Toxic Effects on the Nonspecific Immune System of the Rock Bream *Oplegnathus fasciatus* upon Exposure to Di-2-ethylhexyl Phthalate

Jun-Hwan Kim¹, Dal-Sang Jeong² and Ju-Chan Kang^{1*}

¹Department of Aquatic Life Medicine, Pukyong National University, Busan 608-737, Korea ²Korea National College of Agriculture and Fisheries, Hwaseong 445-760, Korea

Abstract

The aim of this study was to investigate the *in vivo* toxicity of di-2-ethylhexyl phthalate (DEHP), on the immune system of the rock bream, *Oplegnathus fasciatus*. Fish were injected twice intraperitoneally with DEHP (200, 400, and 800 mg/kg BW), and the cellularity and functional activity of phagocytes in the spleen and head kidney were measured. The number of head kidney leukocyte cells was significantly greater in fish treated with 800-mg DEHP/kg BW. Nonspecific immunity, as determined by the phagocytic index, was significantly decreased at 800-mg DEHP/kg BW in the head kidney. A significant reduction in phagocytic capacity was observed in the head kidney at \geq 400-mg DEHP/kg BW. Furthermore, significantly increased levels of serum glutamic oxaloacetate transaminase and glutamic pyruvate transaminase indicated a marked hepatic dysfunction in immunosuppressed fish. Total serum protein was significantly reduced at \geq 400-mg DEHP/kg BW; however, there were no significant changes in lysozyme activity. These results demonstrate that immune responses in the rock bream, *Oplegnathus fasciatus* can predict immunotoxicity at doses \geq 400-mg DEHP/kg BW.

Key words: Oplegnathus fasciatus, Rock bream, Di-2-ethylhexyl phthalate, Immune system

Introduction

Phthalate esters (PEs) are a group of organic chemicals used as plasticizers to increase the flexibility and durability of plastics. The commercial compound di-2-ethylhexyl phthalate (DEHP) has various industrial uses. This compound has been identified in industrial wastewaters, fish, and other aquatic organisms (Scholz et al., 1997; Sonnenschein and Soto, 1998; Acey et al., 2002; Chang et al., 2005). The frequency of exposure to these compounds and their potential effects on the reproduction of aquatic animals has stimulated research into their environmental distribution, bioaccumulation, fate and mechanisms of action (Staples et al., 1997; Kim et al., 2002). These esters are also suspected to be endocrine disruptors and to mimic estrogen (Sonnenschein and Soto, 1998; Shioda

and Wakabayashi, 2000; Tollefsen, 2002). As phthalic acidderived compounds, PEs have been shown to induce estrogen-receptor-mediated responses (Jobling et al., 1995). Stress caused by environmental variables can affect the endocrine system, which in turn, through bi-directional communication, can affect other intimately linked systems, such as the immune and nervous systems, and prevent the maintenance of homeostasis (Balm, 1997).

The immune systems of fish are similar to those of higher vertebrates (Nakanishi et al., 2002). The thymus, spleen and kidney (especially the head kidney) serve mainly as immune organs (Romano et al., 1997). A healthy immune system is essential to protect an organism from infections and diseases

Open Access http://dx.doi.org/10.5657/FAS.2013.0171

This is an Open Access article distributed under the terms of the Creative Commons Attribution Non-Commercial License (http://creativecommons. org/licenses/by-nc/3.0/) which permits unrestricted non-commercial use, distribution, and reproduction in any medium, provided the original work is properly cited. pISSN: 2234-1749 eISSN: 2234-1757 Received 8 May 2013; Revised 20 June 2013 Accepted 19 July 2013

*Corresponding Author E-mail: jckang@pknu.ac.kr and to maintain homeostasis and general good health. Mucus secretions on the gill and digestive tract surfaces in fish act as a first defense and protect them from the invasion of microorganisms, because the surface of the fish body is exposed directly to viruses and bacteria in the surrounding water (Braun et al., 1990). Non-specific biodefense functions, including the phagocytic activity of leucocytes and other components such as lysozymes and complementary systems in fish respond to rapidly remove various xenobiotics, which are recognized as foreign bodies (Bennani et al., 1995).

