7. Drain & sampling port | 8. Overflow line |

Fig. 1. The experimental set-up.

in sample water from the reactor were made at 3 min intervals. The data collection continued until the reading stabilized to a constant value with ± 0.2 . The dependence of NBs oxygen transfer efficiency (OTE) on turbidity and electrolyte concentration was demonstrated through a comparison of mass transfer data collected by adding the known amounts of kaolin and NaCl, respectively. The extent of DO enhancement was analyzed from the DO peak value (DOPV) and the oxygen transfer rate.

2.3. Effect of Glucose and Its Amount on DO Increasing Pattern

The actual COD data of the four major river systems of Korea were assessed to keep the COD of the synthetic water reasonable. COD value of wastewater influent was also taken into account. Extra pure anhydrous glucose from SAMCHUN Pure Chemical Co., Ltd was used to prepare the synthetic wastewater as carbon source. The COD measurement was made using a DR/4000 spectrophotometer (Hach Company, USA). Based on the theoretical concept that a solution with 1 g/L of glucose has a COD of 1,067 mg/L, the main experiment was conducted by adding 0, 0.1, 0.25, 0.5, 1, 2, 3, 4 and 5 g of glucose to the 10 L ultra-pure water in the reactor. The DO measurements in sample water from the reactor were made at 5 min intervals. The NBs were continuously supplied to the glucose solution for 1 h, samples were collected at 0 min, 30 min and 60 min, and then the COD content was analyzed for organic matter degradation.

2.4. Parameters Used for Comparison of the Factors

To analyze the collected DO data, a non-steady mass transfer model (Eq. (1)), by which the volumetric oxygen mass transfer rate (VOTR) is determined [17], was employed.

$$\frac{dC}{dt} = K_L a^* (C_\infty - C_t) \quad (1)$$

where:

dc/dt - transfer rate per unit volume;

$K_L a$ - VOTR (1/time);

C_∞ - near saturation DO concentration at infinite time (mg/L) and

C_t - DO at time t (mg/L).

This standard model assumes that the equilibrium DO concentrations are the same everywhere in the tank, which is reasonable in this work as the reactor is small in size. To visualize the DO-vs-time experimental data, linear regression is used by transforming Eq. (2) to the following form:

$$\ln \left(\frac{C_\infty - C_t}{C_\infty - C_0} \right) = K_L a^* t \quad (2)$$

Here, C_0 is initial concentration at time $t=0$, C_t and $K_L a$ are as defined above. The slope of $\ln((C_s - C_t)/(C_s - C_0))$ vs time plot gives the $K_L a$ value.

The standard method for oxygen transfer measurements assumes overall oxygen transfer to be only from air bubbles, and it considers oxygen transfer from the water's surface to be negligible. In support of this assumption, field oxygen transfer measurement conducted

by He [20] concluded that the liquid-side mass transfer value was only 2% of the bubble side transfer. The surface area of the cylindrical water tank to which the NBs were supplied in this work was small in size. In addition, a large amount of oxygen was transferred to the water across the interfaces as the very slow terminal rising velocity of the tiny bubbles from the bottom to the water's surface gave time for more dissolution. Actually, the liquid-side mass transfer coefficient (k_L) decreases with a decrease in the bubble size [21]. Hence, k_L and interfacial area (a) were not considered here.

To show the extent and persistence of dissolved oxygen enhancement, three parameters - namely, DOPV, average initial DO increasing rate (AIDOIR) and stagnation time (ST) - were used [22]. DOPV is the peak to which the DO concentration rises, or is above the stable saturation value. The AIDOIR is the mean of the DO increase from its initial value to the peak per unit time. It describes the quantity and the speed at which the generator produces NBs, which in turn determine the oxygen mass transfer rate. In addition, the ST is the time taken by the DO to return to the stable saturation value (100%) from its corresponding peak.

3. Results and Discussion

3.1. Variation of DO over Time Due to NBs

DO concentration of the water before nitrogen desorption ranged from 9.10 to 9.64 at 20°C. At the outset, the already dissolved oxygen content was reduced to about 4 mg/L using nitrogen gas, because the tap water showed high DO value and it required a significant amount of time to reduce further, close to zero [19]. After the air NBs supply stopped, the DO concentration stabilized to an average concentration of 13.5 mg/L, for all initial DO values. The concentration was almost unchanged for 24 h after the generator stopped. It has taken two more days for the DO concentration to decrease to the saturated value of about 9.3 mg/L. After an additional day, the concentration reached the saturation equilibrium at 9.1 mg/L.

