

Responses of the Hepatic Microsomal Cytochrome P450 Monooxygenase System in Rock Bream *Oplegnathus fasciatus* Exposed to Tributyltin (TBT)

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Abstract

The study was conducted to investigate the responses of the hepatic microsomal cytochrome P450 monooxygenase system in the rock bream *Oplegnathus fasciatus* after chronic exposure to 0, 1, 2, 4, and 8 µg/L tributyltin (TBT) concentrations for 4 weeks. Hepatic cytochrome 450 content and ethoxyresorufin *O*-deethylation (EROD) activity were found to significantly increase in fish treated with the higher concentration of TBT (≥4 µg/L); however, no significant changes were observed in pentoxyresorufin *O*-deethylation (PROD) activity in all treated groups compared to the control group. These findings suggest that exposure to a low TBT concentration (≥4 µg/L) has the potential to induce cytochrome 450 content and EROD enzyme activity in hepatic tissue in the rock bream.

Key words: *Oplegnathus fasciatus*, Tributyltin, Cytochrome 450, EROD, PROD

Introduction

Many aquatic ecosystems are faced with spatially or temporally alarmingly high levels of xenobiotic chemicals (Brack et al., 2002; Diez et al., 2002). Tributyltin (TBT) is a typical synthetic pollutant, which was selected for investigation in this study. TBT is an organotin compound used primarily as a biocide in antifouling paints for ships, boats, and fishing nets. The use of TBT as an antifouling agent in paint has been banned for small boats and fishing nets in most countries as it has been shown to be toxic to aquatic life and is an endocrine disrupting chemical that generates severe reproductive effects in aquatic animals (Horiguchi et al., 1994). Reproductive impairment in gastropods has been reported at low TBT concentrations (Gibbs et al., 1988). TBT has been shown to

induce imposex and intersex in brosebranches (Oehlmann et al., 1998), and impaired reproductive function in fish has also been described (Mc Allister and Kime, 2003).

Some studies have reported toxic effects of TBT such as morphological and functional alterations of teleost gills in aquatic media (Byrne et al., 1989; Schwaiger et al., 1992; Tsuda et al., 1992; Wang and Huang, 1998). TBT has also been shown to markedly inhibit fish hepatic CYP and in particular the isozyme cytochrome P4501A1 (CYP1A), both in vivo and vitro (Fent and Bucheli, 1994; Morcillo and Porte, 1997). The CYP1A isozyme plays a key role in the biotransformation of xenobiotic compounds, leading to either their detoxification or bioactivation, while other CYP forms are important

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in the metabolism of hormones, fatty acids, and other critical endogenous molecules (Stegeman and Hahn, 1994). Inhibition of the CYP1A-catalyzed 7-ethoxyresorufin *O*-deethylase (EROD) and BaP hydroxylase (BPH) reactions by TBT could also interfere with their use as biomarkers of organic pollution in environmental monitoring (Goksoyr, 1995; Livingstone and Goldfarb, 1998; Whyte et al., 2000).

The rock bream is an economically important fish species in Korea that is commonly cultured for food in marine-based cages. Despite the importance of the rock bream in Korea, relatively little information is available regarding the effects of TBT on the species. Therefore, this study investigated the responses of the xenobiotic biotransformation system in the hepatic tissues of rock bream following chronic waterborne TBT exposure. The study evaluated the effects of TBT on the cytochrome P450 monooxygenase system in rock bream, and the findings suggest that the species can be used as an indicator of the TBT toxicity level in water systems.

Materials and Methods

Experimental fish

Rock bream individuals were obtained from a fish farm in Gyeongsangnam-do, Korea. Fish were held for 3 weeks to acclimatize and to evaluate the overall fish health under laboratory conditions ($20.2 \pm 0.34^\circ\text{C}$) prior to exposure. During the acclimation period, fish were fed with a commercial diet twice daily and maintained on a 12:12 h light/dark cycle at all times. After acclimatization, fish (body length, 12.8 ± 0.32 cm; body weight, 33.7 ± 0.65 g) were selected for the experiments.

