

Evaluation of Badge-Type Diffusive Sampler Performance for Measuring Indoor Formaldehyde

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Abstract

The purposes of this study were to determine the efficiency of using a badge-type diffusive sampler to measure formaldehyde concentrations indoors, and to evaluate the uncertainty associated with the use of data from a diffusive sampler. A diffusive sampler using 2,4-dinitrophenylhydrazine (DNPH) reagent was found to be a suitable tool for measuring the formaldehyde concentration in an indoor environment. The agreement between results of the diffusive sampler and DNPH cartridge were good, showing a correlation coefficient of 0.996. The sampling rate for the diffusive sampler was calculated to be 1.428 L hr⁻¹, with a standard deviation of 0.084 L hr⁻¹. It was found through analysis that the uncertainty associated with the sampling rate and the mass of the formaldehyde transported into the diffusive sampler by diffusion was the dominant contributor to the total.

Keywords: Diffusive monitor, Formaldehyde, Indoor pollution, Passive sampler, Uncertainty

1. Introduction

Formaldehyde is a colorless and highly flammable gas that has a pungent odor. It is used in the preparation of urea-formaldehyde and phenol-formaldehyde resins as adhesives in the manufacture of particle board, fiberboard, and plywood. Such resins are major sources of formaldehyde in indoor environments. Formaldehyde is a pollutant whose presence in indoor environments has caused much concern because of the potential harm it causes to human health [1, 2].

Formaldehyde concentrations in indoor environments have been measured in workplaces, offices, residential buildings, and public spaces. Reportedly, formaldehyde is present in various indoor environments in mean concentrations ranging from 4 to 70 $\mu\text{g m}^{-3}$ [3-6]. In addition, formaldehyde concentrations in indoor environments have been usually found to be greater than those in outdoor environments [4, 7, 8]. For this reason, the measurement of formaldehyde concentrations in indoor environments is of particular interest. The guidelines for residential buildings and public spaces in Korea specify that the concentration of formaldehyde in these places should not exceed 100 $\mu\text{g m}^{-3}$ [9].

Methods for measuring formaldehyde concentrations in indoor and ambient atmospheres have been developed by using both active and passive sampling techniques, employing the reaction of formaldehyde with 2,4-dinitrophenylhydrazine (DNPH) reagent, 3-methyl-2-benzothiazolinone hydrazone hydrochloride reagent, or chromotropic acid [6, 10-13]. Diffusive

sampling techniques have often been used to measure the concentrations of air pollutants in indoor and ambient atmospheres because of its convenience with regard to deployment and collection. Because the technique requires no power supply during sampling, it can be employed almost anywhere. Moreover, the long-term effective concentrations obtained by the diffusive sampling technique are very useful in exposure assessments. However, collecting real-time concentration data of pollutants is not possible [14, 15].

In practice, the uncertainty may arise in the analytical procedure associated with the use of a diffusive sampler, such as sampling, instrument effects, reference values, blank correction, and operator effects [16]. The estimation of the uncertainty is necessary in order to establish the comparability of the measurement results and to improve the confidence in the validity of the results obtained. The evaluation of the uncertainty associated with the analytical method using a thermally desorbable diffusive sampler for measuring benzene was reported by Plaisance et al. [17]. The principal part of the uncertainty budget was accounted for by the variation of the sampling rate in relation to the environmental factors.

In this study, a diffusive sampling technique, using DNPH for determining the formaldehyde concentration in indoor environments, was evaluated. Furthermore, the uncertainty with regard to the collection and analysis of formaldehyde data in such a manner was estimated.

