

# Association between coarse woody debris and small mammals and insectivores in managed forests

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## Abstract

Coarse woody debris (CWD) is generally considered dead woody material in various stages of forest decomposition and has been hypothesized to be an important habitat feature for mammals in forests of the Pacific Northwest, USA. Sherman and pitfall trapping were conducted for 2 years on three paired sites with low and high amounts of CWD. Deer mice was the dominant species with a total capture of 605 (45.6%). Four species of insectivores were captured, including *Sorex monicolus*, *S. trowbridgii*, *S. vagrans*, and *Neurotrichus gibbsii*. A Poisson regression model was used to test whether 11 CWD variables could predict insectivore captures. The volume of logs and mean decay were important variables for deer mice use of CWD. Mean distance from pieces of CWD to the capture point was significantly related to the total number of captures of trowbridge shrew (*Sorex trowbridgii*) and all insectivore species. Vagrant shrews (*Sorex vagrans*) were significantly associated with log volume. Retaining large size CWD should be part of a management plan for ground-dwelling insectivores in forests to secure their biodiversity.

**Key words:** community ecology, deer mice, insectivores, microhabitat, Poisson model

## INTRODUCTION

Retention of coarse woody debris (CWD) is an important issue for sustainable forest management. CWD is any standing dead tree (snag), downed bole, or downed large branch (>10 cm in diameter) (Harmon et al. 1986, Spies et al. 1988, Yan et al. 2006). Many small mammals such as deer mice depend on CWD in forest ecosystems. Studies on small mammals in forests of the Pacific Northwest USA (Maser et al. 1979, Carey and Johnson 1995), southeastern pine forests in the USA (Loeb 1999, Greenberg 2002), and Korean mixed forests (Kim et al. 2006) have indicated that the abundance of CWD is a good predictor of the small mammal abundance in forested ecosystems. Small mammals that live in any given habitat type depend on the differential use of microhabitats (Morrison and Anthony 1989, Carey and Harrington 2001). Additionally, CWD can

provide for groups of taxa that are heavily dependent on moisture as well as temperature sensitive groups such as amphibians and reptiles (Owens et al. 2008). Small mammals use CWD as a necessary ground habitat feature, indicating that microhabitat is important with increasing population stability and reproductive benefits (Lee 1995). Small mammals including insectivores in the forest depend on CWD in the course of their life history, because CWD provides cover and forms survival and breeding microhabitats. Thus, microhabitat studies have clarified small mammal affinities for particular microsite features within a habitat.

The amount of CWD in old-growth forests is frequently sufficient for small mammal survival (West 1991, Ucitel et al. 2003) but may be a limiting factor in second-growth

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forests where CWD loads are lower (Lee 2004). Therefore, it would be instructive to examine the relationships between CWD and small mammals in managed forests, particularly considering that managed forests comprise > 80% of forests in the Pacific Northwest. The objectives of this study were to compare small mammal communities between sites with low and high densities of CWD in managed forests and to describe the associations between CWD and small mammal species. It was hypothesized that variations in CWD influence small mammal communities and microhabitat associations.

## MATERIALS AND METHODS

This study was conducted on the Ft. Lewis Military Reservation, east of the southern tip of Puget Trough Province in Pierce County, WA, USA. The study area was relatively flat with an average elevation of 250 m above sea level, and Douglas fir (*Pseudotsuga menziesii*) dominates the 50-60 year old stands. Predominant understory vegetation included salal (*Gaultheria shallon*), Oregon grape (*Berberis nervosa*), and sword fern (*Polystichum munitum*).

Three paired sites were selected; three with low densities of CWD (treatment) and three with high densities of CWD (control). Sites were paired based on relative proximity (two pairs with sites within 2 km of each other and one pair with sites within 5 km) to minimize the variation between the treatment and control. Low CWD sites had woody material, particularly logs, removed around 1910 to minimize forest fire hazards. Data were collected in 1991-1993 throughout the season with 2-week intervals for 3 days of each trapping period (refer to Fig. 2 in Lee 1995 for population dynamics).