Various chemicals are released and accumulate in the aquatic environment, and coastal areas become a 'sink' for many of them. These chemicals are linked to various adverse effects, including immunotoxicity, in fish (Köllner at al., 2002). However, little is known regarding the effects of chemicals on the immune system, such as the function of leucocytes in fish exposed to PEs *in vivo*. The aim of this study was to investigate the *in vivo* toxicity of DEHP on the immune system of the rock bream, *Oplegnathus fasciatus*. Serum total protein, glutamic oxaloacetate transaminase (GOT) and glutamic pyruvate transaminase (GPT) levels were also measured as indicators of disturbed metabolic functions.

Materials and Methods

Experimental fish

Rock bream, *Oplegnathus fasciatus* weighing on average 46.5 ± 1.02 g, were obtained from a local fish farm in Gyeongsangnam-do, Korea. Fish were held for acclimatization for 2 weeks and to their overall health under laboratory conditions ($18.2 \pm 0.52^{\circ}$ C) prior to exposure. During the acclimatization period, fish were fed a commercial diet twice daily and maintained on a 12:12h light/dark cycle.

Exposure conditions

The fish were exposed to DHEP in 0.5-ton fiberglass-reinforced plastic tanks; each treatment group comprised 20 fish. Each tank received a flow of 7 L/min with continuous aeration. DEHP was purchased from Sigma (St. Louis, MO, USA). DEHP was dissolved in sunflower seed oil (SSO) immediately before intraperitoneal injection. The fish were injected with doses of 200-, 400- and 800-mg DEHP/kg BW. The first injection was given 2 weeks after acclimatization and the second was given 1 week later. The control group was subjected to an identical regimen; however, they were injected with an equal volume of SSO. Blood and lymphoid tissue samples were taken for assessment of blood, spleen and head kidney parameters at 1 and 2 weeks post injection.

Isolation of leucocytes

Leucocytes were isolated from the spleen and head kidney using a modification of the method described by Fatima et al. (2001). Single-cell suspensions were prepared by dissociating the lymphoid tissues, using a cell dissociation sieve-tissue grinder kit (Sigma), in L-15 medium supplemented with 0.1% fetal bovine serum (FBS), 1% streptomycin/penicillin solution (S/P; Gibco, Rockville, MD, USA) and 10 U/mL heparin (Sigma). The resulting suspensions were purified in Percoll (Sigma) density gradients (34/51%), centrifuged at 400 g at 4°C for 30 min, and the buffy coat on the surface was harvested into silicone tubes.

Lymphoid organ cellularity

Cells collected from the interface were washed twice by centrifugation (1,000 g at 4°C for 10 min) in non-supplemented L-15 medium. The final cell pellet was resuspended in 1 mL of L-15 medium (pH 7.2) supplemented with 0.1% FBS, 1% S/P (Gibco) and 10 U/mL heparin (Sigma). Cells were enumerated using a hemocytometer. Cell viability was determined using the trypan blue dye exclusion method. The number of cells was expressed as millions per mL. The leukocyte populations in the spleen and head kidney cell suspensions were determined. Cell numbers were adjusted to the concentration (1×10^7 /mL) appropriate for the assay.

Assay of phagocytic activity

Phagocytic activity was evaluated using a cell suspension, as described by Ahmad et al. (1998). An 0.1-mL aliquot with a cell density of 1×10^{7} /mL in L-15 medium was mixed with an equal volume of L-15 medium containing 20% FBS and 1×10^{8} /mL heat-treated (100°C for 1 h) yeast cells (Saccharomyces cerevisiae). Phagocytosis was allowed to proceed for 1 h at 35°C with occasional shaking. After incubation, 50 µL of this mixture were smeared onto glass slides, air-dried, fixed in methanol and stained with Diff-Quik solution (Sysmex, Kobe, Japan). The slides were air dried and observed under Malinol (MPC, Tokyo, Japan). The average number of yeast cells engulfed per phagocyte was determined by inspecting 500 phagocytic cells in each sample. Phagocytic activity was expressed using the phagocytic index (PI) and as a percentage of the phagocytic capacity (PC). The PI was calculated as:

$PI = A \times B$

, where A = the percentage of phagocytes engulfing at least two yeast cells and B = the average number of yeast cells engulfed by phagocytosis-positive cells. PC is the mean percentage of cells that are engulfed (Bin-Hafeez et al., 2003).