Exposure conditions

The exposure took place in 30-L glass tanks ($500 \times 280 \times 310$ mm) containing 20 fish in each treatment group. Each tank received a water flow of 7 L min^{-1} with continuous aeration. Tributyltin chloride (96% purity) was purchased from Sigma-Aldrich Chemie GmbH (Steinheim, Germany). A stock solution of TBT was prepared by diluting 0.02 g TBT in 1 mL acetone before use. Acetone is known to be nontoxic in terms of histopathological effects (Fent and Meier, 1992). TBT was added to the experimental tanks at final concentrations of 0, 1, 2, 4, and $8 \mu\text{g/L}$ for 4 weeks as a chronic exposure. All experiments were conducted using a static system. TBT solutions in each experimental tank were renewed once daily.

Hepatic somatic index

The hepatic somatic index (HSI) was determined as the ratio of liver weight to body weight.

$$\text{HSI (\%)} = \text{liver wet weight/weight} \times 100$$

Cytochrome P450 assay

Hepatic cytochrome P450 content was determined from the carbon monoxide–oxide difference spectra of dithionite-reduced microsomes using an extinction coefficient of $91 \text{ mM}^{-1} \text{ cm}^{-1}$ for A450–A490 (Omura and Sato, 1964).

De-alkylase assays

Hepatic EROD activity was determined using a modification of the multiwell plate method (Kennedy and Jones, 1994). The following reagent concentrations were used in the EROD reaction mixture: 7-ethoxyresorufin, $1.7 \mu\text{M}$ (Sigma, St. Louis, MO); NADPH, 0.5 mM (Boehringer Mannheim, Mannheim, Germany); MgSO_4 , 17 mM; and HEPES, 0.1 M, pH 7.8 M (Sigma). EROD activity was determined in 48-well culture plates (Costar, Cambridge, MA) using a plate reading fluorometer (BF10001, Packard Bioscience Co., Meriden, CT). Excitation and emission filters were set at 530 nm and 590 nm, respectively. The activity was determined at 25°C using $25 \mu\text{L}$ postmitochondrial supernatant in the reaction mixture. The pentoxyresorufin *O*-deethylation (PROD) activity was assayed in the same systems as EROD. Enzyme activities were calculated using the increase in fluorescence induced by the addition of resorufin. Protein concentration was determined using a method described by Bradford (1976) using bovine serum albumin (Sigma) as a standard.

Statistical analysis

Statistical analysis of the results was performed using the SPSS/PC+ statistical package (SPSS Inc., Chicago, IL). ANOVAs and Duncan's test for multiple comparisons were used to test for significant differences between the control and treatment (Duncan, 1955). The significance level was set at $P < 0.05$.

Results

The effect of waterborne TBT exposure for 4 weeks on the HSI of rock bream is shown in Fig. 1. No differences were observed in the HSI of treated groups compared to the control. Fig. 2. shows the effect of waterborne TBT on hepatic CYP450 content in rock bream. Hepatic CYP450 content in the 0, 1, and $2 \mu\text{g/L}$ -treated groups were stable during the experimental period. However, the hepatic CYP450 content significantly increased in the $4 \mu\text{g/L}$ -treated group after 2 weeks exposure to TBT ($P > 0.05$). The effect of waterborne TBT exposure for 4 weeks on the hepatic EROD activity of rock bream is shown in Fig. 3. Hepatic EROD activity increased insignificantly in the 1 and $2 \mu\text{g/L}$ TBT-treated groups after 2 weeks; however, a significant increase was observed in the 4 and $8 \mu\text{g/L}$ -treated groups. A significant increase in hepatic EROD activity after