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2. Materials and Methods

2.1. Formaldehyde Diffusive Sampler and Chemicals

The badge-type diffusive sampler comprises a sampler end cap (55 mm × 55 mm × 18 mm), sampler body (36-mm i.d.), and a diffusion screen (33-mm i.d.) made of a high-density polyethylene [11]. The sampler body comprises a filter-retaining ring (28-mm i.d.) and a spacer ring (28-mm i.d.), which maintains a diffusion distance of 10 mm, a semi-permeable membrane (Celgard-2500; Celgard LLC, Charlotte, NC, USA), and a collection filter (26-mm, 1 Chr; Whatman, Maidstone, Kent, UK). The filter-retaining ring and spacer ring are made of polyethylene. The semi-permeable membrane has 55% porosity, $0.209 \mu\text{m} \times 0.054 \mu\text{m}$ pore size, and 25- μm thickness. The silicone flat ring (28-mm i.d.), which ensures an airtight connection between the diffusion screen and the sampler body, is mounted on the diffusion screen.

For the experiments, all parts of the diffusive sampler, except the semi-permeable membrane, were thoroughly cleaned with pure water and acetonitrile, using an ultrasonic cleaner (40 kHz; Seong Dong, Seoul, Korea), and were then allowed to dry overnight in a dry oven at 60°C (OF-22; JeioTech, Daejeon, Korea). After drying and cooling, the components were stored in a clean environment to prevent the accumulation of dust or other contaminants.

The collection filter was soaked in an absorbing solution containing 0.1 g of DNPH-HCl ethanol recrystallized twice from DNPH (97%; Sigma-Aldrich, St. Louis, MO, USA), 4 M HCl (35%–37%; Wako, Osaka, Japan) and 0.3 mL of phosphoric acid (85%, Wako) in 20 mL of HPLC-grade acetonitrile (Wako). The coated collection filter was dried for 24 hr under vacuum in a desiccator. After drying the absorbing solution, the components were immediately assembled into the diffusive sampler. The samplers were handled very carefully because they are easily contaminated. All the samplers were stored in sealed aluminum bags in the dark at 4°C before and after use.

2.2. Analytical Procedures

In order to analyze the collected formaldehyde, the collection filter of the diffusive sampler was unloaded into a glass bottle with PTFE-lined screw caps for extracting the formed DNPH derivative; the derivative was extracted with 3 mL of acetonitrile in a shaker (BS-21, JeioTech) having a controlled-temperature incubator ($\pm 1^\circ\text{C}$) with orbital shaking for 1 hr. The agitation in the shaker was adjusted to produce a mild swirling of the extraction solution, and the rotational frequency was about 150 rpm. The temperature of the solution in the extraction vials during the extraction was maintained at about 30°C in order to allow for a stable extraction.

The HCHO-DNPH derivative extracted the diffusive sampler and the DNPH cartridge was analyzed using a high performance liquid chromatograph, equipped with a UV photodetector (UVD-170U; Dionex, Sunnyvale, CA, USA) set at 360 nm, with a C-18 column (4.6 × 250 mm, 5 μm , 120 Å; Acclaim 120, Dionex) used for separation of the HCHO-DNPH derivative, with a gradient elution of acetonitrile-pure water from 50:50(v/v) to 100:0, at a flow rate of 1.0 mL min⁻¹. The total time required for the analysis of HCHO was 15 min, with the retention time of the HCHO-DNPH derivative being 11.4 min.

Quantitation was performed using an aldehyde/ketone-

DNPH derivatives standard solution (15 μg aldehyde mL⁻¹; Restek, Bellefonte, PA, USA). Formaldehyde peaks were quantified from a regression line fitted to a set of external standards ranging from 0.15 to 15 μg mL⁻¹. The detector response (peak area) was linear over this range, as determined by the regression coefficients, with values greater than 0.99 for formaldehyde.

Three blank samplers were prepared and analyzed together with the exposed samplers. The analyzed values of the blank samplers were subtracted from those of the exposed sampler values to obtain the net formaldehyde concentration. All experiments were carried out at least in duplicate.