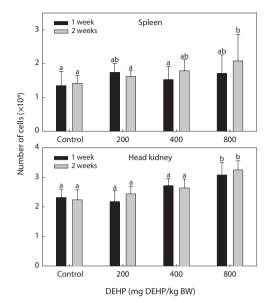


Fig. 1. Mean cellularity in the spleen and head kidney of rock bream, *Oplegnathus fasciatus* exposed to various di-2-ethylhexyl phthalate (DEHP). All data are expressed as mean \pm SE. Different letters significant difference (P < 0.05) between the groups.

Serum analysis

The lysozyme assay was performed according to Ellis (1990), with some modifications. First, 0.2 mg/mL *Micrococcus lysodeikticus* (Sigma) was suspended in 0.05 M sodium-phosphate buffer, pH 6.2. Then, 50 μ L of serum were added to 950 μ L of the bacterial suspension, and the absorbance at 450 nm was measured at 30-s intervals for 3 min at 25°C. One unit (U) of lysozyme activity was defined as the amount of sample causing a reduction in absorbance of 0.001/min. The total protein estimation and the GOT and GPT assays were performed using a diagnostic kit and reagents supplied by Asan Pharm. Co., Ltd. (Hwaseong, Korea).

Statistical analysis

The experiment was repeated three times. Statistical analysis of the results was performed using the SPSS/PC+ statistical package (SPSS Inc., Chicago, IL, USA). ANOVA and Duncan's test for multiple comparisons were used to identify significant differences between the control and treatment groups (Duncan, 1955). The significance level was set at P < 0.05.

Results

Splenic and pronephric cellularity

The total numbers of single cells in the spleen and head

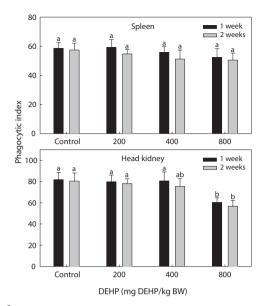


Fig. 2. Phagocytic index in the spleen and head kidney of rock bream, *Oplegnathus fasciatus* exposed to various di-2-ethylhexyl phthalate (DEHP). All data are expressed as mean \pm SE. Different letters significant difference (P < 0.05) between the groups.

kidney of fish exposed to different levels of DEHP are shown in Fig. 1. The number of cells in the head kidney was significantly higher in the fish exposed to 800-mg DEHP/kg than in the control group (P < 0.05). However, no significant difference was observed in cell numbers in the spleens of DEHPtreated and control fish.

Phagocytic functional responses

The dose-dependent effects on the nonspecific immune response and the phagocytic effects of DEHP in rock bream are illustrated in Figs. 2 and 3. The nonspecific immunity of rock bream decreased markedly upon application of 800-mg DEHP/kg BW, as indicated by the decreased PI in the head kidney (P < 0.05). The phagocytic efficiency differed between the control and treatment groups, with a reduction in capacity, especially in head kidney leukocytes, upon treatment with \geq 400-mg DEHP/kg BW, suggesting that this compound has immunosuppressive activity (P < 0.05).

Lysozyme activity

The lysozyme activity in the serum of rock bream exposed to DEHP is presented in Fig. 4. The DEHP decreased the serum lysozyme activity in rock bream in a dose-dependent manner; however, the activity in the treated groups was not significantly different from that in the carrier-injected control group.

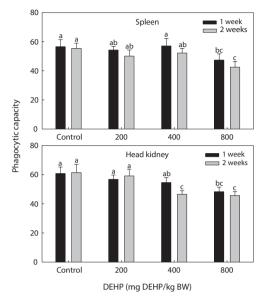


Fig. 3. Phagocytic capacity in the spleen and head kidney of rock bream, *Oplegnathus fasciatus* exposed to various di-2-ethylhexyl phthalate (DEHP). All data are expressed as mean \pm SE. Different letters significant difference (P < 0.05) between the groups.

