

# The Variation in the Species Composition of the Soil Seed Bank in the Natural Flood Plain Vegetation along the Urban Reach of Han River, South Korea

Lee, Hyo Hye Mi<sup>1,2</sup>, Rob H. Marrs<sup>2</sup> and Eun Ju Lee<sup>1,\*</sup>

<sup>1</sup>*School of Biological Sciences, Seoul National University, Seoul 151-742, Korea*

<sup>2</sup>*Applied Vegetation Dynamics Laboratory, School of Environmental Sciences, University of Liverpool, Liverpool, L69 3GP, UK*

We described the above-ground plant species composition and measured a range of soil physico-chemical properties and the composition and size of the soil seed bank in the remnant natural vegetations on the flood plains of the Han River within Seoul, South Korea. We used analysis of variance and multivariate analyses to analyse the data and Sørensen's similarity index to compare the composition of the vegetation and seed banks. The soils were circum-neutral and composed of mainly sand and silt fractions with a very limited clay component; a gradient based on sand/clay proportions was identified. The soil seed banks varied markedly between- and within-sites and had much greater species diversity than the above-ground vegetation. Two of the major dominants in the vegetation (*Miscanthus sacchariflorus* and *Phragmites australis*) were found at very low densities in the seed bank. The site differences appeared to be correlated with the sand-clay gradient, suggesting that the soil properties differentially affected seed inputs into the soil, or that the processes that controlled sediment deposition during floods was also important in differentially affecting seed deposition. Lastly, there was relatively little similarity between the vegetation, dominated mainly by perennials, and the seed bank which contained a relatively large proportion of annuals and biennials. This result suggests that after disturbance caused by flooding there is the potential for many other species to colonize. This may impinge on the regeneration potential of the sites and cause concern for the future conservation of these important remnants of natural vegetation.

**Key words :** soil physico-chemical properties, regeneration, Sørensen's similarity index, life-history traits, conservation

## INTRODUCTION

Riverine systems are important ecosystems from a conservation perspective and are under great threat from development especially in urban areas (Karr and Chu, 2000; Paul and Meyer, 2001; Gergel *et al.*, 2002), where damaging impacts range from complete removal, through various

stages of re-alignment and embankment to direct and diffuse pollution from a variety of sources (Ward *et al.*, 2001). Where there is some natural vegetation present in refuges there is a need to gain a better understanding of the ecological processes that occur in these systems. At the very basic level there is a need to describe the vegetation that occurs in such sites and then to consider processes that will impact on their potential rege-

\* Corresponding author: Tel: 02) 880-6673, Fax: 02) 872-6681, E-mail: ejlee@snu.ac.kr

neration such as the size and composition of the seed bank (Nilsson *et al.*, 1991; Hanlon *et al.*, 1998; Tabacchi *et al.*, 2005; Capon and Brock, 2006). The soil seed bank provides a stock of regeneration potential in many plant communities (Dessaint *et al.*, 1997; Richter and Stromberg, 2005) and is an important component of ecosystem resilience (Thompson, 2000).

The soil seed bank is especially important in flood plains because these ecosystems are affected on a regular basis by disturbance, through periodic flooding (Abernethy and Willby, 1999; Richardson *et al.*, 2007), summer drought (Liu *et al.*, 2006), fire and in some cases cultivation (Leck and Leck, 1998; Holmes, 2002). This disturbance creates bare ground for the colonization by new individuals or species derived either from the seed bank or dispersing via the water (Thompson and Grime, 1979; Richter and Stromberg, 2005; James *et al.*, 2007). Thus, a new succession is created. Flood plain vegetation is, therefore, constantly in transition and the soil seed bank is an important part of the ecosystem that will assist vegetation recovery (Ghorbani *et al.*, 2003; González-Alday *et al.*, 2009), and will be an important mechanism in regulating the biodiversity of flood plain ecosystems.

In South Korea there is a lack of information on the vegetation of flood plain systems and the regeneration potential available within the soil seed banks. The Han River is a major river, and the second longest river in South Korea. The river has been subject to substantial technological management for the control of floods during the heavy monsoonal rainfall period, and for securing irrigation water during drought periods (Woo, 2010). Within the city of Seoul the river has been channelized and embanked since the mid-1960s. Since the 1980's the narrow flood plain of the river has been developed as public and ecological parks, roads, and as car parks (Woo, 2010). As a result, the previous ecological functions derived from the natural riparian vegetation have been lost within the urban reach of the Han River. However, there are a few refuge areas within the confines of the city where natural vegetation remains or has regenerated after construction work. The remnant vegetation in the disturbed flood plains are very important with high ecological and environmental values from a conservation viewpoint due to their potential values of the species source and the restoration prototype for the future river restora-

tion (Kim and Ju, 2005; Jeon *et al.*, 2008).

The aims of this study were, therefore, to: (1) provide a description of the vegetation in the refuges within the Seoul stretch of the Han River, (2) assess the size and composition of the seed bank in each area, and this was determined by germinating the seedling in the soil (Dalling *et al.*, 1995; Hanlon *et al.*, 1998; Ter Heerdt *et al.*, 1999; Ghorbani *et al.*, 2003; Kim and Ju, 2005; Richter and Stromberg, 2005; González-Alday *et al.*, 2009), and (3) to test whether there was a correlation between the vegetation and soil seed bank. We hope that this preliminary assessment would provide information on the natural potential vegetation and soil seed bank in the urban food plains under severe disturbance to help inform their conservation planning.

## METHODS

### 1. Study sites

This study was conducted on the flood plains of the Han River in Seoul, South Korea. The urban reach of the river in the Seoul metropolitan area has been subject to technological management such as channelization, embankment, damming and developing the leisure facility. There are many dams upstream from Seoul used for flood control, water supply and power generation (for example, the Soyang, Chungju and Paldang Dams), and at the seaward end there are two under-water barriers, the Shingok and Jamsil weirs, which regulate salt-water inflow and maintain water levels (Kim, 2008).