2.3. Optimization of the Diffusive Sampling

The diffusive samplers were tested in a cylindrical exposure chamber (0.07 m³), which employed a mixing fan and sensors (TSI model 8386; TSI Inc., Shoreview, MN, USA) for monitoring the temperature and relative humidity. The formaldehyde was generated from formaldehyde solution (38% in water, Sigma-Aldrich) by introduction of pure air (flow rate, 1 L min⁻¹). The diluted formaldehyde was introduced into the chamber at a flow rate of 5 L min⁻¹ and its concentration was measured using the DNPH cartridge.

To determine the sampling rate of the diffusive sampler, a DNPH cartridge (capacity, 75 μg /cartridge; background, <0.06 μg formaldehyde/cartridge; Supelco, Bellefonte, PA, USA) was employed for active sampling. Diffusive and active sampling methods were carried out simultaneously in the exposure chamber to estimate the sampling rate. The exposure experiment, using four replicate diffusive samplers in the chamber, was performed. The temperature and relative humidity in the chamber throughout the exposure period were held constant at $25 \pm 3^\circ\text{C}$ and $40 \pm 10\%$, respectively. Four samplers were installed in the chamber, and exposed for periods of between 3 and 24 hr. All experiments were carried out at least in duplicate.

The optimum quantity of DNPH as a collection medium should result in the sensitive detection of formaldehyde, while remaining applicable to long exposure times. Furthermore, the material should exhibit low blank levels. The collection filter is completely covered by the absorption solution. To obtain the optimum dipping time for the absorption of DNPH onto the collection filter, the collection filter was dipped in DNPH solution for around 0.5 to 24 hr. The diffusive samplers assembled with these filters were exposed to an average formaldehyde concentration of $250 \pm 50 \mu\text{g m}^{-3}$ in the chamber. The sampling was performed for 6 to 24 hr. The blank value of the diffusive sampler was also estimated. The collection filter coated with the absorbing solution was immediately extracted in acetonitrile and analyzed by the procedure already described.

The relationship between the collected formaldehyde mass and the dipping time of the collection filter was investigated by carrying out a series of exposure experiments using a formaldehyde concentration of about $250 \mu\text{g m}^{-3}$. The collected mass increased with dipping time up to about 4 hr. Subsequently, however, there was no further increase in the collected mass. The mean relative standard deviation for the total data was about 13% (that for 4 hr of dipping time was 8%). The variation of the analyzed peak area of DNPH as a function of the dipping time was the same as that of the collected mass of formaldehyde.

The blank samplers were used to evaluate the contamination of samplers, as well as to estimate the detection limits and errors for the diffusive samplers. The blank values had a mean form-

aldehyde mass of $0.028 \pm 0.003 \mu\text{g}$ with a relative standard deviation of 10% ($n = 31$), and they showed a normal distribution. The blank values did not increase after four weeks of storage in a refrigerator at 4°C in the dark.

2.4. Uncertainty Sources

In this study, the formaldehyde concentration in the sample, expressed in $\mu\text{g m}^{-3}$, can be given as follows:

$$\text{Concentration} = \frac{1,000 Q_{\text{DS}}/t}{\text{SR}} \quad (1)$$

where, Q_{DS} is the mass of formaldehyde collected by the diffusive sampler (μg), t is the sampling time (hr), SR is the sampling rate (L hr^{-1}) and the scale factor of 1,000 converts liters into cubic meters.

The estimation of the overall uncertainty of an analytical result should consider all sources of systematic and random errors that are associated with the applied measurement and analytical method [16]. Errors in the determination of formaldehyde concentration using the diffusive sampler are mainly affected by the following sources: 1) standard preparation, calibration, extraction and analysis ($u^2(Q_{\text{DS}})$); 2) sampling time ($u^2(t)$); and 3) determination of the sampling rate ($u^2(\text{SR})$).