### CWD estimation

Length, diameter at both ends, decay class (Sollins 1982), species, and location of downed logs were measured. The amount of CWD was also determined. Downed pieces  $\geq 10$  cm diameter at the small end were counted as downed logs. Circumference of the base and height of all snags and diameter and height of all stumps was measured. Some volumes of CWD (volume of logs, stumps, and snags) were measured following Lee (1995).

### Animal capture using Sherman and Pitfall traps

Small mammal trapping was conducted on a 1-ha grid at each of the six sites. Each sampling grid consisted of a

10 × 10 array of stations with 10 m spacing. One Sherman collapsible trap (7.6 × 8.9 × 22.9 cm) was placed within a 2 m radius of each station. Each trap contained synthetic bedding and rolled oats for bait. The traps were baited and opened in late afternoon and checked and closed in the early morning. Each trap grid was normally run for 2 consecutive days twice per month for 25 months (June 1991-June 1993) except for November- February when trapping occurred once per month. Trapping was not conducted in January and February 1993 due to unusually cold conditions.

The sex, age (adult, subadult), weight, and reproductive condition of all trapped animals were individually recorded. Males were described as either scrotal (descended testes) or non-scrotal (abdominal testes). Females were described as either nonreproductive or reproductive (pregnant or lactating).

Twenty-five pitfall traps, spaced 10 m apart, were placed in a 5 × 5 grid within the center of a trapping grid. Pitfalls consisted of no. 10 tin cans (15 cm in diameter) buried flush with the surface of the soil and protected with a piece of wood propped one to several cm over the can lip. Pitfall traps were filled with water to a depth of 5-10 cm to reduce escapes and hasten mortality. Pitfalls were opened 2 consecutive nights every 2 weeks.

### Data analyses

The abundance of each species in the community was analyzed using a paired *t*-test, because sites were carefully chosen with much similarity in each pair except for the different amounts of CWD. The same numbers of Sherman and pitfall traps were used at each site to minimize effects from trap differences. Insectivores were most likely to be captured with the use of pitfall traps due to their creeping behavior, and the trapped insectivores did not make it overnight so that capture-marking-recapture method was only limited to the deer mice study. For these reasons, the deer mice and insectivore data were separated and explained separately in the results. The number of captured animals was used for the community analysis (Table 1), and the capture points of the animals were used for the microhabitat analysis.

To determine the ecological relationship of deer mice to CWD, Poisson regression was selected to model seven response variables (total captures, captures of males, captures of females, proportion of males captured, captures of adults, captures of subadults, and captures by season [summer vs. winter]) as a function of CWD, pairs of sites, and low vs. high CWD sites (Zar 1999). Summer was de-

fined as May-October and winter as November-April. Thirteen variables were used as predictor variables for CWD function (total volume of CWD, volume of logs, log volume multiplied by distance, volume of stumps, stump volume multiplied by distance, total number of CWD, number of logs, number of stumps, mean decay class, mean distance to capture point, volume of snag, and block, treatment). Block and treatment effects were considered macrohabitat factors in all models.

To determine the ecological relationship of CWD with insectivores, a Poisson regression was also selected to model the number of insectivore captures as a function of CWD, pairs of sites, and low vs. high CWD sites (Zar 1999). Five response variables were used in the analyses (total number of insectivore captures, captures of shrew-moles [*Neurotrichus gibbsii*], montane shrews [*Sorex monticolus*], Trowbridge shrews [*S. trowbridgii*], and vagrant *S. vagrans* and as predictor variables, 11 characteristics of CWD (total volume of CWD, volume of logs, volume of stumps, total number of pieces of CWD, number of logs, number of stumps, mean decay class, mean distance to trap station, volume of snags, paired sites, low vs. high CWD) were measured within a 5 m radius for each trapping station.