Looking at the variation in oxygen gas NBs DO concentration over time (Fig. 2), it can be seen that the value reached peak

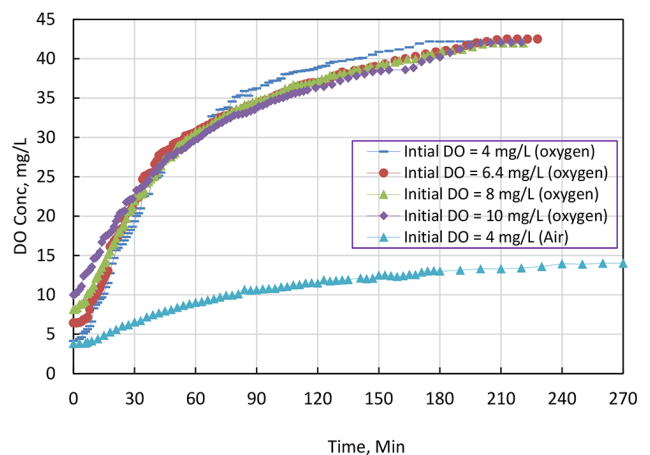


Fig. 2. Variation of dissolved oxygen (DO) with time, for the different initial DO concentrations.

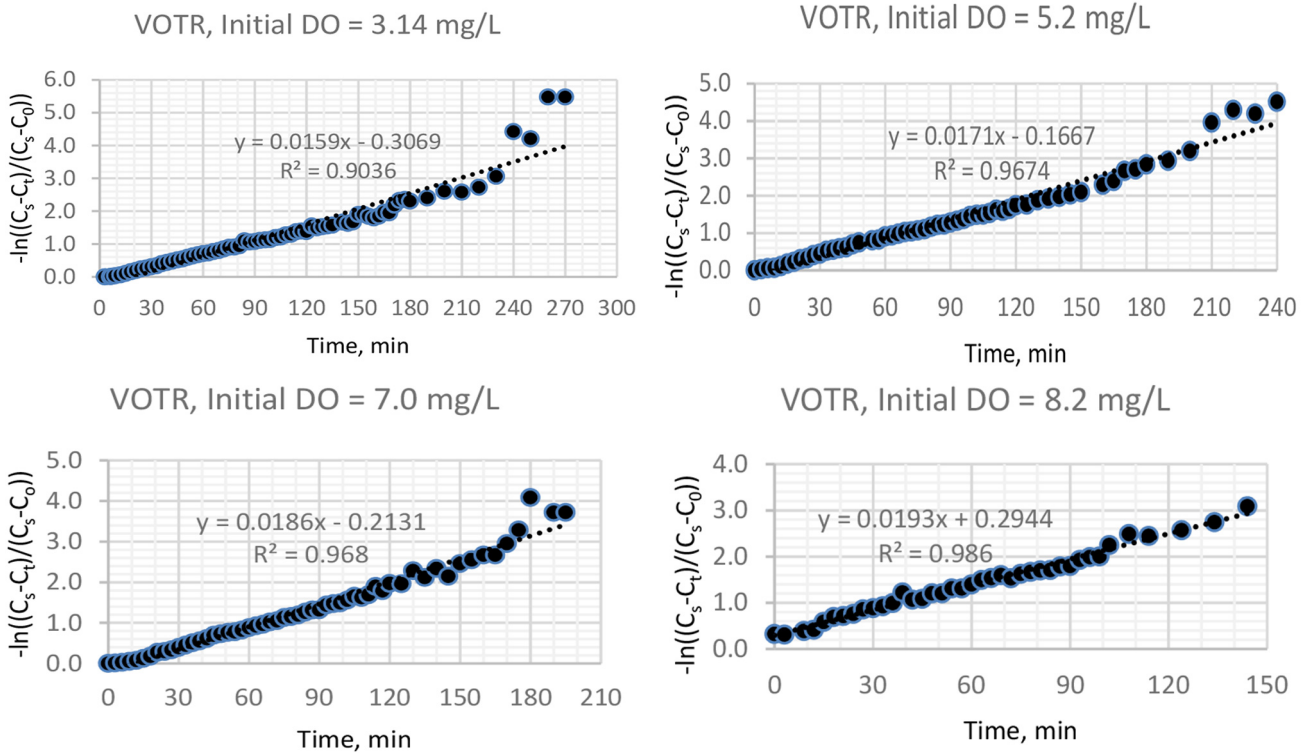


Fig. 3. Volumetric oxygen transfer rate (VOTR) by air NBs for different initial oxygen concentrations.

supersaturated level of 42.0 mg/L. It was again almost unchanged after 1 d, and then slowly decreased to the saturation equilibrium of 9.1 mg/L at 20°C. The DO concentration approached saturation condition (9.3 mg/L) on the 9th day, and after one more day it reached the saturation equilibrium of 9.1 mg/L. In this study, immediately after the NBs generation started, the DO increased relatively rapidly, until it reached 30 mg/L after about an hour. However, the rate of increase was reduced afterwards, and oscillated slightly around the peak value of 42.0 mg/L.