Four study sites, the only ones with semi-natural vegetation remaining, were selected along the flood plains of the Han River which goes through Seoul City (Fig. 1, Table 1). Janghang is located downstream of the Shingok weir and therefore has a tidal influence (Kim, 2008). This site has a flat, silty shore line and there is a succession of plant communities from the waterfront to the landward edge, dominated in order by *Phragmites australis*, *Salix* spp. and *Miscanthus sacchariflorus*. The *Salix* community, comprised of *S. nipponica* and *S. purpurea* var. *japonica*, has developed into the largest *Salix* forest along the Han River. The other three sites are above the Shingok weir and hence are unaffected by tidal inundation. The flood plain wetland in Kangseo was used for agri-

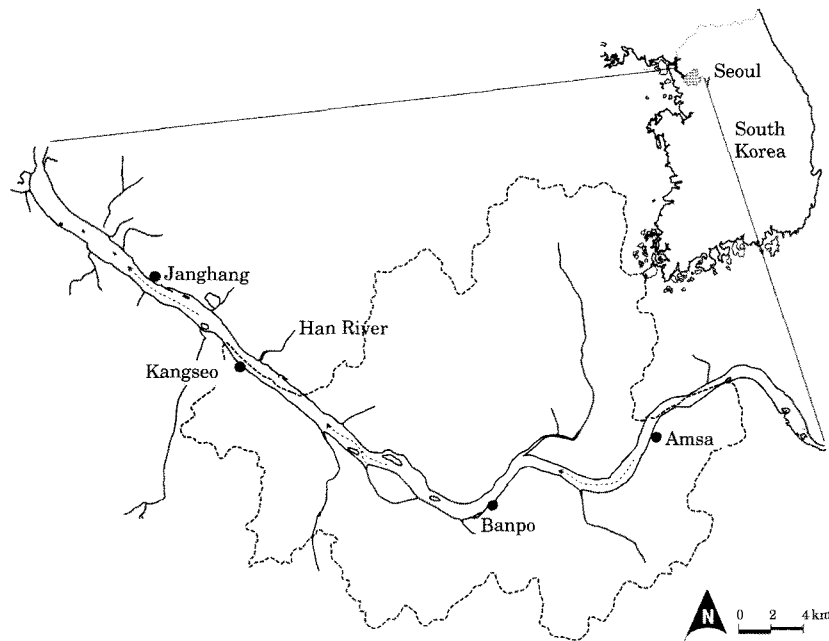


Fig. 1. Location of study sites along the Han River in Seoul, South Korea.

culture until 2006 and since then has been developed into a wetland ecological park. There was a small amount of extant natural vegetation (*Phragmites australis*, *Miscanthus sacchariflorus* and *Salix* spp.) but at the time of sampling time had communities dominated by invasive alien plant species such as *Ambrosia trifida*. The dominant communities at Kangseo were *Ambrosia trifida*, *Miscanthus sacchariflorus*, *Phragmites australis*, *Phalaris arundinacea* and *Salix* woodland (*S. koreensis*, *S. nipoponica* and *S. pseudo-lasiogyne*). The Banpo wetland is within a public park that is mainly used for leisure activity and car parking, but there is a small area with natural flood plain vegetation. Here the plant communities were dominated by *Bidens frondosa*, *Humulus japonica*, *Miscanthus sacchariflorus* and *Phragmites australis*. Finally, the Amsa wetland (0.1 km<sup>2</sup>) was designated as a conservation area in 2002, and has been fenced to prevent human access. Amsa is close to natural riverine wetland upstream of the Shingok weir, and the communities were dominated by *Chenopodium album* var. *centrorubrum*, *Chenopodium ficifolium*, *Humulus japonica*, *Miscanthus sacchariflorus*, *Persicaria perfoliata*, *Phragmites australis* and *Salix* woodland (*S. koreensis*, *S. nipoponica* and *S. pseudo-lasiogyne*).

In each of the study sites there were communities dominated by either *Phragmites australis* or

Table 1. The four river flood plain study sites along the Han River, Seoul, South Korea, along with the dominant vegetation types sampled; sites are in seaward to landward order (Fig. 1).

Flood plain site	Dominant plant community sampled	Code
Janghang	Waterfront	WF
	<i>Phragmites australis</i>	Pa
	<i>Salix</i> spp.	Ss
	<i>Miscanthus sacchariflorus</i>	Ms
Kangseo	<i>Phragmites australis</i>	Pa
	<i>Salix</i> spp.	Ss
	<i>Miscanthus sacchariflorus</i>	Ms
	<i>Ambrosia trifida</i>	At
Banpo	<i>Phragmites australis</i>	Pa
	<i>Miscanthus sacchariflorus</i>	Ms
Amsa	Waterfront	WF
	Low Channel Bank	Bn
	<i>Salix</i> spp.	Ss
	<i>Phragmites australis</i>	Pa
	<i>Miscanthus sacchariflorus</i>	Ms

*Miscanthus sacchariflorus*. At each site between 2~5 different plant community types were identified and a total of 15 plant communities were sampled (Table 1). The presence of all higher plant species was recorded in five 0.25 m<sup>2</sup> quadrats positioned randomly in each of the 15 plant communities.

## 2. Assessment of soil physico-chemical properties

Soils were sampled in March 2005 for soil chemical analysis. Within each of the 15 plant communities (Table 1) seven positions were located randomly and soil samples collected for analysis. Soil pH and electrical conductivity were measured in a 1:2 mixture of fresh soil in distilled water. Soil moisture content was then measured by weighing before and after drying a sub-sample at 105°C for 48 hours. The soil organic matter content was estimated by loss-on-ignition (Heiri *et al.*, 2001). Soil texture analysis was carried out by shaking 40 g in 300 mL of 5% sodium hexametaphosphate solution overnight. Following dilution to 1,000 mL with the distilled water and shaking for 1 minute at 60 rpm density measurements were made using a Bouyoucos hydrometer (USA/ASTM 151H) after 40 seconds and 7 hours. Isoamyl alcohol was fallen in drops for removing bubbles if necessary (Sheldrick and Wang, 1993). Total soil nitrogen was analysed using a protein/nitride auto analyzer (Kjeldahl Protein/Nitrogen Analyzer (Automatic); Kjeltec Auto 1035/1038 System). Extractable soil phosphorus (EP) and exchangeable cations (Na, Mg, K, Ca) were extracted using Mehlich III solution (Sims *et al.*, 2002); EP was analyzed using the ascorbic acid method (APHA, 1992) and exchangeable cations using an induced coupled plasma (ICP) Emission Spectrometer (Shimadzu/ICPS-7510).

## 3. Assessment of species in the seed bank

The main purpose of this study was to determine the size and composition of the germinable seed bank, and therefore the soils were sampled in March 2006 at the beginning of Spring when all seed available for germination during the sampling year should be available. Within each of the 15 plant communities (Table 1) five replicate 1 m<sup>2</sup> quadrats were located randomly. At each of these quadrats five soil cores were taken using a soil corer (Eijkelkamp BV, Netherlands; corer dimensions=5 cm diameter, 5 cm deep). As the seed density is usually greatest in the uppermost soil layer (Roberts, 1981; Galatowitsch and Valk, 1996) we sampled the surface layer of soil with 5 cm deep core. The layers were pooled by quadrat to produce a surface and sub-surface sample for each quadrat. There were in total 75 soil samples for seed bank analysis (15 communities × 5 repli-

cate quadrats). All samples were stored in at 4°C until the seedling emergence experiment started (Looney and Gibson, 1995).

Initially, each soil was prepared by hand-removal of roots and plant stems from the sample.