Metabolic indicators

The serum total protein concentration of rock bream treated with DEHP is shown in Fig. 4. The total protein concentration in rock bream exposed to \geq 400 mg DEHP/kg BW DEHP was significantly suppressed compared to the control (P < 0.05). Moreover, GOT and GPT activity increased significantly upon application of 800-mg DEHP/kg BW (P < 0.05) (Fig. 4).

Discussion

The immune systems of fish are similar to those of higher vertebrates, and tissues such as the thymus, spleen and kidney (especially the head kidney) function mainly as immune organs. Healthy immune systems are essential to prevent infection and diseases in fish and to maintain homeostasis and general good health. Markers of nonspecific immunity appear to be successful indicators of xenobiotic-induced stress in laboratory fish. Considerable evidence now supports links between environmental changes (including exposure to contaminants) and noninfectious diseases resulting from immunosuppression (Zelikoff et al., 2000; Fatima et al., 2001). The total number of leucocytes in fish hematopoietic organs is considered to be a more sensitive indicator of chemical toxicity

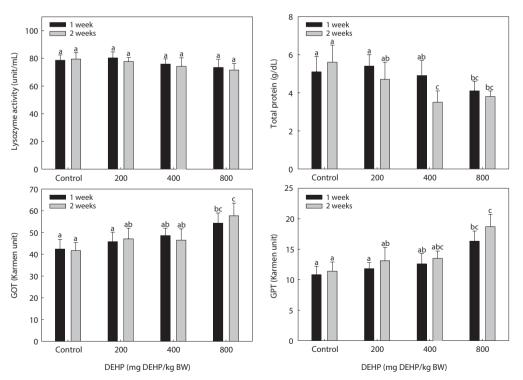


Fig. 4. Levels of lysozyme, total protein, glutamic oxaloacetate transaminase (GOT) and glutamic pyruvate transaminase (GPT) in the serum of rock bream, *Oplegnathus fasciatus* exposed to various di-2-ethylhexyl phthalate (DEHP). All data are expressed as mean \pm SE. Different letters significant difference (P < 0.05) between the groups.

than the activity of phagocytic cells contained within these organs (Hart et al., 1998). In lymphoid organs, high doses of DEHP can cause marked cellularity in the head kidney. The hypercellularity observed in the fish head kidney associated with treatment with 800-mg DEHP/kg BW is in agreement with the dose-dependent induction of thymocyte populations in mice by dioctyl phthalate (Dogra et al., 1993).

DEHP had an immunomodulatory effect on the phagocytic efficiency of rock bream leukocytes in vivo. High doses of DEHP had an immunostimulatory effect on the PI in the head kidney phagocytes. A significant reduction in the PI was observed in the head kidney at the highest DEHP concentration (800 mg/kg BW). A dose-dependent response of phagocytic activity has also been reported in fish treated with gonadal steroids (Law et al., 2001; Yamaguchi et al., 2001). Immunosuppressive effects, in terms of PC, were observed in the spleen and head kidney following treatment with the highest dose of DEHP. In the present study, ≥400-mg DEHP/kg BW particularly influenced the killing activity of head kidney leukocytes and contributed to an affinity for the plasma sex-steroid-binding PE protein. PEs have been identified as xenoestrogenic compounds that mimic endogenous estrogens and exert direct effects on cells via estrogen receptors (Tollefsen, 2002). Watanuki et al. (2003) reported the expression levels of estrogen receptors by brain, liver, and fish kidney leucocytes. The in vivo results presented here revealed immunosuppression in exposed fish, especially those treated with the highest dose of DEHP (800 mg/kg BW). The reductions in nonspecific immune responses might be caused by the release of hormones such as cortisol, progesterone, and testosterone, which exert immunosuppressive effects by directly affecting leukocytes via androgen receptors (Slater et al., 1995; Yamaguchi et al., 2001).

The dose-dependent reduction in total protein and immunomodulation by DEHP suggested a disturbed physiological mechanism and a specific, receptor-mediated induction by PEs. The high GOT and GPT activity in rock bream treated with 800-mg DEHP/kg BW indicated a greater degree of hepatic dysfunction in immunosuppressed fish. The different responses of the immune system in spleen and head kidney leukocytes might have resulted from differences in the toxicity and distribution pattern of PEs in the organs of fish, as suggested by Menghi et al. (2002). Furthermore, the lack of significant changes in GOT and GPT activity also suggested normal hepatic function in fish treated with low DHEP concentrations. We conclude that exposure of the rock bream Oplegnathus fasciatus to a high concentration (>400 mg/kg BW) of DEHP results in significant alterations in a number of parameters of the nonspecific immune system.