Fig. 3 shows that the higher the initial air DO, the higher the kLa value, although the trend is one of a gradual increase. Generally, the kLa value for nano-bubbling was not very high (0.0159-0.0193 min^{-1} for air) because the NB generator had limited capacity. When oxygen was used instead of air, the kLa value decreased slightly with the increase in initial DO, but still had a magnitude as high as 3 times that of air. Thus, the dissolution of oxygen from NBs to the water depends on the gas type and the initial concentration of oxygen. Oxygen dissolution in water follows Henry's law, with the dissolved amount proportional to the partial pressure of oxygen in air.

The AIDOIR indicates the mass transfer rate of the NBs (Fig. 4). It is related to the generator capacity, and gives the number, the rate and the size of bubbles generated. Li et al. [22] explained that mass transfer rate will be increased by ten thousand if one 1 mm-diameter air bubble is divided into eight hundred thousand 10 μm -diameter air bubbles, to determine why AIDOIR of micro-nano bubbles was much higher than the macro bubbles. Different time duration was taken to reach the peak saturation concentration that AIDOIR was not just difference between the

DO concentration values. DOPV was the basic result of AIDOIR over time.

Fig. 5 shows the effect of initial oxygen concentration in terms of OTE, just to compare the AIDOIR result with another universal parameter. OTE is the ratio of mass rate of oxygen transferred to mass rate of oxygen supplied. Thus, the bar graphs show the apparent effect of gas type and initial DO.

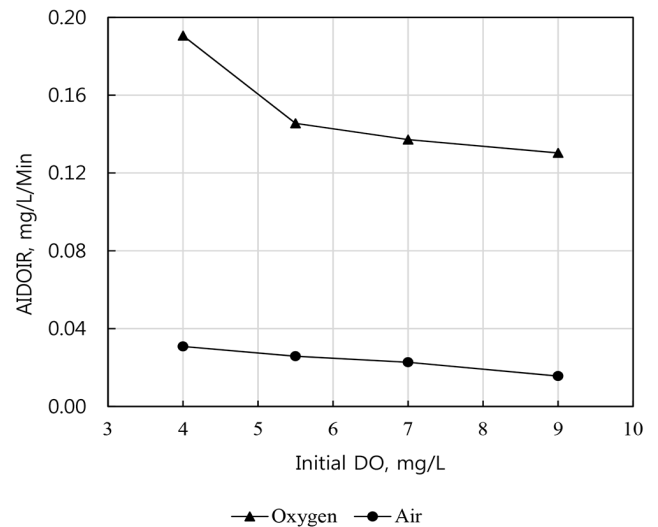


Fig. 4. Variation of average initial dissolved oxygen increasing rate (AIDOIR) based on initial DO.

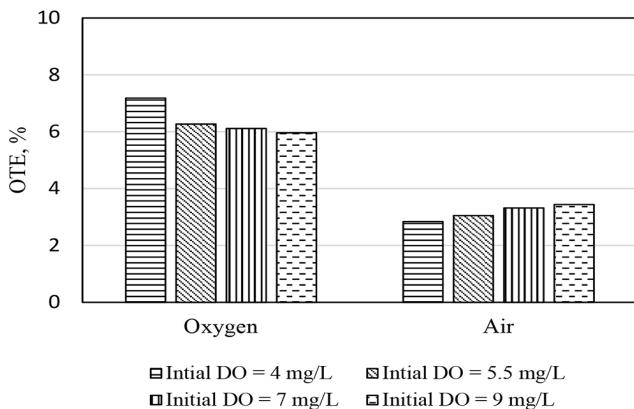


Fig. 5. Effect of initial dissolved oxygen concentration and gas type on oxygen transfer efficiency, OTE (%).

3.2. Stagnation Time of NBs

The life span of the NBs staying dissolved in water is given by ST. ST is associated with bubble size, the initial DO, gas type and ratio of gas-water exchange area to the water volume. Thus, ST is a parameter that roughly shows the durability of DO caused by the NBs. When the DOPV was reached, the generation stopped; at this point, the DO decreased due to the oxygen in the bubbles having been dissolved into the water and finally diffused to the atmosphere. Fig. 6 presents the stagnation duration for different initial DOs. The higher the initial DO, the longer the DO stayed beyond the equilibrium saturation value, provided that the samples considered were of the same size.

Recently, it has been considered that the internal pressure of NBs should be much lower than predicted by the Young-Laplace equation, and thus the equation should no longer hold for NBs. Instead, the liquid-gas interface may play the central role in the stabilization of NBs. Ushikubo et al. [12] explained the stability of nano-sized bubbles according to the high dissolved gas concentration in water and the electrically charged interface of the bubbles. At a higher initial DO concentration, particles were observed for a longer period of time. Thus, high DO concentration could extend the particle stability. This verifies the possibility of the existence of NBs, since at a higher DO concentration, the oxygen diffusion would take more time. The high gas concentration in the liquid explains the extension of NB stability.