Thereafter the soils were placed in individual seed trays (44 cm × 30 cm × 7 cm) and transferred to the glasshouse. The soils were watered daily. The minimum and maximum temperatures recorded in the glasshouse were 10.4°C and 35.8°C respectively. Seedling emergence was observed for 16 weeks. Emergent seedlings were identified using (Asano, 2005) and counted every two weeks; where identification was not possible the seedling was transferred to a separate pot and grown on until identification was possible. The seed numbers emerging were converted to seed densities on an area basis (m<sup>-2</sup>).

## 4. Statistical analysis

The species detected in each of the 15 communities are presented in full in Appendix I. Summary data on number of species (total and sub-divided into annuals, biennials and perennials) were analysed using analysis of variance (ANOVA; SAS Institute, 2009). The inter-relationships of the soil environmental physico-chemical factors were investigated using the principle component analysis (PCA; SAS Institute, 2009). Multivariate analysis of the soil seed bank species composition was also carried out unconstrained analysis. Decorana (DCA, Hill and Gauch, 1980) was used to analyse the species composition using the Vegan package (function 'decorana', Oksanen, 2005) within the R statistical environment (Team, 2008). Species present in less than five quadrats were excluded and the downweighting option was used. The environmental factors were then correlated with the DCA ordination using the 'envfit' function within Vegan (Oksanen, 2005); here significance was assessed using a permutation test ( $n=1,000$ ) and only variables significant ( $P<0.05$ ) are discussed. The species composition of the vegetation and seed bank were compared using Sørensen's similarity index (Hopfensperger, 2007; Boudell and Stromberg, 2008).

# RESULTS

## 1. Above-ground vegetation

The total number of species in each community

**Table 2.** Number of plant species found in the above-ground vegetation at the 15 communities along the Han River, Seoul, South Korea. Total numbers are presented along with numbers and % of the species in each life-history category (annual, biennial and perennial plants). Community codes are given in Table 1.

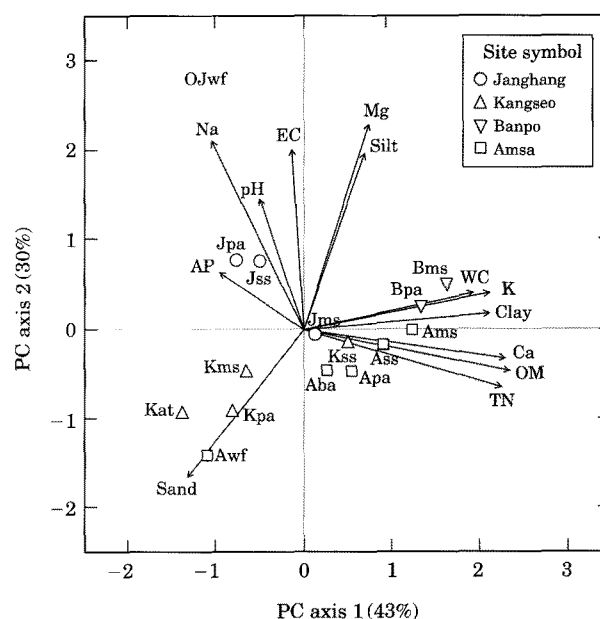
Life history		Janghang				Kangseo			Banpo		Amsa					
		WF	Pa	Ss	Ms	Pa	Ss	Ms	At	Pa	Ms	WF	Bn	Ss	Pa	Ms
Annuals	Species no	1	4	3	4	10	14	11	8	11	9	15	20	27	13	4
	%	25	36	30	19	33	47	41	38	50	47	44	53	48	48	33
Biennials	Species no	1	2	4	10	10	7	9	6	6	7	9	10	12	7	5
	%	25	18	40	48	33	23	33	29	27	37	26	26	21	26	42
Perennials	Species no	2	5	3	7	10	9	7	7	5	3	10	8	17	7	3
	%	50	45	30	33	33	30	26	33	23	16	29	21	30	26	25
Total		4	11	10	21	30	30	27	21	22	19	34	38	56	27	12

varied between 4~56 with a mean of 24 in five 0.25 m<sup>2</sup> quadrats (Table 2 and Appendix 1). The site nearest the sea (Janghang) had the lowest species number (4~21) and the greatest species numbers were found in Kangseo and Amsa which had the greatest protection. Banpo was intermediate and this is a small site with considerable human disturbance. The distribution of life-forms types varied also between the different communities. The disturbed Banpo and Amsa sites had a relatively high proportion of annuals (33~53%), otherwise there was no real consistent pattern in the distribution of life-form types (Table 2).

## 2. Soil sampling and analysis of physico-chemical properties

The soils were circum-neutral with a pH between 6.4 and 7.4 and this was reflected in relatively high concentrations of exchangeable cations in particular Ca and Mg. Exchangeable Na was only elevated at the waterfront community at Janghang, the site nearest the sea (Table 3). The soils were predominantly composed of sand and silt with a maximum of 9% clay content. The organic matter content and total nitrogen content was also relatively low peaking at 9% and 2% respectively.

The first two axes of the PCA of the soil data produced eigenvectors of 5.439 and 3.909 which explained 43% and 30% of the variance; two obvious gradients were detected on these axes (Fig. 2). The first axis represented a gradient from sandy conditions at the negative end through to soils with a greater content of clay, organic matter, total nitrogen and exchangeable K and Ca at the positive end. Also correlated with the positive end was an increased water content. Axis 2 showed a

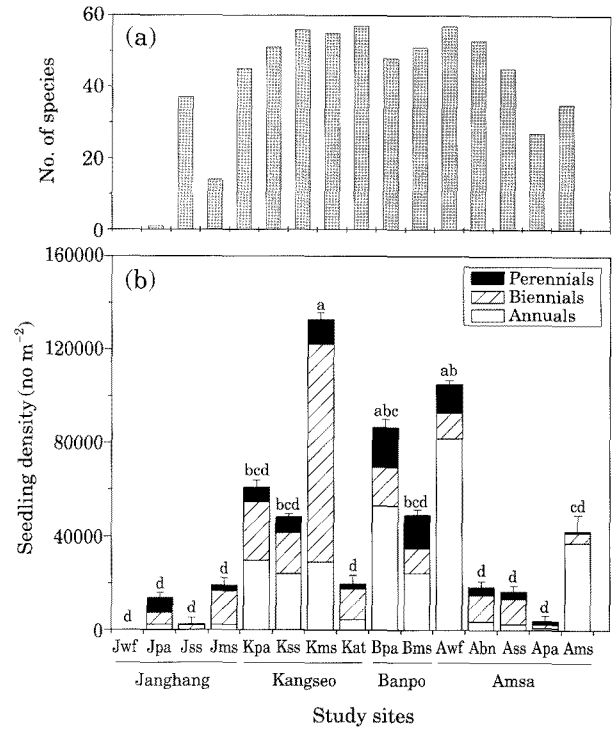


**Fig. 2.** Ordination biplot derived from the first two axes of principal components analysis (PCA) of the soil data from 15 plant communities along the Han River, Seoul, South Korea. Site codes: A=Amsa; B=Banpo; J=Janghang; K=Kangseo. Community codes: at=*Ambrosia trifida*; WF=waterfront; bn=Bank; pa=*Phragmites australis*; ms=*Miscanthus sacchariflorus*; ss=*Salix* spp. Soil chemical property codes: EP=extractable phosphorus; EC=electrical conductivity; OM=organic matter; TN=total nitrogen; WC=water content; Na=exchangeable Na; Mg=exchangeable Mg; K=exchangeable K; Ca=exchangeable Ca.

gradient of increasing soil pH, electrical conductivity silt content and high exchangeable Mg and exchangeable P (EP) concentrations. The most maritime sites were correlated with high electrical conductivity, pH and Na, Banpo the most dis-

**Table 3.** Physico-chemical properties of the soils at 15 plant communities along the Han River, Seoul, South Korea. The soil properties are coded: WC=Water content; OM=Organic matter content; EC=Electric conductivity; TN=Total Nitrogen; EP=Extractable Phosphorous. Community codes (Com) are given in Table 1. Mean values (mean ± SE; n=7).