The overall combined uncertainty ($u(\text{CON})$) of the analytical result can be calculated as follows:

$$u(\text{CON}) = \sqrt{u^2(Q_{\text{DS}}) + u^2(\text{SR}) + u^2(t)} \quad (2)$$

The formaldehyde mass (Q_{DS}) collected in the diffusive sampler is calculated as follows:

$$Q_{\text{DS}} = C_{\text{stock}} \cdot \frac{V_s}{(V_s + V_d)} \cdot \frac{(R_x - R_b)}{R_s} \cdot V_x \cdot \frac{M_{\text{HCHO}}}{M_{\text{DNPH}}} \cdot f_{\text{lin}} \quad (3)$$

where C_{stock} is the concentration of the standard solution ($\mu\text{g mL}^{-1}$), V_s is the volume of the standard solution (mL), V_d is the volume of a dilute solution (mL), R_x is the analytical response value of a sample, R_b is the analytical response value of a blank, R_s is the analytical response value of the standard solution, V_x is the extraction volume for the collection filter in the diffusive sampler (mL), M_{HCHO} is the molar mass of formaldehyde, M_{DNPH} is the molar mass of HCHO-DNPH derivative, and f_{lin} is the slope of the calibration curve.

The sampling of formaldehyde in the chamber was simultaneously performed with a DNPH cartridge and pump as an active sampling method to determine the sampling rate (SR) of the diffusive sampler. The exposure experiment in the chamber was performed with concentrations ranging from 15 to $480 \mu\text{g m}^{-3}$. Four diffusive samplers were installed in the chamber and exposed for periods of 3 to 24 hr.

The collected mass of formaldehyde was corrected by subtracting the average mass measured for the blanks. The reference formaldehyde concentration was determined from the mass of formaldehyde collected by the DNPH cartridge in relation to the volume of air drawn by the pump. The sampling rate of the diffusive sampler can be calculated as follows:

$$\text{Sampling rate (L hr}^{-1}\text{)} = \frac{Q_{\text{DS}}}{C_{\text{HCHO}} \cdot t} \quad (4)$$

where Q_{DS} is the mass of the formaldehyde transported into the diffusive sampler by diffusion (μg), t is the exposure time for the diffusive sampler (hr), and C_{HCHO} is the formaldehyde concentra-

tion ($\mu\text{g m}^{-3}$) measured by the DNPH cartridge with active sampling.

3. Results and Discussion

3.1. Estimation of the Standard Uncertainty

The uncertainty associated with the purity of the standard solution (HCHO-DNPH derivative), $u(C_{\text{stock}})$, was estimated based on the difference of the purity level stated by the manufacturer. The concentration and purity of the standard solution were stated to be $15 \mu\text{g mL}^{-1}$ and 99%, respectively. The standard uncertainty for the purity of the standard solution was assumed to have a rectangular distribution and was divided by $\sqrt{3}$ [16], because the manufacturer gave no further information concerning the uncertainty value.

$$u(C_{\text{stock}}) = \frac{(1-\text{purity})/2}{\sqrt{3}} \cdot C_{\text{stock}} \quad (5)$$

The standard uncertainty for the concentration of the standard solution associated with the purity was $0.0433 \mu\text{g mL}^{-1}$.

The uncertainties associated with the volume measurement, $u(V_s)$ and $u(V_d)$, exert the main influences like those on calibration and repeatability. The standard uncertainties of the volumetric calibration for 0.2 and 1.0 mL pipettes, $u(V_{\text{cal}})$ were assumed to have a triangular distribution and were divided by $\sqrt{6}$ [16]. The manufacturer's tolerance values for these pipettes were 0.1% (0.2 mL) and 0.13% (1 mL), respectively.

$$u(V_{\text{cal}}) = \frac{(T_{\text{pipette}}/100)}{\sqrt{6}} \cdot V_{\text{pipette}} \quad (6)$$

where T_{pipette} is the manufacturer-reported tolerance value of the pipette and V_{pipette} is the volume of solution taken with a pipette. The values of $u(V_{\text{cal}})$ for pipettes of 0.2 and 1.0 mL were 0.0816×10^{-3} and 0.4246×10^{-3} , respectively.