The magnitudes of zeta potential of oxygen nano-bubbled water were possibly higher than those of air nano-bubbled water. At a high magnitude zeta potential value, the electrically charged particles tend to repel each other, avoiding aggregation. For bubble dispersion, the high zeta potential could create repulsion forces that would avoid the coalescence of bubbles, thus contributing to the stabilization of the bubbles. Therefore, the oxygen NBs could be stabilized in water by the electrically charged surfaces, while the nano-bubbled air would be below the limit of the stabilization by repulsion forces.

Furthermore, the big difference between the periods in which bubbles could be observed as stable in the oxygen nano-bubbled water (10-15 d) and in that of air nano-bubbled (3-5 d) is likely

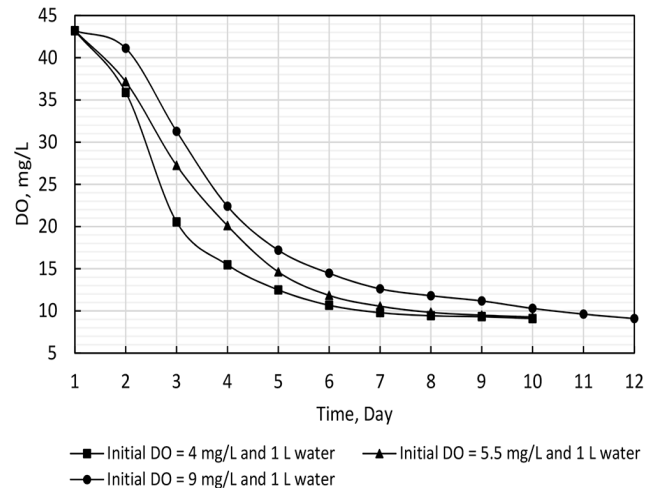


Fig. 6. Graphs showing stagnation time for different initial dissolved oxygen (DO).

related to the lower dissolution of the air in water in relation to the dissolution of oxygen. At a lower concentration of dissolved gas, the concentration gradient across the gas-liquid interface of the bubble was likely higher, leading to a more efficient gas transfer to the liquid phase.

The dependence of the ST on the ratio of gas-water exchange area to the water volume (A/V) was observed by storing nano-bubbled water, immediately after its generation, in beakers of different sizes. Three surface areas of 127 cm², 183 cm², and 183 cm² with corresponding volumes of 1 L, 1.5 L and 2 L, respectively, were used. The DO in the holder of A/V ratio 0.091 per cm took two more days to reach the equilibrium condition than the one kept in A/V ratio of 0.130 per cm, both having the same surface area but different volumes. As such, because of the increased ratio, the DO ST was affected in such a way that it approached the equilibrium faster. The persistence decreased from the 1 L to 2 L volume, showing the effect of the interface rather than water volume. This was also suggested by Ushikubo et al. [12].

3.3. Influence of Turbidity and Its Amount on DO Increasing Pattern

Instead of adding arbitrary amounts of kaolin to attain a certain turbidity, an experiment was carried out to determine the amount of kaolin clay per liter that should be added to achieve a desired turbidity. Thus, the experiment was conducted so that the amount of kaolin suspension of the synthetic water resulted in turbidity in the actual range of 0 NTU and 1,107 NTU. The actual range was taken from the recently available daily minimum and maximum turbidities of all stations of the four major basins of South Korea. Based on the experimental data collected, the relationship between concentration of clay and turbidity was determined (Fig. 7). From the data points, two separate linear regression relations were derived, one up to 336 NTU and the other for greater values.

The effect of varying the amount of kaolin added while making the synthetic water and then applying the NBs was then examined. The rate of DO increase was slightly affected by low turbid water.

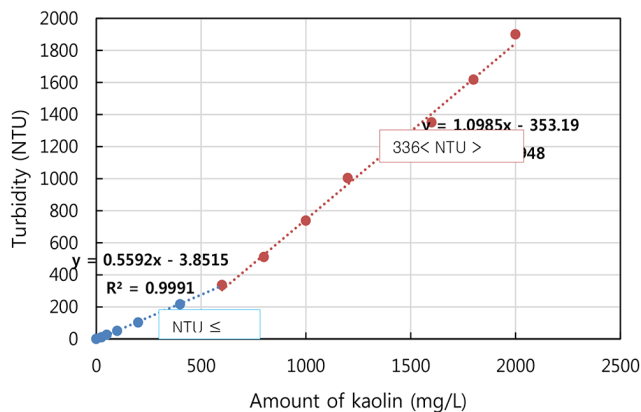


Fig. 7. Relationship between amount of kaolin clay (mg/L) and the corresponding turbidity (NTU).