Site	Com	pH	WC			Sand (%)	Silt	Clay	TN (mg g <sup>-1</sup> )	EC (mS m <sup>-1</sup> )	EP	K	Na (µg g <sup>-1</sup> )	Ca	Mg
			WC	OM	EC										
Janghang	WF	7.4±0.04	24.7±0.5	1.2±0.1	31±4	64±3	6±0	0.2±0.02	61.4±3.5	27.8±0.5	190±5	823±24	532±23	439±4	
	Pa	7.2±0.06	26.9±0.5	2.3±0.1	30±1	65±1	4±0	0.3±0.05	11.4±0.6	20.5±1.1	164±9	479±10	710±7	295±8	
	Sa	6.5±0.07	27.0±0.3	2.1±0.1	20±1	72±1	8±0	0.4±0.04	25.0±1.7	22.9±0.8	148±4	473±12	881±73	298±24	
	Ms	6.4±0.10	27.4±0.5	4.7±0.4	28±4	56±3	17±1	1.0±0.07	11.7±1.6	24.9±1.8	210±23	426±5	1076±99	241±25	
Kangseo	Pa	6.7±0.08	21.1±2.1	2.6±0.5	53±5	38±4	9±1	0.8±0.09	8.2±0.5	20.9±2.9	99±12	366±8	1048±59	95±6	
	Ss	6.7±0.03	26.5±1.5	6.0±0.5	36±3	49±3	16±1	1.3±0.14	19.7±0.8	25.8±1.1	258±4	393±4	2085±129	184±7	
	Ms	6.6±0.08	23.7±0.5	4.1±0.2	51±3	41±4	7±1	0.9±0.07	17.4±1.8	31.6±2.5	197±14	355±7	907±46	129±19	
	At	6.8±0.06	12.3±0.4	2.6±0.1	68±3	24±2	9±1	0.4±0.02	12.1±1.1	28.2±2.3	168±8	399±4	881±44	80±6	
Banpo	Pa	6.6±0.07	33.3±0.2	6.8±0.1	14±1	59±1	27±0	1.4±0.04	22.7±1.3	20.3±1.0	281±17	349±2	1993±32	285±10	
	Ms	6.9±0.05	35.6±0.3	8.7±0.1	15±2	59±2	26±1	1.4±0.05	25.9±9.0	21.5±0.8	370±41	357±4	1988±29	296±6	
Amsa	WF	6.7±0.05	29.5±1.3	3.0±0.4	85±2	12±1	3±1	0.6±0.05	13.6±1.6	16.8±1.4	62±6	368±3	662±83	100±11	
	Bn	6.7±0.03	31.7±0.6	5.9±0.3	57±5	34±4	9±1	1.3±0.14	25.6±1.3	18.6±0.6	233±14	363±1	1476±59	228±12	
	Ss	6.8±0.03	29.9±0.4	6.0±0.2	47±3	42±2	11±1	1.9±0.09	23.3±1.7	16.9±1.0	389±33	386±11	1888±68	252±18	
	Pa	6.9±0.02	30.1±1.7	6.0±0.5	50±7	39±5	11±2	1.5±0.20	16.9±2.3	18.9±1.4	335±42	370±4	1533±171	186±25	
	Ms	6.7±0.05	34.0±0.4	8.2±0.3	33±3	50±3	17±1	1.8±0.11	14.3±0.6	21.3±1.5	307±24	412±12	1925±89	289±23	



**Fig. 3.** Summary on seed bank composition from 15 plant communities along the Han River, Seoul, South Korea. (a) species number, and (b) total seed density (number of seedlings m<sup>-2</sup>), separated into different life-history strategies. Communities are coded (at: *Ambrosia trifida*; ba: Bank; pa: *Phragmites australis*; ms: *Miscanthus sacchariflorus*; ss: *Salix* spp.; wf: waterfront). Values are mean density ± standard error (n=5). The same letter are not significantly different by Tukey's test ( $\alpha=0.05$ ).

turbed site had been correlated with both a high clay content and water holding capacity, Kangseo was correlated with high sand content and Amsa showed a gradient across the communities from sandy soils through to soils with a relatively greater clay and silt content.

### 3. Seed banks

The diversity of the seed banks in terms of species number varied markedly both between and within sites (Fig. 3). Kangseo and Banpo were the most diverse having between 42 ~ 57 species in all communities, Amsa was intermediate; there three communities (water front, river bank, *Salix* spp.) had more than 40 species but the *Phragmites australis* and *Miscanthus sacchariflorus* communities had 35 and 30 species respectively. The

**Table 4.** Similarity between the species detected in the seed bank with and field vegetation in 15 plant communities along the Han River, Seoul, South Korea. The number of species in each, the number of common species and Sørensen's similarity index are presented. Community codes are given in Table 1.

Contents	Site																			
	Janghang					Kangseo					Banpo			Amsa						
	WF	Pa	Ss	Ms	Total	Pa	Ss	Ms	At	Total	Pa	Ms	Total	WF	BN	Ss	Pa	Ms	Total	
Species no																				
in																				
Seed bank	1	36	13	45	61	51	56	54	55	82	48	49	61	57	52	44	27	34	83	
Vegetation	4	11	10	21	39	30	30	27	21	55	22	19	24	34	38	56	27	12	75	
Common	0	4	2	9	18	17	15	20	16	42	13	11	15	19	24	26	11	8	50	
Sørensen's similarity index (%)	0	17	17	27	36	42	35	49	42	61	37	32	35	42	53	52	41	35	63	

Janghang site had the lowest species number: two communities had fewer than five species (Water front and *Phragmites australis*), the *Miscanthus sacchariflorus* one had 12 and only the *Salix* dominated community had a modest species count (35 species).