## RESULTS

### Small mammal community

The total volume of all CWD was higher ( $P < 0.05$ ) in the high CWD sites than in the low CWD sites (refer to Table 1 in Lee 1995). A total of 1,326 individual small mammals were trapped during the 2-year study (Table 1). *Sorex trowbridgii* ( $N = 330$ , 24.9%) was the most frequent insectivore, followed by *N. gibbsii*, *S. monticolus*, and *S. vagrans*. Deer mice (*Peromyscus maniculatus*) was the most common small mammal ( $N = 605$ , 45.6%). *Sorex vagrans*, *T. townsendii*, and *C. gapperi* showed a difference in numbers ( $P < 0.05$ ) between low and high CWD (Table 1).

### Microhabitat analysis of deer mice and insectivores

The regression models used 582 individual (380 Sherman traps and 242 pitfall) insectivores from Table 1. Data from 1,425 captured deer mice (data not shown) were used to assess microhabitat structure. All Poisson regression models indicated a significant difference between treatment and control sites; that is, treatment was always

**Table 1.** Total number of individual small mammals captured in both live and pitfall traps for each site and paired t-test results between high and low coarse woody debris (CWD) sites

Species	Paired sites						Total	t-test	P-value
	I		II		III				
	High CWD	Low CWD	High CWD	Low CWD	High CWD	Low CWD			
<i>Sorex monticolus</i>	15 (6)	3 (2)	23 (18)	11 (2)	15 (11)	26 (4)	93 (43)	0.57	0.62
<i>Sorex trowbridgii</i>	44 (23)	15 (3)	69 (48)	69 (45)	54 (37)	79 (47)	330 (203)	0.04	0.97
<i>Sorex vagrans</i>	17 (10)	35 (13)	0	8 (5)	0	17 (7)	77 (35)	-4.51	0.05*
<i>Neurotrichus gibbsii</i>	5 (4)	11 (11)	22 (16)	17 (9)	11 (6)	16 (13)	82 (59)	-0.32	0.78
Unidentified <i>Sorex</i>	1	1 (1)	1 (1)	0	2 (1)	4 (3)	9 (6)	-	
<i>Tamias townsendii</i>	2	6	1	7	3	7	26	-7.00	0.02*
<i>Tamias douglasii</i>	0	1	2	2	1	1	7	-1.00	0.42
<i>Glaucomys sabrinus</i>	0	0	0	0	2	0	2	-	
<i>Microtus oregoni</i>	2	7 (2)	2	7 (2)	2	50 (8)	70 (12)	-1.35	0.31
<i>Clethrionomys gapperi</i>	6	0	9	2	5	0	22	10.39	0.01**
<i>Peromyscus maniculatus</i>	159 (14)	60 (1)	107 (1)	100 (5)	105 (1)	74 (4)	605 (26)	1.66	0.24
<i>Mustela ernimea</i>	3	0	0	0	0	0	3	-	
Total	254 (57)	139 (33)	236 (84)	223 (68)	200 (56)	274 (86)	1,326 (384)		

Number of pitfall trap captures is in parentheses. - indicates no test was performed. \* $P < 0.05$ , \*\* $P < 0.01$ .

a significant variable (Table 2). Of the 13 microhabitat predictor variables, only six were related to microhabitat use of CWD (Table 2). The predictor variables of the total number of CWD pieces and mean decay were significant for predicting total captures of deer mice (Table 2). Total captures of males was significantly related to the number of CWD pieces and mean decay but captures of females were not significantly related to any of the 11 variables. The proportion of males captured was significantly related to both the volume of logs and stumps (Table 2). Microhabitat use by age class showed that total adult captures were significantly related only to CWD. In contrast, the number of logs, mean decay class, and mean distance from the capture points had a significant effect on sub-adult captures. The proportion of adults was 39.5%. Microhabitat use between summer and winter was not different (data not shown).

In the insectivore analyses, the Poisson model showed that only *S. vagrans* was significantly different between low and high CWD (Table 2). The mean distance of CWD to capture points had a significant effect on both total captures of all insectivores and those of *S. trowbridgii*. This was not surprising because most insectivores (203/346, 59%) were *S. trowbridgii*. Log volume had a significant effect on *S. vagrans* captures.