However, highly turbid water, especially with turbidity above 336 NTU, significantly affected the DOPV reached and the transfer rate. For water turbidity of 10 and 26 NTU, the calculated oxygen transfer rate was raised to 0.0473 mg/L/min and 0.0553 mg/L/min, respectively, compared to 0.0405 mg/L/min of the turbidity free water. On the other hand, for turbidity values equal to or greater than 336 NTU, the peak DO was limited to 11.0 mg/L. However, it was able to reach a peak of 13.32 mg/L in limited turbidity cases.

Oxygen was more easily dissolved into water with low levels of suspended solids. The dissolved and suspended particles which significantly contribute to turbidity depress the hydrodynamic activity and offer barriers to the passage of gas molecules at the gas-liquid interface [23, 24]. When present in water, suspended and dissolved solids, which influence the surface tension and viscosity, affect the size of air bubbles released from a generator and reduce the transfer of oxygen into the solution. Solid particles collect on the bubble surface, forming a wall that interferes with gas diffusion. The reduction in oxygen transfer varies with the extent of surface tension change.

Thus, the interaction of the suspended particles and of NBs affects the dissolution of oxygen. The data analyses of the stream stations recording turbidity showed that only five of the thirty-six registered values were above 336 NTU. To increase the efficiency of oxygen supply using nano-technology, it is important to consider suspension carried by the water to be aerated. By decreasing the suspended solids or selecting proper timing related to the suspension amount, i.e. turbidity, the oxygen dissolution by NBs can be improved.

3.4. Effect of NBs on Glucose Degradation

According to recent real-time water quality information, COD of Korea's four major rivers is in the range of 0 to 35 mg/L, with most of the streams showing values below 10 mg/L. To include the application of NBs for wastewater treatment, COD values up

to 500 mg/L were considered in this study. Glucose decomposition by air NBs is depicted in table 1. Based on the results, percent reductions of 2.76 to 7.16 in COD content were achieved during the 1 h NBs supply period. Li et al. [3] achieved comparable results after the 1 h. mark by using a micro-bubbles collapse technique on phenol degradation. However, in the present experiment, the increase in glucose concentration did not show a significant trend, as the values in Table 1 show. Oxygen nano-bubbling facilitates the chemical reaction at the gas-liquid interface and the formation of OH⁻ radicals, which in turn aids organic decomposition [25, 26]. The bubbles react with the organic matter contained in the liquid, whereby the organic matter is oxidized and degraded.

Takahashi and Li [25] coupled the application of NBs with the UV process, which dramatically increased the efficiency of organic matter removal. In this study, hydrogen peroxide is combined with the nano-bubbling, as it is often combined with other oxidizing processes, to enhance the decomposition effect. In addition, hydrogen peroxide, H₂O₂, was selected because it shows immediate and unselective reactivity with the majority of organic compounds [3]. It has also been used as a DO supplement for BOD and COD [27]. Thus, to analyze the combined effect of NBs and H₂O₂ on the degradation, 100 mg/L of glucose was used to form the synthetic water. 10 mg/L of H₂O₂, which is equivalent to 0.089 mL (density of 1.12 g/mL) was also spiked into the solution.

The H₂O₂ clearly affected the DO. The DO is improved simply by the H₂O₂ addition, or without NBs application. In addition, the DO increased to a very high value of 14.7 mg/L, compared to the 13.5 mg/L without H₂O₂. Furthermore, the glucose degradation was better when hydrogen peroxide was used, even though the improvement was not significantly high. Hydrogen peroxide decomposes to form water and oxygen, and thus can increase the DO concentration above the steady state concentration [26]. In this way, it provides a source of reactive oxygen species, and can be used as an oxygen supplement.

3.5. Effect of Electrolyte Concentration

The minimum and maximum values at each of the stations of the major rivers and their tributaries for the duration of 31 mon were considered. 44 μS/cm and 2,801 μS/cm EC, respectively, of the Han River were taken to represent the overall range. Based on the 'rule of thumb' for conductivity vs. concentration at 25°C, the EC range corresponds to NaCl concentration of 22 mg/L - 1,420 mg/L. However, rather than fixing the NaCl concentration based on a general theoretical concept, a simple experiment was performed to determine what amount of NaCl had to be added per liter to create a desired conductivity. The linear regression equation representing this relationship between concentration, C (mg/L), of the solute and EC (spc-μS/cm), (C = 0.0005EC - 0.0309) was used for the main experiment to achieve a certain conductivity.