The uncertainty associated with volumetric repeatability, $u(V_{\text{rep}})$, was obtained from the repeated filling and weighting of the pipette volume. The standard deviations from the repeatability experiment for pipettes of 0.2 and 1.0 mL were 0.9428×10^{-3} mL and 1.5239×10^{-3} mL, respectively, and can be used directly as a standard uncertainty.

The standard uncertainties associated with the volume measurement for the dilution of the standard solution, $u(V_s)$ and $u(V_d)$, were calculated by simply finding the square root of the sum of the squares of the contributions of calibration and repeatability.

$$u(V_s) \text{ or } u(V_d) = \sqrt{u^2(V_{\text{cal}}) + u^2(V_{\text{rep}})} \quad (7)$$

The calculated standard uncertainties for the volume measurement of the standard and dilution solutions were 0.9463×10^{-3} and 1.5819×10^{-3} mL, respectively.

In order to evaluate the uncertainty associated with the repeatability of the analytical response values (e.g., peak area) obtained in the analytical process at the same concentration level, the standard uncertainties, $u(R_s)$, $u(R_b)$, and $u(R_x)$, were estimated by calculating the standard deviations of repeated samples. The standard deviations obtained from the repeated analysis of samples, blanks, and standards were 0.2614, 0.0338, and 0.6743, respectively.

$$u(R_x), u(R_b), \text{ or } u(R_s) = \frac{\sigma}{\sqrt{n}} \quad (8)$$

where σ is the standard deviation of response values for the samples, blanks, and standards. n is the number of their replicates analyzed in routine analysis ($n = 10$). The results obtained for each response value were 0.0827, 0.0107, and 0.2132, respectively.

A multiple-point calibration process is based on the linear regression curves obtained by plotting the peak area of a chromatogram versus the standard concentrations. Seven calibration standards were prepared from the HCHO-DNPH standard solution of $15 \mu\text{g mL}^{-1}$ and were measured nine times each. The calibration curve was given by (peak area) = $1.5787 + 5.1199 \times (\text{HCHO conc.})$ with a correlation coefficient of 0.997. The uncertainty associated with the determination of concentration by transforming the chromatographic signals in a calibration curve, $u(f_{\text{lin}})$, was defined as follows [18]:

$$u(f_{\text{lin}}) = \frac{1}{b} \cdot \sqrt{\sigma_{\text{resid}}^2 \cdot 1/n + (C_0 - C_{\text{mean}})^2 \cdot \sigma_b^2} \quad (9)$$

where b is the slope of the calibration curve, σ_{resid} is the standard deviation of residuals, n is the number of measurements for the calibration, C_0 is the formaldehyde concentration at each calibration level, C_{mean} is the mean value of the different calibration standards, and σ_b is the standard deviation of the slope of the calibration curve. The standard deviations of the slope and residuals obtained from the linear regression of each calibration curve were 0.1690 and 0.8084, respectively. When the formaldehyde concentration at each calibration level was $9 \mu\text{g mL}^{-1}$, the mean value of the different calibration standards (n , number of measurements) was determined to be $9.2932 \pm 0.3225 \mu\text{g mL}^{-1}$. The obtained result for the uncertainty in determining the concentration by the calibration curve was 0.0535.

The uncertainty associated with the final volume upon the extraction of the collection filter with acetonitrile, $u(V_x)$, has two major influences on calibration and repeatability. The volumetric calibration standard uncertainty of a 3 mL pipette, $u(V_{\text{cal}})$, was calculated to be 0.0016 mL, based on the manufacturer's tolerance value of 0.13%, and using the assumption of a triangular distribution, as shown in Eq. (6). The volumetric repeatability, $u(V_{\text{rep}})$, was obtained from the repeated filling and weighting of the pipette volume (3 mL). The standard deviations obtained from the repeatability experiment for $u(V_{\text{rep}})$ was 0.0096 mL, and can be used directly as a standard uncertainty. The standard uncertainty associated with the final volume upon the extraction of the collection filter, $u(V_x)$, was calculated to be 0.0097 mL by combining the contributions of calibration and repeatability, as shown in Eq. (7).