The results for the total number of seedlings to some extent reflected these sites and community differences (Fig. 3). Overall Kangseo and Banpo had greater than 40,000 seeds  $m^{-2}$  in all but one community (*Ambrosia trifida* at Kangseo). Amsa had more than 40,000 seeds  $m^{-2}$  in two communities, the waterfront and the *Miscanthus sacchariflorus* communities and much less than 20,000 seeds  $m^{-2}$  in the other three communities. Janghang had less than 20,000  $m^{-2}$  seeds in all communities, and seedling densities were especially low at the water front. The distribution of life forms also differed between sites. At Kangseo and Banpo there was a mixture of annuals, biennials and perennials (mainly biennials), whereas at Janghang and Amsa most communities were composed of mainly biennials and perennials. There were few annuals at Janghang. However, the *Miscanthus sacchariflorus* community at Amsa had a substantive annual component, but whilst the seed density of annuals in this community was high it was mainly of a single species (*Chenopodium ficifolium*).

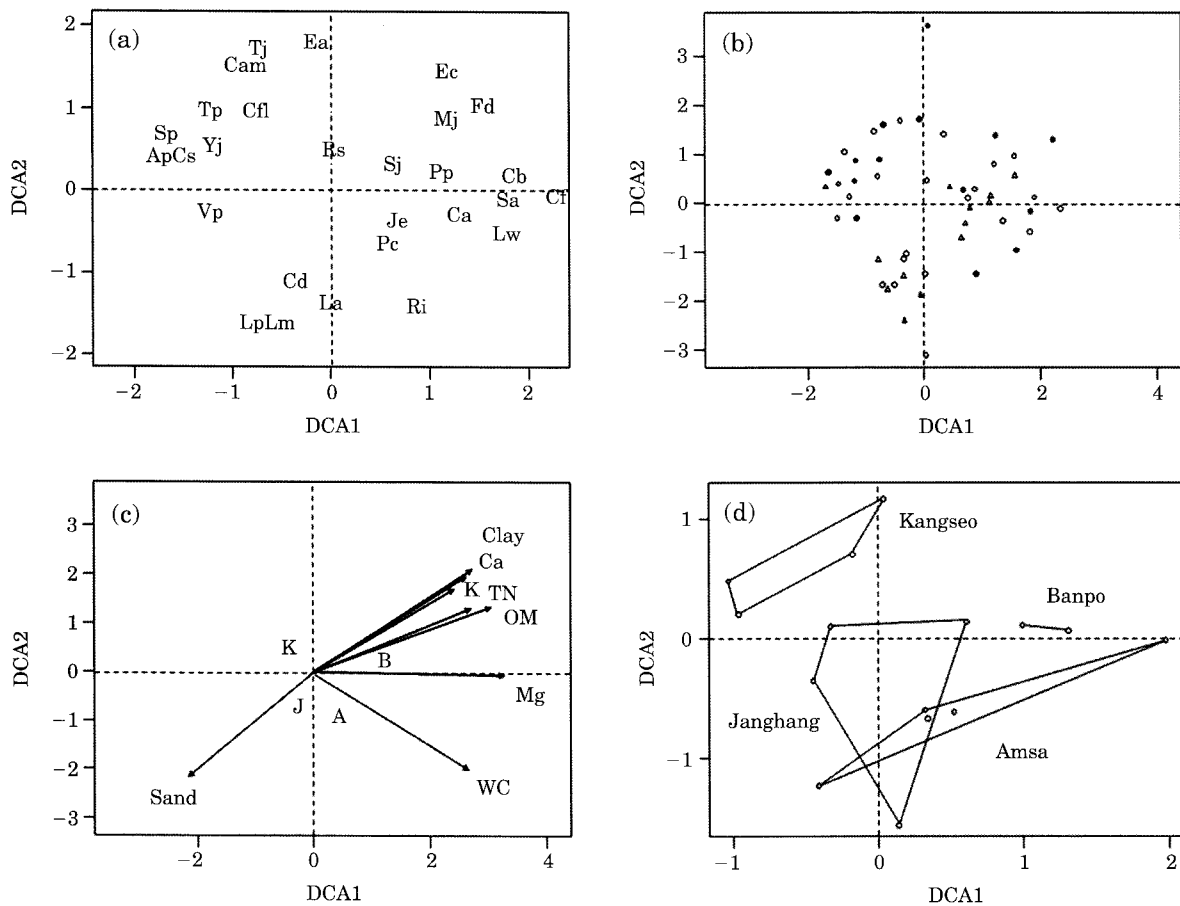
The seed bank showed a relatively poor similarity with the vegetation data (Table 4). At the site level Janghang and Banpo had the lowest similarity coefficients (35~36%) whereas Janghang and Amsa had greater values (61~63%). However, at the individual community level the similarity coefficients were much small ranging from 0~27% at Janghang to a maximum of 53% on the bank at Amsa.

**Table 5.** Correlation between the soil physico-chemical properties and the first two axes of the Decorana analysis of the soil seed bank composition in 15 plant communities along the Han River, Seoul, South Korea. The correlation coefficients were derived from a permutation test ( $n=1,000$ ). The site centroids were also significantly correlated with the species composition ( $r^2=0.6589$ ;  $P=0.021$ ). Significance level: \*\*\* $P<0.001$ , \*\* $P<0.01$ , \* $P<0.05$ .

Soil variable	Axis 1	Axis 2	$r^2$	$P$
Clay content	0.794	0.608	0.759	0.004**
Exchangeable Mg	1.000	-0.013	0.735	0.007**
Water content	0.799	-0.602	0.737	0.006**
Exchangeable Ca	0.797	0.604	0.738	0.014*
Exchangeable K	0.816	0.578	0.579	0.046*
Organic matter content (LOI)	0.916	0.400	0.728	0.011*
Sand content	-0.715	-0.699	0.593	0.044*
Total N	0.902	0.432	0.591	0.040*
Extractable P	-0.459	0.889	0.512	0.102
Electrical conductivity	0.984	0.177	0.255	0.345
Exchangeable Na	0.444	-0.896	0.024	0.903
pH	0.509	-0.861	0.099	0.735
Silt content	0.655	0.756	0.425	0.162

#### 4. Relationship between the seed bank composition and soil characteristics

The DCA analysis produced eigenvalues of 0.567, 0.394, 0.292 and 0.073 and gradient lengths of 3.04, 2.71, 2.33 and 2.68 for the first four axes respectively. The distribution of species in the seed bank was not easily explained in terms of the three major species composition or the three life-forms tested (Fig. 4a). Apart from two outlier species, the annual species *Digitaria violascens* and the biennial species *Oenothera odorata*, there was