Table 1. Percent Reduction of Chemical Oxygen Demand (COD) after 30 Min and 1 H. of NBs Application (with and without Hydrogen Peroxide, H₂O₂)

Time (min)	Concentration of glucose, mg/L								
	10	50	100	150	200	300	400	500	100 + H ₂ O ₂
30	1.00	1.94	2.87	4.35	2.41	4.05	2.22	1.30	5.52
60	3.23	5.52	3.59	4.61	5.36	7.16	2.76	3.47	8.12

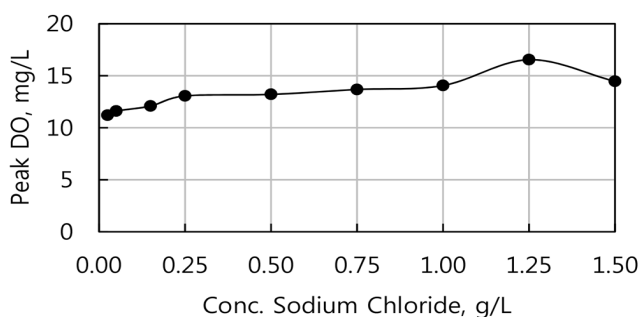


Fig. 8. The influence of electrolyte concentration on DO peak value.

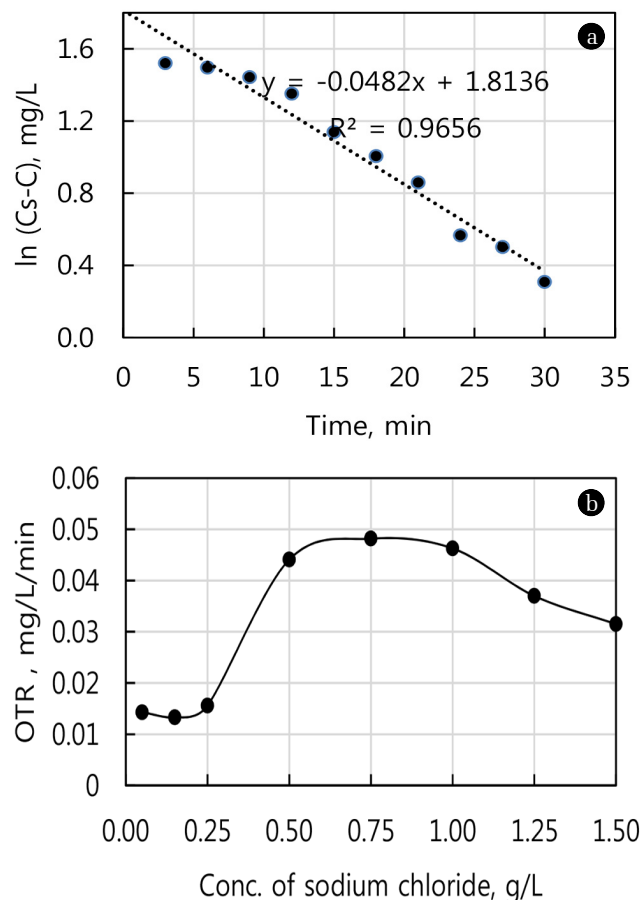


Fig. 9. (a) The oxygen mass transfer rate for NaCl concentration of 0.75 g/L and (b) The influence of electrolyte concentration on oxygen transfer rate per unit volume (OTR).

It can be recognized that, without the addition of NaCl, the DO ascended to a value of 13.21 mg/L before stabilizing at 12.38 mg/L. When solute was added, the DO increased continuously to a peak value and then dropped to a stable value, in a relatively short span of time. This DO variation pattern was similar for all the different concentrations considered. The DO increased to the peak value of 16.54 mg/L and then quickly stabilized at 11.20 mg/L, for a concentration of 1.25 g/L.

Fig. 8 and Fig. 9 show the effects of the different amounts of NaCl considered on the DOPVs and the oxygen mass transfer rate. The concentration of 1.25 g/L was the turning point, with a peak value of 16.54 mg/L. The influence of the electrolytic solution concentration was decreased thereafter. On the other hand, the oxygen transfer rate reached the maximum value of 0.0482 mg/L/min for a concentration of 0.75 g/L. The transfer rate was comparatively small with concentrations below 0.25 g/L. Besides, the rate showed a decreasing trend when the concentration was increased beyond 0.75 g/L.

Hydration energy, which is energy released when ions come into contact with water molecules, is the governing mechanism. The hydration energy of OH^- is much lower than that of H^+ , keeping a greater number of OH^- at the gas-water interface. The interface charge is easily influenced by such parameters as electrolyte concentration. Thus, when water into which NaCl is dissolved is nano-bubbled, Na^+ is adsorbed as a counter ion for the excess OH^- at the interface [28]. In this study, at salinity of 1.5 g/L, an equilibrium state at the interface was likely achieved with minimum zeta potential value. Increasing the salinity further, Cl^- would replace adsorbing at the interface in excess of Na^+ . Above 1.5 g/L, the zeta potential gradually increased, and this in turn caused the release of oxygen to the atmosphere and thus lowered DO peak. This causes the increase in zeta potential, and thus a decrease of DOPV. Hence, the use of NBs for bioremediation and other cases requiring massive oxygen transfer is greatly affected by the salinity of the water.