From the International Union of Pure and Applied Chemistry-recommended atomic-weight values [19, 20], the atomic weights and uncertainties for the constituent elements of formaldehyde and HCHO-DNPH were calculated. The calculated molar masses of formaldehyde and HCHO-DNPH were found to be 30.0260 and 210.1469 g mol^{-1} , respectively. The standard uncertainty for each constituent element of formaldehyde and HCHO-DNPH was determined by dividing these values by $\sqrt{3}$, assuming a rectangular distribution. By multiplying the standard uncertainty for each element (e.g., C, H, O, and N) by the number of atoms, the standard uncertainties of element contributions to the molar mass were determined. Then, the standard uncertainties associated with the molar mass of formalde-

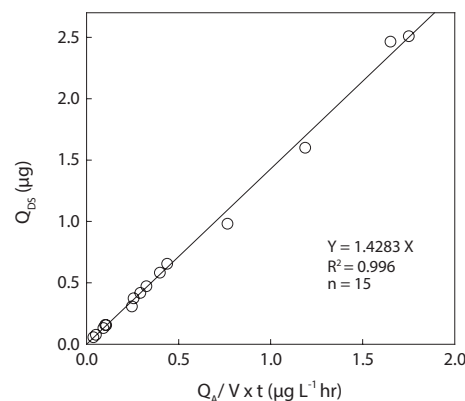


Fig. 1. Comparison of performance of the diffusive samplers and 2,4-dinitrophenylhydrazine (DNPH) cartridge. Q_{DS} and Q_{A} are the mass of the HCHO collected by the diffusive sampler and the DNPH cartridge, respectively (μg), V is the total air volume come into the DNPH cartridge during sampling (L), and t is the sampling time (hr).

hyde and HCHO-DNPH, $u(M_{\text{HCHO}})$ and $u(M_{\text{DNPH}})$, were calculated to be 0.00072 and 0.00463, respectively, by combining each element contribution, as shown in Eq. (9).

The interim combined standard uncertainty associated with the determination of formaldehyde mass, $u(Q_{\text{DS}})$, was calculated according to [16]:

$$u(Q_{\text{DS}}) = \sqrt{\sum_{i=1}^N [c_i \cdot u(x_i)]^2} \quad (10)$$

where c_i is the sensitive coefficients, that is, the change in Q_{DS} produced by a small change Δx_i in input value x_i ($(\Delta Q_{\text{DS}})_i / \Delta x_i$). $u(x_i)$ are the standard uncertainties of the input values estimated for the uncertainty associated with the determination of the formaldehyde mass, which has previously been described.

The formaldehyde mass (Q_{DS}) calculated from Eq. (3) with the input value for each component was $2.9403 \mu\text{g}$. The interim combined standard uncertainty, $u(Q_{\text{DS}})$, calculated from Eq. (10) with the standard uncertainty and sensitive coefficient for each component was $0.0525 \mu\text{g}$. Among the uncertainty contributions assessed in the estimation of the formaldehyde collection, the contribution of the uncertainty associated with the repeatability of the peak area obtained in the process of sample analysis, $u(R_x)$, was the largest (29.6%), and that associated with the determination of the concentration by transforming chromatographic signals in a calibration curve, $u(f_{\text{lin}})$, was similar (27.7%). The contributions of the uncertainties associated with the molar mass of formaldehyde and HCHO-DNPH were similar, and would have virtually no influence on the overall uncertainty.