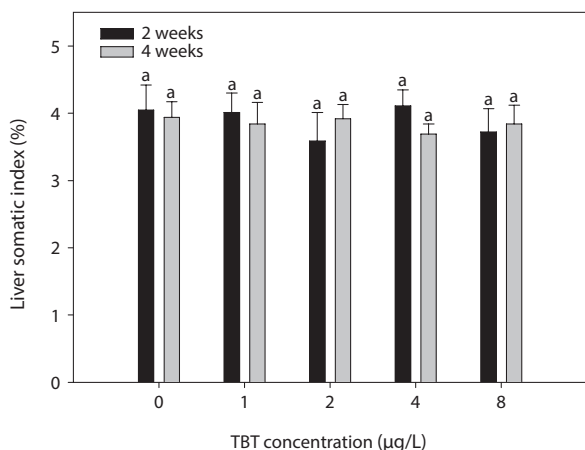


Fig. 1. Hepatic somatic index of *Oplegnathus fasciatus* exposed to the various TBT for 4 weeks. All data are expressed as mean±SE. Different letters indicates significant difference ($P < 0.05$) between the groups.

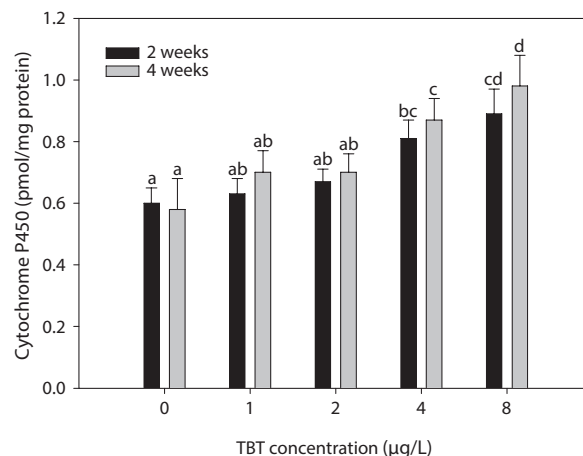


Fig. 2. Hepatic Cytochrome P450 contents of *Oplegnathus fasciatus* exposed to the various TBT for 4 weeks. All data are expressed as mean±SE. Different letters indicates significant difference ($P < 0.05$) between the groups.

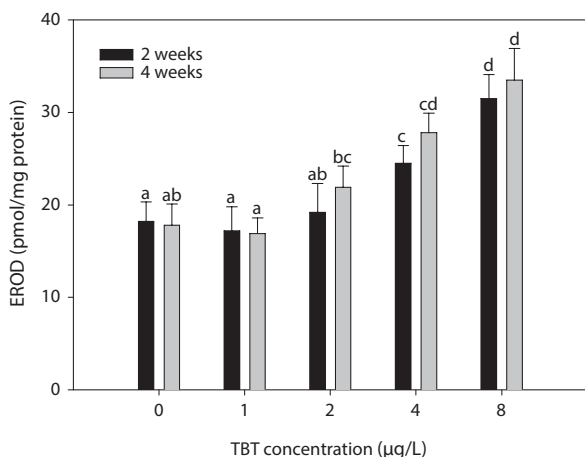


Fig. 3. Hepatic EROD activity of *Oplegnathus fasciatus* exposed to the various TBT for 4 weeks. All data are expressed as mean±SE. Different letters indicates significant difference ($P < 0.05$) between the groups.

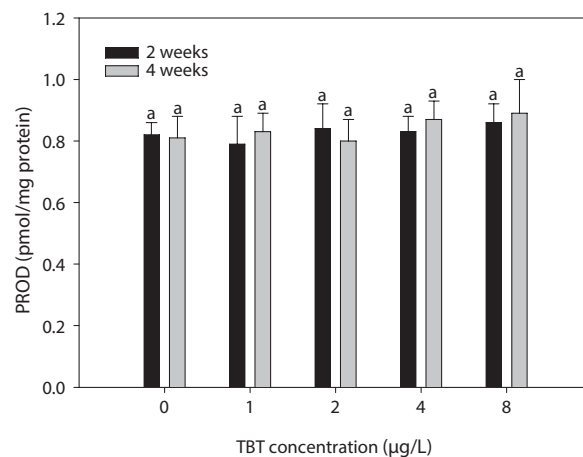


Fig. 4. Hepatic PROD activity of *Oplegnathus fasciatus* exposed to the various TBT for 4 weeks. All data are expressed as mean±SE. Different letters indicates significant difference ($P < 0.05$) between the groups.