**Fig. 4.** Detrended correspondence analysis (DCA) ordination of the seed bank species in the 15 plant communities along the Han River, Seoul, South Korea. (a) the most abundant species; (b) all species with more than 5 occurrences displayed by life-form ( $\circ$ =annuals,  $\bullet$ =biennials,  $\triangle$ =perennials); (c) significant ( $P < 0.05$ ) environmental variables, these are coded: OM=organic matter, TN=total nitrogen, WC=water content, Mg=exchangeable Mg, K=exchangeable K, Ca=exchangeable Ca, sites coded: A=Amsa, B=Banpo, J=Janghang, K=Kangseo; (d) community centroids linked in site clusters. Species are shown: Ap=*Artemisia princeps* var. *orientalis*; Cam=*Cyperus amuricus*; Cb=*Capsella bursa-pastoris*; Cd=*Cyperus difformis*; Cf=*Chenopodium ficifolium*; Cfl=*Cardamine flexuosa*; Cs=*Cyperus sanguinolentus*; Ea=*Erigeron annuus*; Ec=*Erigeron canadensis*; Fd=*Fimbristylis dichotoma*; Je=*Juncus effusus* var. *decipiens*; La=*Lindernia attenuata*; Lm=*Lindernia micrantha*; Lp=*Lindernia procumbens*; Lw=*Ludwigia prostrata*; Mj=*Mazus japonicus*; Pc=*Potentilla paradoxa*; Pp=*Poa sphondylodes*; Ri=*Rorippa islandica*; Rs=*Ranunculus sceleratus*; Sa=*Stellaria aquatica*; Sj=*Sagina japonica*; Sp=*Salvia plebeia*; Tj=*Torilis japonica*; Tp=*Trigonotis peduncularis*; Vp=*Veronica peregrina*; Yj=*Youngia japonica*.

considerable overlap between the three groups (Fig. 4b). However, there were significant differences between the four sites tested and some of the soil physico-chemical properties (Table 5); there was a major gradient across the sites based on substrate physical composition across the ordination space (Fig. 4c) from a more sandy soil (lower left quadrant) to a more clay soil (upper right quadrant). The high clay content is also correlated with high concentrations of exchangeable Ca and Mg and organic matter and nitrogen content.

The water holding capacity of the soils measured at the time of sampling was orthogonal to this main gradient. The four sites and their constituent communities occupied different parts of the ordination space showing a gradation from Janghang and Banpo at the sandy end of the spectrum through to Kangseo and Amsa with a greater clay content (Fig. 4d). Amsa, however, showed the largest variation along this sand-clay gradient.



## DISCUSSION

The flood plain ecosystems along the urban stretch of the Han River within Seoul has been considerably altered by river engineering (Kim, 2008). There are few remnants of flood plain vegetation remaining and these represent small foci of natural habitats within this managed system which provide some conservation interest (Kim and Ju, 2005). Here, fifteen different community types were identified across four of these remnant sites and there were considerable difference in plant species number in these areas. The soils in the communities were sampled and their physico-chemical properties analysed. The soils were composed mainly of sand and silt with a maximum of 9% clay content. The organic matter content was relatively low and they were circum-neutral. Multivariate analysis identified sand-clay and sand-silt gradients across these communities, and there was an increase in most physico-chemical properties associated with the soils with the greatest silt and clay contents (Cho, 1995).

The soil seed banks varied markedly between- and within-sites and this will almost certainly impinge on regeneration after disturbance, a common feature in flood plain situations (Dessaint *et al.*, 1997; Richter and Stromberg, 2005). In most communities studied here, the seed bank composition had a greater species richness than the above-ground vegetation; this is in contrast to some other studies where more species were detected in the above-ground vegetation (Valk and Davis, 1978; Thompson and Grime, 1979; Bakker and Berendse, 1999). At the site level there was a gradation in terms of species number and seed density in the seed bank from the site with the most saline influence having the lowest at Kanseo and Banpo having the greatest. The seed banks were dominated by a few species and the species were mainly annual and biennial ones; this appears typical; of flood plain soils (Van Der Valk and Davis, 1979; McDonald *et al.*, 1996; Abernethy and Willby, 1999; Richter and Stromberg, 2005; Boudell and Stromberg, 2008). For example, the seed bank in the perennial *Miscanthus sacchariflorus* community at the Amsa was dominated by the annual *Chenopodium ficifolium* (84% of the total seed density). Even where *Miscanthus sacchariflorus* and *Phragmites australis* were the dominant species in the vegetation, these species

had a very low seedling density in the seed bank (Van Der Valk and Davis, 1979; Baldwin *et al.*, 2010). It is possible that some of the species found in the above-ground vegetation did not emerge as seedlings either because they have short-lived seed longevity (*Salix* spp., Niiyama, 1990; Karrenberg *et al.*, 2002) or the experimental conditions were not suitable for their germination, i.e. the aquatic species, *Spirodela polyrhiza*. However, the tests used here provide a reasonable approximation of the geminate seed in the soil bank that will be available for regeneration after disturbance.

One of the main results to emerge from this study was that the seed bank showed a very poor similarity with the overlying vegetation with a maximum of 53% on the bank at Amsa. This implies that there is only limited potential for direct regeneration of the current vegetation from the seed bank. In all communities, there was a mixture of annuals, biennials and perennials with some site and community differences. One reason for this low similarity is that the riverine vegetation was dominated by perennial species (e.g. *Miscanthus sacchariflorus*, *Phragmites australis* and *Salix* spp.), whereas the seed banks contain a relatively high number of annual and biennial species.

The site differences in seed bank composition were highlighted in the unconstrained multivariate analysis where the sites were clearly separated in seed bank species space. Moreover, the distribution was correlated with a sand-clay gradient which also reflected a fertility gradient for most soil variables measured. This is an intriguing result because the soil seed bank in these riverine systems must have two major inputs, from aquatic sources and from the overlying vegetation. Given the relatively weak similarity between seed bank and vegetation this result suggests that deposition processes results in differential seed addition to the seed bank. Clearly aquatic-borne inputs will depend on (a) upstream distance from natural habitat sources (and downstream distance in the case of the tidal Janghang), (b) current flows, and (c) the vegetation structure of the ecosystem which will regulate sediment (and seed) capture. The results here suggest the process that control sediment deposition also affects the deposition of seed of different species differentially. Irrespective, there are clear differences in regeneration potential in the different sites and communities and

this will inevitably impinge on the resilience of these flood plain ecosystems to disturbance.

From a management perspective it is very difficult to be able to predict the way communities will recover after major disturbance. The perennial species will be able to regenerate from stem bases (Uchida *et al.*, 2003; Cho and Cho, 2005) but the potential for colonization by species that are currently not present in the above-ground vegetation is considerable and at the moment unpredictable. Further experimental studies are, therefore, needed to assess the effects of such disturbance on both the vegetation and the role of the seed bank in its regeneration.

### ACKNOWLEDGEMENTS

This study was supported by the 2006 Core Construction Technology Development Project (06 KSHS-B01) through ECORIVER21 Research Center in KICTTEP of MOCT Korea. We thank Dr. Uham Song, Jung Soo Park and Kyu Lee from the Plant Ecology Laboratory of Seoul National University for technical support.