The NaCl ions interfere with the performance of the NBs by exerting resistance to the diffusion of the gases contained in the bubbles. The resistance increased with concentration of the ions up to an equilibrium state of 0.75 g/L, after which it started to decline, and so was the DO stagnation in the water. However, in the deionized or ultra-pure water, the increasing pattern was different, as the DO became stabilized around the corresponding peak value. Generally, both the DOPV and the oxygen mass transfer rate showed a rising pattern up to a certain concentration and then started to decrease. Therefore, electrolyte concentration affects the interaction of NBs in oxygen mass transfer.

4. Conclusions

When applying microbubbles to water to supply oxygen, the bubbles last in water for only a few minutes, but NBs can last in water for days. Thus, DO can be well-maintained for a long time with NBs. The effectiveness of NBs depends directly on the number of NBs and on their residence time. This study showed that the initial DO concentration affected the time taken to reach the saturation equilibrium, possibly because a high initial DO caused a greater number of particles to nucleate. At low DO concentration, cavitation bubbles are harder to generate and the distance between bubbles is too long to keep the size of bubbles small, as there is less coalescence of bubbles. At higher DO concentration, coalescence between bubbles is repeated to increase the number of bubbles. Thus, at higher DO concentration, a great number of bubbles with good size distribution were likely generated following coalescence between the bubbles.

Gas type also plays a major role in mass transfer and ST, as oxygen NBs have higher OTE and more durable DO enhancement

than air NBs. Therefore, providing oxygen in the form of NBs is promising in terms of oxygen supply processes such as for water body oxygenation, which can be used to control algal bloom and related water quality problems. DO increased with electrolyte concentration as the tiny bubbles were stabilized, and in this way, their life span in water was extended. On the other hand, the gas liquid interface charge was affected by electrolyte concentration in such a way that when salinity was increased above 1.5 g/L, Cl^- would replace, adsorbing at the interface in excess of Na^+ . This contributed to the release of oxygen to the atmosphere and thus lowered DO peak.

A high concentration of particles, which contributes to turbidity, affects the NBs effect in oxygen mass transfer to water by presenting a barrier to the passage of gas molecules at the gas-liquid interface. Even though turbidity in the water column is mainly caused by suspended solids, other constituents, such as organic matter and microorganisms, contribute to turbidity of water; the latter is believed to consume DO. Therefore, the situation of actual water should be considered in applications.

The application of NBs can also play a role in the mineralization of organic compounds so that effectiveness and economic feasibility of biological processes involving aeration are optimized. NBs have been shown to be useful for improving the degradation of organic matter, as the formation of OH radicals on NBs contributes to the removal of organic matter. Through NBs application, up to 7.2% decomposition of glucose as representative organic matter was achieved.

Thus, even if thermodynamics theory cannot describe the internal pressure of the NB, its stability and its release of OH radicals can contribute to a sustainable environmental treatment. In this study, the persistence of nano-sized bubbles in water was examined in terms of DO concentration and its prolonged existence beyond the saturation level. The quantitative description of the physical properties of the NBs such as the zeta potential and particle size provides a full explanation of the factors.

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Common Symbols used

Symbol	Unit	Description
dc/dt	-	Oxygen transfer rate per unit volume
$k_L a$	1/time	Volumetric oxygen transfer rate (VOTR)
C_∞	mg/L	Near saturation dissolved oxygen (DO) conc. at infinite time
C_t	mg/L	DO at time t
C_0	mg/L	initial DO concentration at time t = 0
EC	spc- $\mu\text{S}/\text{cm}$	Electrical conductivity
BOD	mg/L	Biochemical oxygen demand
COD	mg/L	Chemical oxygen demand