References

Acey RA, Bailey S, Healy P, Jo C, Unger TF and Hudson RA. 2002. A

butyrylcholinesterase in the early development of the brine shrimp (*Artemia salina*) larvae: a target for phthalate esters embryo embryotoxicity? Biochem Biophy Res Commun 299, 659-662.

- Ahmad I, Fatima, M Athar M, Khan NZ and Raisuddin S. 1998. Responses of circulating fish phagocytes to paper mill effluent exposure. Bull Environ Contam Toxicol 61, 746-753. http://dx.doi. org/10.1007/s001289900824.
- Balm PHM. 1997. Immune-endocrine interactions. In: Fish Stress and Health in Aquaculture. Iwama GK, Pickering AD, Sumpter JP and Schreck CB, eds. Cambridge University Press, Cambridge, GB, pp. 195-221.
- Bennani N, Schmid-Alliana A and Lafaurie M. 1995. Evaluation of phagocytic activity in a teleost fish, *Dicentrarchus labrax*. Fish Shellfish Immunol 5, 237-246. http://dx.doi.org/10.1016/S1050-4648(05)80017-8.
- Bin-Hafeez B, Haque R, Parvez S, Pandey S, Sayeed I and Raisuddin S. 2003. Immunomodulatory effects of fenugreek (*Trigonella foenum graecum* L.) extract in mice. Int Immunopharmacol 3, 257-265. http://dx.doi.org/10.1016/S1567-5769(02)00292-8.
- Braun R, Arnesen JA, Rinne A and Hjelmeland K. 1990. Immunohistological localization of trypsin in mucus-secreting cell layers of Atlantic salmon, *Salmo salar* L. J Fish Dis 13, 233-238. http://dx.doi. org/10.1111/j.1365-2761.1990.tb00778.x.
- Chang BV, Liao CS and Yuan SY. 2005. Anaerobic degradation of diethyl phthalate, di-*n* butyl phthalate, and di-(2-ethyhexyl) phthalate from river sediment in Taiwan. Chemosphere 58, 1601-1607. http://dx.doi.org/10.1016/j.chemosphere.2004.11.031.
- Dogra RK, Khanna S, Srivastava SN, Shukla LJ, Chandra K, Saxena G and Shanker R. 1993. Immunomodulation due to coexposure to styrene and dioctyl phthalate in mice. Immunopharmacol Immunotoxicol 15, 491-514. http://dx.doi.org/10.3109/08923979309035242.
- Ellis AE. 1990. Lysozyme assays. In: Techniques in Fish Immunology. Vol. I. Stolen JS, Anderson DP, Roberson BS and van Muiswinkel WB, eds. SOS Publications, New Haven, NJ, US, pp. 101-103.
- Fatima M, Ahmad I, Siddiqui R and Raisuddin S. 2001. Paper and pulp mill effluent-induced immunotoxicity in freshwater fish *Channa punctatus* (Bloch). Arch Environ Contam Toxicol 40, 271-276.
- Hart LJ, Smith SA, Smith BJ, Robertson J, Besteman EG and Holladay SD. 1998. Subacute immunotoxic effects of the polycyclic aromatic hydrocarbon 7,12-dimethylbenzanthracene (DMBA) on spleen and pronephros leukocytic cell counts and phagocytic cell activity in tilapia (*Oreochromis niloticus*). Aquat Toxicol 41, 17-29. http:// dx.doi.org/10.1016/S0166-445X(97)00075-1.
- Jobling S, Reynolds T, White R, Parker MG and Sumpter JP. 1995. A variety of environmentally persistent chemicals, including some phthalate plasticizers, are weakly estrogenic. Environ Health Perspect 103, 582-587.
- Kim EJ, Kim JW and Lee SK. 2002. Inhibition of oocyte development in Japanese medaka (*Oryzias latipes*) exposed to di-2-ethylhexyl phthalate. Environ Int 28, 359-365. http://dx.doi.org/10.1016/ S0160-4120(02)00058-2.
- Köllner B, Wasserrab B, Kotterba G and Fischer U. 2002. Evaluation of immune functions of rainbow trout (*Oncorhynchus mykiss*): how can environmental influences be detected? Toxicol Lett 131, 83-

95. http://dx.doi.org/10.1016/S0378-4274(02)00044-9.