## DISCUSSION

The insectivorous community in this study was dominated by *S. trowbridgii*. This species is known to be a habitat generalist (Carey and Harrington 2001). *Sorex vagrans* was more abundant in sites with low CWD than in sites with high CWD. *Sorex vagrans* increased in frequency in open grassy areas on well drained sites. At the same time, the population of *S. vagrans* fluctuated greatly at sites with low amounts of CWD, indicating that the population was less stable (Lee 1995).

Although the number of deer mice was not different between sites with low and high CWD, within-site use signified the importance of the CWD. Capture frequency of *P. maniculatus* showed significant differences in variables with a greater total number of CWD. Additionally, the number of deer mice was higher in sites with advanced CWD decay. Favorable microsites for *P. maniculatus* were characterized by advanced decay class and an abundance of CWD. At the site level, volume of CWD was most often a good predictor of CWD use by deer mice *P. leucopus* habitat use was strongly influenced by CWD abundance as it reduces the risk of predation and provides a wider range of areas to forage more efficiently (Fitzerald and Wolff 1988, Barnum et al. 1992).

**Table 2.** A list of significant coefficients from the generalized linear model for deer mice (*Peromyscus maniculatus*) and insectivores

Response variable	Variables							
	Volume logs	Volume stumps	Number CWD	Number logs	Mean decay	Mean distance	Pairs	Low vs High CWD
Deermice								
Total captures			0.01**		0.39*			0.31***
Males			0.02**		0.47**			0.32***
Females								0.32***
Proportion of males captured	0.01*	-0.03**						
Adults captured			0.01*					0.51***
Subadults captured				0.02**	0.44**	-0.17*		0.24***
Captures by season								0.32***
Insectivores								
Total captures						0.23**	0.17*	
Shrew-moles captured								
Montane ( <i>Sorex montanus</i> ) shrews captured								0.27***
Trowbridge ( <i>S. trowbridgii</i> ) shrews captured						0.15**	0.37***	
Vagrant ( <i>S. vagrans</i> ) shrews captured	0.02*						0.53***	

Poisson regression models were used for all responses except for proportion of males, which was modeled by binomial regression. Thirteen variables were tested, but coefficients were for only those variables significant at  $\alpha = 0.05$ . \* $P < 0.05$ , \*\* $P < 0.01$ , \*\*\* $P < 0.001$ .

We also found that subadult deer mice were closely associated with some characteristics of the CWD (number of logs, mean decay class, and mean distance). It seemed that the area covered by adults was larger than that of subadults, resulting in a better correlation between habitat characteristics of the subadult population and CWD.

Although a 5 m radius was used to estimate the total effect of CWD on insectivores, the mean distance from CWD to the capture station appeared to be the only significant variable for *S. trowbridgii*. Total captures of all insectivores were also highly associated with mean distance, suggesting that the proximity of CWD was important to insectivore abundance. *Sorex trowbridgii* was the most abundant species among the four insectivores and strongly influenced the results of all insectivores. Similarly, Carey and Harrington (2001) also observed an increased number of insectivores at high CWD sites.

Two conclusions can be reached: 1) the number of pieces of CWD and mean decay class, but not volume, were important for *P. maniculatus* habitat use; 2) mean distance to the capture point was a significant habitat variable for insectivores, suggesting the importance of proximity to CWD. We also found that volume of CWD was a good predictor of deer mice abundance among sites. Thus, it is critical to retain a sufficient number of logs and snags in forest ecosystems. Unfortunately, current logging practices greatly reduce the amount of CWD (Spies et al. 1988). The size of logs and snags determines the duration of their use, because large CWD generally lasts longer than smaller CWD. In this regard, an uneven-aged structure may best ensure continual input and existence of sufficient CWD. Retaining large logs is also important in nutrient cycling and cover for larger animals. This study used fixed trapping points but monitoring individual animals (e.g., radio-telemetry) would determine CWD use more clearly.

In conclusion, the importance of CWD as a habitat feature for small mammals was tested at sites with high and low amounts of CWD. Analysis of within-site variability revealed that the number of pieces of CWD and mean decay class were positively related to deer mice captures. Captures of insectivores at trapping stations were highly associated with mean distance from CWD indicating that proximity to CWD was important to insectivores. Retention of large CWD can be an important for rodent management in forests of the Pacific Northwest.