References

1. Michioku K, Sakatani Y, Matsuo K, Oda T, Hara Y. Reservoir purification by using micro-bubble aerator. In: 7th International Conference on Hydrosience and Engineering (ICHE); 10-13 September 2006; Philadelphia, USA.
2. Srithongouthai S, Endo A, Inoue A, et al. Control of dissolved oxygen levels of water in net pens for fish farming by a microscopic bubble generating system. *Fisheries Sci.* 2006;72:485-493.
3. Li P, Takahashi M, Chiba K. Degradation of phenol by the collapse of microbubbles. *Chemosphere* 2009;75:1371-1375.
4. Kim KH, Lee KH, Adachi Y, Pyun CK. Field investigation of water purification technique in the lagoon. In: 6th International Conference on Offshore and Polar Engineering; 28 May-2 June 2006; California, USA.
5. Burns SE, Yiaccoumi S, Tsouris C. Microbubble generation for environmental and industrial separations. *Sep. Purif. Technol.* 1997;11:221-232.
6. Takahashi M. Nanobubbles: An introduction. In: Tsuge H, ed. Micro- and nanobubbles: Fundamentals and applications. Singapore: Pan Stanford Publishing; 2014. p. 307-315.
7. Calgaroto S, Wilberg KQ, Rubio J. On the nanobubbles interfacial properties and future applications in flotation. *Miner. Eng.* 2014;60:33-40.
8. Agarwal A, Ng WJ, Liu Y. Principle and applications of micro-bubble and nanobubble technology for water treatment. *Chemosphere* 2011;84:1175-1180.
9. Cho SH, Kim JY, Chun JH, Kim JD. Ultrasonic formation of nanobubbles and their zeta-potentials in aqueous electrolyte and surfactant solutions. *Colloid. Surface. A.* 2005;269:28-34.
10. Kikuchi K, Ioka A, Oku T, Tanaka Y, Saihara Y, Ogumi Z. Concentration determination of oxygen nanobubbles in electrolyzed water. *J. Colloid Interf. Sci.* 2009;329:306-309.
11. Uchida T, Oshita S, Ohmori M, et al. Transmission electron microscopic observations of nanobubbles and their capture of impurities in wastewater. *Nanoscale Res. Lett.* 2011;6:295.
12. Ushikubo FY, Furukawa T, Nakagawa R, et al. Evidence of the existence and the stability of nano-bubbles in water. *Colloid. Surface. A.* 2010;361:31-37.
13. Wu C, Nasset N, Masliyah J, Xu Z. Generation and characterization of submicron size bubbles. *Adv. Colloid Interf. Sci.* 2012;179-182:123-132.
14. Najafi AS, Drelich J, Yeung A, Xu Z, Masliyah J. A novel method of measuring electrophoretic mobility of gas bubbles. *J. Colloid Interf. Sci.* 2007;308:344-350.
15. Yount, DE. On the elastic properties of the interfaces that stabilize gas cavitation nuclei. *J. Colloid Interf. Sci.* 1997;193:50-59.
16. Hampton MA, Nguyen AV. Nanobubbles and the nanobubble bridging capillary force. *Adv. Colloid Interf. Sci.* 2010;154:30-55.
17. Jiang P. Gas transfer parameter estimation: Applications and implications of classical assumptions [dissertation]. USA: Univ. of California; 2010.
18. Muroyama K, Imai K, Oka Y, Hayashi J. Mass transfer properties in a bubble column associated with micro-bubble dispersions. *Chem. Eng. Sci.* 2013;100:464-473.
19. Sadatomi M, Kawahara A, Matsuura H, Shikatani S. Micro-bubble generation rate and bubble dissolution rate into water by

- a simple multi-fluid mixer with orifice and porous tube. *Exp. Therm. Fluid Sci.* 2012;41:23-30.
20. He Y. An alternative mathematical model for oxygen transfer evaluation in clean water. Case Study; c2014 [cited 20 June 2014]. Available from: <http://www.wateronline.com/doc/an-alternative-mathematical-model-for-oxygen-transfer>.
21. Kawahara A, Sadatomi M, Matsuyama F, Matsuura H, Tominaga M, Noguchi M. Prediction of micro-bubble dissolution characteristics in water and seawater. *Exp. Therm. Fluid Sci.* 2009;33: 883-894.
22. Li H, Hu L, Song D, Al-Tabbaa A. Subsurface transport behavior of micro-nano bubbles and potential applications for groundwater remediation. *Int. J. Environ. Res. Public Health* 2014;11: 473-486.
23. Ljunggren S, Eriksson JC. The lifetime of a colloid-sized gas bubble in water and the cause of hydrophobic attraction. *Colloid. Surface. A.* 1997;129-130:151-155.
24. Leu HG, Lin SH, Shyu CC, Lin CM. Effects of surfactants and suspended solids on oxygen transfer under various operating conditions. *Environ. Technol.* 1998;19:299-306.
25. Takahashi M, Li P. Base and technological application of micro-bubble and nanobubble. *J. Phys. Chem.* 2009;22:2-19.
26. Ghadimkhani A, Zhang W, Marhaba T. Ceramic membrane defouling (cleaning) by air nano bubbles. *Chemosphere* 2016;146:379-384.
27. Riser-Roberts E. Remediation of petroleum contaminated soils; biological, physical and chemical processes. USA: Lewis publishers; 1998.
28. Li H, Hu L, Xia Z. Impact of groundwater salinity on bioremediation enhanced by micro-nano bubbles. *Materials* 2013;6:3676-3687.