### LITERATURE CITED

- Abernethy, V.J. and N.J. Willby. 1999. Changes along a disturbance gradient in the density and composition of propagule banks in floodplain aquatic habitats. *Plant Ecology* **140**: 177-190.
- APHA. 1992. Standard Methods for the Examination of Water and Wastewater. 18<sup>th</sup> edition. APHA, Washington, D.C.
- Asano, S. 2005. Seeds/Fruits and Seedlings of Plants in Japan. Zenkoku-Noson-Kyoiku-Kyokai, Tokyo.
- Bakker, J.P. and F. Berendse. 1999. Constraints in the restoration of ecological diversity in grassland and heathland communities. *Trends in Ecology & Evolution* **14**: 63-68.
- Baldwin, A.H., K.M. Kettenring and D.F. Whigham. 2010. Seed banks of *Phragmites australis* dominated brackish wetlands: Relationships to seed viability, inundation, and land cover. *Aquatic Botany* **93**: 163-169.
- Boudell, J.A. and J.C. Stromberg. 2008. Flood pulsing and metacommunity dynamics in a desert riparian ecosystem. *Journal of Vegetation Science* **19**: 373-380.
- Capon, S.J. and M.A. Brock. 2006. Flooding, soil seed bank dynamics and vegetation resilience of a hydrologically variable desert floodplain. *Freshwater Biology* **51**: 206-223.
- Cho, D.S. 1995. A study on the distribution of stream-side vegetation in Kyonganchon. *Korean Journal of Ecology* **18**: 55-62.
- Cho, H.-J. and K.-H. Cho. 2005. Responses of riparian vegetation to flooding disturbance in a sand stream. *KSCE Journal of Civil Engineering* **9**: 49-53.
- Dalling, J.W., M.D. Swaine and N.C. Garwood. 1995. Effect of soil depth on seedling emergence in tropical soil seed-bank investigations. *Functional Ecology* **9**: 119-121.
- Dessaint, F., R. Chadoeuf and G. Barralis. 1997. Nine years' soil seed bank and weed vegetation relationships in an arable field without weed control. *Journal of Applied Ecology* **34**: 123-130.
- Galatowitsch, S.M. and A.G.v.d. Valk. 1996. The vegetation of restored and natural prairie wetlands. *Ecological Applications* **6**: 102-112.
- Gergel, S.E., M.G. Turner, J.R. Miller, J.M. Melack and E.H. Stanley. 2002. Landscape indicators of human impacts to riverine systems. *Aquatic Sciences-Research Across Boundaries* **64**: 118-128.
- Ghorbani, J., P.M. Das, A.B. Das, J.M. Hughes, H.A. McAllister, S.K. Pallai, R.J. Pakeman, R.H. Marrs and M.G. Le Duc. 2003. Effects of restoration treatments on the diaspore bank under dense *Pteridium* stands in the UK. *Applied Vegetation Science* **6**: 189-198.
- González-Alday, J., R.H. Marrs and C. Martínez-Ruiz. 2009. Soil seed bank formation during early revegetation after hydroseeding in reclaimed coal wastes. *Ecological Engineering* **35**: 1062-1069.
- Hanlon, T.J., C.E. Williams and W.J. Moriarity. 1998. Species composition of soil seed banks of allegheny plateau riparian forests. *Journal of the Torrey Botanical Society* **125**: 199-215.
- Heiri, O., A.F. Lotter and G. Lemcke. 2001. Loss on ignition as a method for estimating organic and carbonate content in sediments: reproducibility and comparability of results. *Journal of Paleolimnology* **25**: 101-110.
- Hill, M.O. and H.G. Gauch. 1980. Detrended correspondence analysis: An improved ordination technique. *Plant Ecology* **42**: 47-58.
- Holmes, P.M. 2002. Depth distribution and composition of seed-banks in alien-invaded and uninvaded fynbos vegetation. *Austral Ecology* **27**: 110-120.
- Hopfensperger, K.N. 2007. A review of similarity between seed bank and standing vegetation across ecosystems. *Oikos* **116**: 1438-1448.
- James, C., S. Capon, M. White, S. Rayburg and M. Thoms. 2007. Spatial variability of the soil seed bank in a heterogeneous ephemeral wetland system in semi-arid Australia. *Plant Ecology* **190**: 205-217.
- Jeon, S., J. Nam, H. Choi, E. Ju, J. Yoon and J. Kim. 2008. Monitoring for wetlands ecosystem conservation and management in Urban -A case study at Dunchon-dong wetland. *Korean Journal of Nature Conservation* **6**: 127-142.