4 weeks was observed in the groups treated with TBT concentrations of more than 2 µg/L. Fig. 4. shows the effects of TBT on PROD activity in the hepatic microsomes of rock bream. No significant changes were observed in all the treated groups compared to the control group ($P > 0.05$).

Discussion

The results clearly indicate that exposure of rock bream to TBT induced hepatic cytochrome P450 and de-alkylase activity. However, no significant change occurred in the HSI. A similar result for was observed previously in common carp supplemented with anthraquinone (Xie et al., 2008).

The duration and intensity of action of xenobiotics within a biological system are determined by the rate of their biotransformation to pharmacologically active or inactive metabolites. Cytochrome P450, a heme protein, is a heterogeneous system of microsomal enzymes responsible for the oxidative biotransformation of many chemicals to polar metabolites, thereby facilitating the pharmacological inactivation of these chemicals and their elimination from the body (Lu and West, 1980; Wrighton and Stevens, 1992). Cytochrome P450 occurs in multiple forms and the composition of these isozymes, as well as their relative concentrations in tissues, are influenced by treatment with different chemicals (Lu and West, 1980). Several studies have observed that TBT suppresses cytochrome P450 1A in fish. The influence of TBT on hepatic cytochrome

P450 levels in aquatic organisms has been studied in practical situations (Rice and Roszell, 1998).

In this study, a significant increase in hepatic cytochrome P450 level was observed in groups exposed to TBT. The hepatic cytochrome P450 content of the 4 µg/L-exposure group was significantly different from the control for an exposure duration of either 2 or 4 weeks. At the end of the experiment, the cytochrome P450 content of the 8 µg/L-exposure group was higher than in the control group. A pronounced increase in hepatic cytochrome P450 induction occurred in the presence of TBT, which confirms that cytochrome P450 induction in the rock bream can be a reliable indicator of contamination of the aquatic environment by compounds known to induce CYP1A, including many xenobiotics (Colloer et al., 1995; Yolanda et al., 2004).

Hepatic EROD activity in rock bream is increasingly used to indicate the presence or effects of certain organic contaminants. This is because laboratory studies have established a strong causal link between exposure of the fish to contaminants and the expression of cytochrome P4501A1 and EROD activity (Haasch et al., 1989; Skaare et al., 1991). Although monitoring of cytochrome P4501A1 and EROD activity has been widely used as biomarkers of aquatic pollution, the effect of TBT on these parameters has not yet been established. Our results have shown that following waterborne exposure of TBT (8 µg/L), significant induction of hepatic EROD occurred after 4 weeks.

In this study, no significant difference between TBT-treated and control groups in terms of hepatic PROD activity was observed. PROD activity is a catalytic probe for determining the induction response of CYP2B class isozymes in mammals. In the P450 system of fish, the phenobarbital (PB)-type inductive response appears to be completely absent (Goksøyr and Förlin, 1992). CYP2B genes have been established to be present and expressed in fish, but also that they are nonresponsive to PB-type compounds (Stegeman et al., 1990). In a study by Brown (1992), a large series of gas chromatograms of polychlorinated biphenyl (PCB) residues in 32 species of teleost fish was investigated. CYP2B-like alteration patterns were observed in only four species. In this study, insignificant PROD induction was detected in the rock bream exposed to TBT. Therefore, TBT does not induce hepatic CYP2B class isozyme. In conclusion, waterborne TBT significantly affected the cytochrome P450 content and EROD activity. The level of the decrease induced by chronic exposure to TBT can be used as a biomarker in coastal and estuarine risk assessments.