- Law WY, Chen WH, Song YL, Dufour S and Chang CF. 2001. Differential *in vitro* suppressive effects of steroids on leukocyte phagocytosis in two teleosts, tilapia and common carp. Gen Comp Endocrinol 121, 163-172. http://dx.doi.org/10.1006/gcen.2000.7593.
- Menghi G, Sabbieti MG, Menghi M, Roda A, Materazzi S and Marchetti L. 2002. Immunohistochemical detection of phthalate esters in the alimentary canal of *Tilapia* spp. J Fish Biol 61, 265-271. http:// dx.doi.org/10.1111/j.1095-8649.2002.tb01751.x.
- Nakanishi T, Fischer U, Dijkstra JM, Hasegawa S, Somamoto T, Okamota N and Ototake M. 2002. Cytotoxic T cell function in fish. Dev Comp Immunol 26, 131-139. http://dx.doi.org/10.1016/S0 145-305X(01)00055-6.
- Romano N, Taverne-Thiele JJ, Van Maanen JC and Rombout JHMW. 1997. Leucocyte subpopulations in developing carp (*Cyprinus carpio* L.): immunocytochemical studies. Fish Shellfish Immunol 7, 439-453.
- Scholz N, Diefenbach R, Rademacher I and Linnemann D. 1997. Biodegradation of DEHP, DBP, and DINP: poorly water soluble and widely used phthalate plasticizers. Bull Environ Contam Toxicol 58, 527-534. http://dx.doi.org/10.1007/s001289900367.
- Shioda T and Wakabayashi M. 2000. Effect of certain chemicals on the reproduction of medaka (*Oryzias latipes*). Chemos 40, 239-243. http://dx.doi.org/10.1016/S0045-6535(99)00235-0.

Slater CH, Fitzpatrick MS and Schreck CB. 1995. Characterization of

an androgen receptor in salmonid lymphocytes: possible link to androgen-induced immunosuppression. Gen Comp Endocrinol 100, 218-225. http://dx.doi.org/10.1006/gcen.1995.1151.

- Sonnenschein C and Soto AM. 1998. An updated review of environmental estrogen and androgen mimics and antagonists. Mol Biol 65, 143-150. http://dx.doi.org/10.1016/S0960-0760(98)00027-2.
- Staples CA, Adams WJ, Parkerton TF, Gorsuch JW, Biddinger GR and Reinert KH. 1997. Aquatic toxicity of eighteen phthalate esters. Environ Toxicol Chem 16, 875-891. http://dx.doi.org/10.1002/ etc.5620160507.
- Tollefsen KE. 2002. Interaction of estrogen mimics, singly and in combination, with plasma sex steroid-binding proteins in rainbow trout (*Oncorhynchus mykiss*). Aquat Toxicol 56, 215-225. http://dx.doi. org/10.1016/S0166-445X(01)00154-0.
- Watanuki H, Gushiken Y and Sakai M. 2003. *In vitro* modulation of common carp (*Cyprinus carpio* L.) phagocytic cells by Di-n-butyl phthalate and Di-2-ethylhexyl phthalate. Aquat Toxicol 63, 119-126. http://dx.doi.org/10.1016/S0166-445X(02)00172-8.
- Yamaguchi T, Hironobu W and Sakai M. 2001. Effects of estradiol, progesterone and testosterone on the function of carp, *Cyprinus carpio*, phagocytes *in vitro*. Comp Biochem Physiol C Toxicol Pharmacol 129, 49-55. http://dx.doi.org/10.1016/S1532-0456(01)00176-4.
- Zelikoff JT, Raymond A, Carlson E, Li Y, Beaman JR and Anderson M. 2000. Biomarkers of immunotoxicity in fish: from the lab to the ocean. Toxicol Lett 112-113, 325-331.