- Karr, J.R. and E.W. Chu. 2000. Sustaining living rivers. *Hydrobiologia* **422/423**: 1-14.
- Karrenberg, S., P.J. Edwards and J. Kollmann. 2002. The life history of Salicaceae living in the active zone of floodplains. *Freshwater Biology* **47**: 733-748.
- Kim, J.G. and E.J. Ju. 2005. Soil seed banks at three ecological preservation areas in Seoul. *Journal of Ecology and Field Biology* **28**: 271-279.
- Kim, S.J. 2008. A study on the variation of groundwater level in the Han River Estuary (The Effect of the Removing of a Weir). *Journal of Korean Society of Coastal and Ocean Engineers* **20**: 589-601.
- Leck, M.A. and C.F. Leck. 1998. A ten-year seed bank study of old field succession in Central New Jersey. *Journal of the Torrey Botanical Society* **125**: 11-32.
- Liu, G.-H., W. Li, J. Zhou, W.-Z. Liu, D. Yang and A.J. Davy. 2006. How does the propagule bank contribute to cyclic vegetation change in a lakeshore marsh with seasonal drawdown? *Aquatic Botany* **84**: 137-143.
- Looney, P.B. and D.J. Gibson. 1995. The relationship between the soil seed bank and above-ground vegetation of a coastal barrier island. *Journal of Vegetation Science* **6**: 825-836.
- McDonald, A.W., J.P. Bakker and K. Vegelin. 1996. Seed bank classification and its importance for the restoration of species-rich flood-meadows. *Journal of Vegetation Science* **7**: 157-164.
- Niiyama, K. 1990. The role of seed dispersal and seedling traits in colonization and coexistence of *Salix* species in a seasonally flooded habitat. *Ecological Research* **5**: 317-331.
- Nilsson, C., M. Gardfjell and G. Grelsson. 1991. Importance of hydrochory in structuring plant communities along rivers. *Canadian Journal Botany* **69**: 2631-2633.
- Oksanen, J. 2011. Multivariate analysis of ecological communities in R: vegan tutorial. <http://cc.oulu.fi/~jarioksa/opetus/metodi/vegantutor.pdf>.
- Paul, M.J. and J.L. Meyer. 2001. Streams in the urban landscape. *Annual Review of Ecology and Systematics* **32**: 333-365.
- Richardson, D.M., P.M. Holmes, K.J. Esler, S.M. Galatowitsch, J.C. Stromberg, S.P. Kirkman, P. Pyšek and R.J. Hobbs. 2007. Riparian vegetation: degradation, alien plant invasions, and restoration prospects. *Diversity and Distributions* **13**: 126-139.
- Richter, R. and J. Stromberg. 2005. Soil seed banks of two montane riparian areas: implications for restoration. *Biodiversity and Conservation* **14**: 993-1016.
- Roberts, H.A. 1981. Seed banks in soils. *Advances in applied biology* **6**: 1-55.
- SAS Institute. 2009. SAS Software version 9.2. SAS Institute Inc., Cary, NC.
- Sheldrick, B.H. and C. Wang. 1993. Particle size distribution, p. 499-509. *In: Soil Sampling and Methods of Analysis* (Carter, M.R., ed.). Canadian Society of Soil Science, Lewis Publishers, Boca Raton.
- Sims, J.T., R.O. Maguire, A.B. Leytem, K.L. Gartley and M.C. Pautler. 2002. Evaluation of Mehlich 3 as an agri-environmental soil phosphorus test for the Mid-Atlantic United States of America. *Soil Science Society of America Journal* **66**: 2016-2032.
- Tabacchi, E., A.M. Planty-Tabacchi, L. Roques and E. Nadal. 2005. Seed inputs in riparian zones: implications for plant invasion. *River Research and Applications* **21**: 299-313.
- Team, R.D.C. 2008. R: A Language and Environment for Statistical Computing. R Foundation for Statistical Computing, Vienna Austria.
- Ter Heerdt, G.N.J., A. Schutter and J.P. Bakker. 1999. The effect of water supply on seed-bank analysis using the seedling-emergence method. *Functional Ecology* **13**: 428-430.
- Thompson, K. 2000. The functional ecology of soil seed banks, p. 215-236. *In: Seeds: the ecology of regeneration in plant communities* (Fenner, M., ed.). CAB International, London.
- Thompson, K. and J.P. Grime. 1979. Seasonal variation in the seed banks of herbaceous species in ten contrasting habitats. *Journal of Ecology* **67**: 893-921.
- Uchida, T., F. Tazak, J. Maruyama and Y. Sato. 2003. Regeneration responses of *Phragmites australis* (Cav.) Trin. and *P. japonica* Steud. to above-ground organs loss. *Journal of the Japanese Society of Re-vegetation Technology* **29**: 74-79.
- Valk, A.G.v.d. and C.B. Davis. 1978. The role of seed banks in the vegetation dynamics of prairie glacial marshes. *Ecology* **59**: 322-335.
- Van Der Valk, A.G. and C.B. Davis. 1979. A reconstruction of the recent vegetational history of a prairie marsh, Eagle Lake, Iowa, from its seed bank. *Aquatic Botany* **6**: 29-51.
- Ward, J.V., K. Tockner, U. Uehlinger and F. Malard. 2001. Understanding natural patterns and processes in river corridors as the basis for effective river restoration. *Regulated Rivers: Research & Management* **17**: 311-323.
- Woo, H. 2010. Trends in ecological river engineering in Korea. *Journal of Hydro-environment Research* **4**: 269-278.

(Manuscript received 8 February 2011,  
Revision accepted 14 March 2011)









## Appendix 2. Continued.

Life history	Species	Janghang				Kangseo				Banpo		Amsa				
		WF	Pa	Ss	Ms	Pa	Ss	Ms	At	Pa	Ms	WF	Bn	Ss	Pa	Ms
	<i>Erigeron annuus</i>				○	○	○	○	○	○	○	○	○	○	○	○
	<i>Erigeron canadensis</i>				○		○	○	○	○	○		○	○	○	○
	<i>Galinsoga ciliata</i>				○		○		○			○	○			
	<i>Galium spurium</i>		○			○		○								
	<i>Hemistepta lyrata</i>														○	
	<i>Lactuca indica</i> var. <i>laciniata</i>							○								
	<i>Leonurus sibiricus</i>									○	○		○			
	<i>Oenothera odorata</i>						○	○	○	○	○	○	○	○	○	○
	<i>Ranunculus chinensis</i>			○												
	<i>Rorippa islandica</i>	○		○	○	○			○		○		○	○	○	
	<i>Rorippa islandica</i>									○	○	○	○	○		
	<i>Stellaria alsine</i> var. <i>undulata</i>				○	○		○				○		○		
	<i>Stellaria aquatica</i>					○			○			○	○	○	○	
	<i>Trigonotis peduncularis</i>				○	○	○	○					○	○		
	<i>Veronica peregrina</i>				○	○		○				○				○
	<i>Youngia japonica</i>												○	○		
Perennials	<i>Achyranthes japonica</i>							○		○				○		
	<i>Artemisia princeps</i> var. <i>orientalis</i>				○				○						○	
	<i>Artemisia selengensis</i>											○	○	○	○	
	<i>Aster pilosus</i>					○		○								
	<i>Boehmeria nivea</i>														○	
	<i>Calamagrostis epigeios</i>					○	○	○	○							
	<i>Calystegia sepium</i> var. <i>americana</i>									○	○					
	<i>Chrysanthemum boreale</i>														○	
	<i>Cirsium japonicum</i> var. <i>ussuriense</i>					○		○								
	<i>Equisetum arvense</i>				○	○	○		○							
	<i>Geranium koraiense</i>														○	
	<i>Gnaphalium japonicum</i>														○	
	<i>Isachne globosa</i>					○	○									
	<i>Ixeris chinensis</i> var. <i>strigosa</i>							○								○
	<i>Ixeris dentata</i>						○								○	
	<i>Lycopus ramosissimus</i> var. <i>japonicus</i>			○												
	<i>Mentha arvensis</i> var. <i>piperascens</i>		○		○							○	○	○		
	<i>Metaplexis japonica</i>					○			○							
	<i>Miscanthus sacchariflorus</i>		○		○	○		○		○	○		○		○	○
	<i>Phalaris arundinacea</i>					○	○				○	○	○	○	○	
	<i>Phragmites australis</i>		○			○			○	○	○	○	○	○	○	○
	<i>Plantago asiatica</i>													○		
	<i>Poa pratensis</i>				○											
	<i>Poa sphondylodes</i>							○	○	○	○					
	<i>Potentilla paradoxa</i>					○		○				○	○	○	○	
	<i>Robinia pseudo-acacia</i>		○													
	<i>Rorippa globosa</i>											○	○	○		
	<i>Rumex crispus</i>						○		○			○		○	○	
	<i>Salix gracilistyla</i>											○		○		
	<i>Salix koreensis</i>						○							○		
	<i>Salix nipponica</i>				○		○							○		
	<i>Salix pseudo-lasiogyne</i>						○							○		
	<i>Scirpus radicans</i>											○				
	<i>Sium suave</i>			○												
	<i>Spirodela polyrhiza</i>	○														
	<i>Trifolium repens</i>				○											
	<i>Typha angustifolia</i>		○													
	<i>Veronica americana</i>													○		
	<i>Zizania latifolia</i>	○										○				