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Reference

- Bradford M. 1976. A rapid and sensitive method for the quantitation of microgram quantities of protein utilizing the principle of protein-dye binding. *Anal Biochem* 72, 248-245.
- Brack W, Schirmer K, Kind T, Schrader S, Schüürmann G. 2002. Effect-directed fractionation and identification of cytochrome P450A-inducing halogenated aromatic hydrocarbons in a contaminated sediment. *Environ Toxicol Chem* 21, 2654-2662.
- Brown JF. 1992. Metabolic alteration of PCB residues in aquatic fauna: distributions of cytochrome P4501A1- and P4502B-like activities. *Mar Environ Res* 34, 261-266.
- Byrne P, D Speare, HW Ferguson. 1989. Effects of a cationic detergent on the gills and blood chemistry of rainbow trout *Salmo gairdneri*. *Dis Aquat Org* 6, 195-196.
- Collier TK, BF Anulacion, JE Stein, A Goksøyr and U Varanasi. 1995. A field evaluation of cytochrome P4501A as a biomarker of contaminant exposure in three species of flat fish. *Toxicol Chem* 14, 143-152.
- Diez S, Abalos M, Bayona JM. 2002. Organotin contamination in sediments from the Western Mediterranean enclosures following 10 years of TBT regulation. *Water Res* 36, 905-918. 199-213.
- Fent K, Bucheli TD. 1994. Inhibition of hepatic microsomal monooxygenase system by organotins *in vitro* in freshwater fish. *Aquat Toxicol* 28, 107-126.
- Fent K and Meier. 1992. Tributyltin-induced effects on early life stages of minnows, *Phoxinus phoxinus*. *Arch Environ contam Toxicol*, 22, 429-438.
- Gibbs PE, Pascoe PL, Burt GR. 1988. Sex change in the female dogwhelk, *Nucella lapillus* around southwest England: Evidence for the effect of tributyltin from antifouling paints. *J Mar Biol Assoc UK* 68, 715-731.
- Goksøyr A. 1995. Use of cytochrome P450 1A (CYP1A) in fish as a biomarker of aquatic pollution. *Arch Toxicol* 17, 80-95.
- Goksøyr A and L Förlin. 1992. The cytochrome P450 system in fish, aquatic toxicology and environmental monitoring. *Aquat Toxicol* 22, 287-312.
- Haasch ML, PJ Wejksnora, JJ Stegeman and JJ Lech. 1989. Cloned rainbow trout liver P450 complementary DNA as a potential environmental monitor. *Toxicol Appl Pharmacol* 98, 362-368.
- Horiguchi T, Hiroaki S, Makoto S, Sunao Y and Masatoshi M. 1994. Organotin compounds and their effects on aquatic organisms, focusing on imposex in gastropods. *Main Group Metal Chem*, 17(1-4), 81-100.
- Kennedy SW and Jones SP. Simultaneous measurement of cytochrome P4501A catalytic activity and total protein concentration with a fluorescence plate. *Anal. Biochem* 222, 217-223.
- Livingstone DR, Goldfarb PS. 1998. Aquatic environmental biomonitoring: Use of cytochrome P450 1A and other molecular biomarkers in fish and mussels. In Lynch J Wiseman A eds *Environmental Biomonitoring: The Biotechnology Ecotoxicology Interface Vol 6-Biotechnology Reserch Series Cambridge University Press Cambridge UK*, 101-129.
- Lu AYH and SB west. 1980. Multiplicity of mammalian microsomal