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## LITERATURE CITED

- Barnum SA, Manville CJ, Tester JR, Carmen WJ. 1992. Path selection by *Peromyscus leucopus* in the presence and absence of vegetative cover. *J Mammal* 73: 797-801.
- Carey AB, Harrington CA. 2001. Small mammals in young forests: implications for management for sustainability. *For Ecol Manag* 154: 289-309.
- Carey AB, Johnson ML. 1995. Small mammals in managed, naturally young, and old-growth forests. *Ecol App* 5: 336-352.
- Fitzgerald VJ, Wolff JO. 1988. Behavioral responses of escaping *Peromyscus leucopus* to wet and dry substrata. *J Mammal* 69: 825-828.
- Greenberg CH. 2002. Response of white-footed mice (*Peromyscus leucopus*) to coarse woody debris and micro-site use in southern Appalachian treefall gaps. *For Ecol Manag* 164: 57-66.
- Harmon ME, Franklin JF, Swanson FJ, Sollins P, Gregory SV, Lattin JD, Anderson NH, Cline SP, Aumen NG, Sedell JR, Lienkaemper GW, Cromack K Jr, Cummins KW. 1986. Ecology of coarse woody debris in temperate ecosystems. *Adv Ecol Res* 15: 133-302.
- Kim RH, Son Y, Lim JH, Lee IK, Seo KW, Koo JW, Noh NJ, Ryu SR, Hong SK, Ihm BS. 2006. Coarse woody debris mass and nutrients in forest ecosystems of Korea. *Ecol Res* 21: 819-827.
- Lee SD. 1995. Comparison of population characteristics of three species of shrews and the shrew-mole in habitats with different amounts of coarse woody debris. *Acta Theriol* 40: 415-424.
- Lee SD. 2004. Population dynamics and demography of deermice (*Peromyscus maniculatus*) in heterogeneous habitat: role of coarse woody debris. *Pol J Ecol* 52: 55-62.
- Loeb SC. 1999. Responses of small mammals to coarse woody debris in a southeastern pine forest. *J Mammal* 80: 460-471.
- Maser C, Anderson RG, Cromack K Jr, Williams JT, Martin RE. 1979. Dead and down woody material. In: *Wildlife Habitats in Managed Forests: The Blue Mountains of Oregon and Washington*. USDA Agricultural Handbook No. 553 (Thomas JW, ed). Wildlife Management Institute, Washington, DC, pp 78-95.
- Morrison ML, Anthony RG. 1989. Habitat use by small mam-

- mals on early growth clear-cuttings in western Oregon. *Can J Zool* 67: 805-811.
- Owens AK, Moseley KR, McCay TS, Castleberry SB, Kilgo JC, Ford WM. 2008. Amphibian and reptile community response to coarse woody manipulations in upland loblolly pine (*Pinus taeda*) forests. *For Ecol Manag* 256: 2078-2083.
- Sollins P. 1982. Input and decay of coarse woody debris in coniferous stands in western Oregon and Washington. *Can J For Res* 12: 18-28.
- Spies TA, Franklin JF, Thomas TB. 1988. Coarse woody debris in Douglas-fir forests of western Oregon and Washington. *Ecology* 69: 1689-1702.
- Ucitel D, Christian DP, Graham JM. 2003. Vole use of coarse woody debris and implications for habitat and fuel management. *J Wildl Manag* 67: 65-72.
- West SD. 1991. Small mammal communities in the Southern Washington Cascade Range. In: *Wildlife and Vegetation of Unmanaged Douglas-fir Forests*. Genetical Technical Report PNW-GTR-285 (Ruggiero LF, Aubry KB, Carey AB, Huff MH, eds). US Department of Agriculture, Forest Service, Pacific Northwest Research Station, Portland, OR, pp 269-284.
- Yan E, Wang X, Huang J. 2006. Concept and classification of coarse woody debris in forest ecosystems. *Front Biol China* 1: 76-84.
- Zar JH. 1999. *Biostatistical Analysis*. 4th ed. Prentice Hall, Inc., Englewood Cliffs, NJ.