Uncertainty associated with the sampling time (t) is based on the accuracy of the digital timer (EW-94460-04; Cole-Parmer, Vernon Hills, IL, USA). According to the manufacturer's certificate, the accuracy of the digital timer was given as $\pm 0.07 \text{ sec hr}^{-1}$. The standard uncertainty, $u(t)$, was then calculated using the assumption of a rectangular distribution for the sampler time variation and the value of 0.07 sec hr^{-1} was divided by $\sqrt{3}$. The obtained result was 0.00067 hr , based on a 24-hr measurement.

Fig. 1 shows the comparison of the performance of the diffusive samplers and active cartridges. The results of the comparative measurement of the diffusive samplers and active cartridges over the range of tested formaldehyde concentrations in the two

methods are in close agreement. These data were highly correlated with a coefficient of correlation of 0.996 ($n = 15$). As shown in the relationship described Eq. (4), the sampling rate for the diffusive sampler was calculated to be 1.428 L hr^{-1} with a standard deviation of 0.084 L hr^{-1} . The obtained standard deviation can be used directly as a standard uncertainty associated with the determination of sampling rate.

3.2. Estimation of the Combined Standard and Expanded Uncertainty

The combined standard uncertainty, $u(\text{CON})$, considered all the uncertainty components associated with the determination of the formaldehyde concentration, and was calculated by grouping all these components in three terms, as shown in Eq. (2). The results are summarized in Table 1. The formaldehyde concentration measured with the diffusive sampler was calculated as $85.8 \mu\text{g m}^{-3}$ by applying Eq. (1). The overall combined standard uncertainty, $u(\text{CON})$, in the determination of the formaldehyde concentration with the diffusive sampler was $5.24 \mu\text{g m}^{-3}$. The two major contributors to the combined standard uncertainty were identified as the sampling rate and the mass of the formaldehyde transported into the diffusive sampler by diffusion. This is because the sampling rate is evaluated by the comparison of a reference method (e.g., DNPH cartridge) with a diffusive sampling method, and as a result, the accuracy of the diffusive sampling method needs to be improved to diminish this uncertainty. The contribution of the uncertainty associated with the sampling duration was determined to be insignificant.

The expanded uncertainty, U , was calculated by multiplying the combined uncertainty by a coverage factor ($k_{95} = 2$) which gives a level of confidence of approximately 95%:

$$U = k_{95} \cdot u(\text{CON}) \quad (11)$$

The final result for formaldehyde analysis using the diffusive sampler could be expressed as $85.8 \pm 10.48 \mu\text{g m}^{-3}$ with a relative expanded uncertainty of 12.2%.

4. Conclusions

The results presented here show that the diffusive sampler, using a DNPH as a substrate, is a satisfactory method for determining atmospheric formaldehyde, within ranges which are commonly encountered in indoor air. Furthermore, the diffusive

sampler could easily be used for the evaluation of indoor pollution by the formaldehyde generated from the major sources, such as furniture or building materials.

The uncertainty related to the collection and analysis of formaldehyde was estimated. The uncertainty associated with the determination of the sampling rate and the mass of the formaldehyde transported into the diffusive sampler by diffusion is likely the dominating factor affecting uncertainty of measurement. The quality of analytical measurement results could be assured quantitatively by estimating uncertainty associated with measurement.

To evaluate the contribution of the sampling rate on the uncertainty related to diffusive sampling, the sampling rate should be evaluated under various environmental conditions, such as wind speed, humidity, and temperature. This is a further research subject of this laboratory

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Table 1. Input value, standard uncertainty, and sensitivity coefficient for each component for calculating the combined standard uncertainty associated with determination of HCHO concentration, $u(\text{CON})$

Description	Value	Standard uncertainty	Sensitivity coefficient
$u(\text{CON})$			
Q_{DS} (μg)	2.9403	0.0525	29.1783
t (hr)	24	0.00067	-3.5729
SR (L hr^{-1})	1.428	0.084	-59.6553
Concentration ($\mu\text{g m}^{-3}$)	85.8	5.24	

Q_{DS} : the mass of formaldehyde collected by the diffusive sampler, t : sampling time, SR: sampling rate.

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