- cytochrome P-450. *Pharmacological Reviews* 31, 277-295.
- Mc Allister B, Kime ED. 2003. Early life exposure to environmental levels of the aromatase inhibitor tributyltin causes masculinisation and irreversible sperm damage in zebrafish (*Danio rerio*). *Aquat Toxicol* 65, 309-316.
- Morcillo Y, Porte C. 1997. Interaction of tributyltin- and triphenyltin with the microsomal monooxygenase system of mollusks and fish from the NW Mediterranean. *Aquat Toxicol* 38, 35-46.
- Oehlmann J, Bauer B, Minchin D, Schulte-Oehlmann U, Fioroni P, Markert B. 1998. Imposex in *Nucella lapillus* and intersex in *Littorina littorea*: interspecific comparison of two TBT-induced effects and their geographical uniformity. *Hydrobiologia* 378, 199-213.
- Omura T and Sato R. 1964. The carbon monoxide-binding pigment of liver microsomes. 2. Solubilization, purification and properties. *J Biol Chem* 239, 2379-2385.
- Ric CD and Roszell LE. 1998. Tributyltin modulates 3,3',4,4',5-pentachlorobiphenyl (PCB-126)-induced hepatic CYP1A activity in channel catfish, *Ictalurus punctatus*. *Journal of Toxicology and Environmental Health, Part A* 55, 197-212.
- Schwaiger J, F Bucher, H Ferling. 1992. A prolonged toxicity study on the effects of sublethal concentration of bis (tri-*n*-butyltin) oxide (TBTO): histopathological and histochemical findings in rainbow trout (*Oncorhynchus mykiss*). *Aquat Toxicol* 23, 31-48.
- Skaare JU, EG Jensen, A Goksoeyr, E Egaas. 1991. Response of xenobiotic metabolizing enzymes of rainbow trout (*Oncorhynchus mykiss*) to the mono-ortho substituted polychlorinated PCB congener 2, 3', 4, 4', 5-pentachlorobiphenyl, PCB-118, detected by enzyme activities and immunochemical methods. *Arch Environ Contam Toxicol* 20, 349-352.
- Stegeman JJ, Hahn ME. 1994. Biochemistry and molecular biology of monooxygenases: Current perspectives on forms, functions, and regulation of cytochrome P450 in aquatic species. In Malins DC Ostrander GK eds *Aquatic Toxicology Molecular Biochemical and Cellular Perspectives* Lewis Boca Raton FL USA, 87-206.
- Stegeman JJ, BR Woodlin and RM Smolowitz. 1990. Structure, function and regulation of cytochrome P450 forms in fish. *Biochem Soc Trans* 18, 19-21.
- Tsuda T, S Aoki, M Kojima, T Fujita. 1992. Accumulation and excretion of tri-*n*-butyltin chloride and tributyltin chloride by willow shiner. *Comp Biochem Physiol* 101C, 67-70.
- Wang DY, BQ Huang. 1998. Toxic effects of tributyltin (TBT) on early life stage of thornfish (*Terapon jarbua* Forsskål). *J Fish Soc Taiwan* 25, 15-25.
- Whyte JJ, Jung RE, Schmitt CJ, Tillitt DE. 2000. Ethoxyresorufin-*O*-deethylase (EROD) activity in fish as a biomarker of chemical exposure. *Crit Rev Toxicol* 30, 347-570.
- Wrighton SA and JC Stevens. 1992. The human hepatic cytochrome P-450 involved in drug metabolism. *Crit Rev Toxicol* 22, 1-22.
- Xie J, Liu B, Zhou Q, Su Y, He Y, Pan L, Ge X, and Xu P. 2008. Effects of anthraquinone extract from rhubarb *Rheum officinal* Bail on the crowding stress response and growth of common carp *Cyprinus carpio* var. Jian. *Aquaculture* 281, 5-11.
- Yolanda M. Gemma J. Sean CM. Hara O. David RL. And Cinta P. 2004. Interaction of tributyltin with hepatic cytochrome P450 and uridine diphosphate-glucuronosyl transferase system of fish. *Environ Toxicol and Chem*, 23(4), 990-996.