# Temporal Variation in Species Composition and Abundance of Fish Assemblages in Masan Bay 

Seok Nam K wak* and Sung-Hoi Huh ${ }^{1}$<br>Marine Eco-Technology Institute Co., Ltd., 485-1 Yongdang-dong, Namgu, Busan 608-830, K orea, ${ }^{1}$ Department of Oceanography, Pukyong National University, 599-1 Daeyeon-Dong, Namgu, Busan 608-737, K orea


#### Abstract

A total of 27 fish species were collected by a gill net in Masan Bay. The dominant fish species were Konosirus punctatus, Mugil cephalus, Engraulis japonicus, Psenopsis anomala, Lateol abrax japonicus, Leiognathus nuchalis, Chelon affinis, Trachurus japonicus, and these accounted for $87.6 \%$ of the total numbers. The numerically dominant fish species made up $95.4 \%$ of biomass. These were primarily juvenile species or early larger species. Temporal variation in both species composition and abundance was large: the peak number of fishes occurred in March 2005 and J uly 2005, whilst biomass of fishes was the highest in September 2005 and November 2005. Fish numbers as well as biomass was lowest in November 2005. Temporal changes in the abundance of fishes corresponded with temperature. A gill net can be used as an alternative fishing gear to collect pelagic fishes commercially, although a gill net has a strong selectivity for the target fish species or for size.


Key words: Konosirus punctatus, Mugil cephalus, Engraulis japonicus, Psenopsis anomala, Lateolabrax japonicus, Leiognathus nuchalis, gill net, fish assemblage, Masan Bay

## Introduction

The coastal areas generally support different fish species, a greater diversity and abundance of fish, and larger numbers of juveniles on a worldwide scale. Many fish species use coastal areas as feeding and nursery grounds, including many economically important fishes (J enkins and Wheatley, 1998; Lazzari et al., 1999; Guidetti, 2000; Paperno et al., 2001; Kwak and Klumpp, 2004). Recent studies of fish assemblages on coastal areas in K orea have reported seasonal variation in species composition and abundance of fishes with using small trawl (Lee, 1989; Cha an Park, 1997; Huh and Kwak, 1998a; Lee and

[^0]Gil, 1998; Huh and Chung, 1999; Huh and An, 2000; Oh, 2003), gill net (Kim and Kang, 1991; Hwang et al., 1997; Cha, 1999; Han et al., 2002), and set nets (Lee and Seok, 1984; Go and Shin, 1988; Huh and Kwak, 1998b; Lee, 1998; Hwang, 1998; Hwang et al., 1998; Huh and An, 2002).
Masan Bay have been known one of heavily polluted area in the southern part of Korea. Environmental disturbances such as red tide and pollutant load have occurred every year due to industrial complex around coastal areas since 1980's, however, improving of water quality with construction of sewage treatment plant were in Masan Bay nowadays (Kim, 2003; Oh et al., 2005). To date the study in Masan Bay has been confined to sustainable management for target water quality (J u et al., 2000; Lee and Park, 2003; Oh et al., 2005), and far less is known about the
studies on fish assemblages compared with other regions of K orea except the report was community patterns of demersal fishes by the baited traps and bottom gill nets (Youm, 1997). Few studies have been conducted on fish assemblages with using gill nets. These were included studies on the seasonal variation in species composition and abundances of fishes in the coastal off Shinsudo, Samchonpo (Kim and Kang, 1991), Heunghae (H wang et al., 1997), Geoje Island (Cha, 1999), and Gadeok-do (An and Huh, 2003).

The objective of this study was to examine the temporal variation in species composition and abundance of fishes inhabiting in Masan Bay, K orea and to determine the relationships between environmental factors and fish abundance.

## Methods

Masan Bay was a semi-closed bay and were located several small islands (Fig. 1). Fish samples were collected by drift gill net (multi-filament, 22~150 mm stretched mesh size) with 2~ 3 months intervals from September 2004 to J anuary 2006. Nets were set at 06:00h and fish retrieved at 19:00 h at night.

Specimens were preserved immediately in 10\% formalin after capture and later transferred to $70 \%$ isopropanol. These samples were identified according to Masuda et al. (1984), Y oon (2002), and Kim et al. (2005), and weighed to the nearest gram in wet weight. Specimens were measured to the nearest mm (standard length SL). Water temperature (by thermometer) and salinity (by salinometer) were monitored on each sampling occasion.

The fish data was analysed to obtain the following community variables. Diversity H' (Shannon and Weaver, 1949) was calculated as:

$$
\mathrm{H}^{\prime}=-\Sigma(\mathrm{ni} / \mathrm{N}) \log (\mathrm{ni} / \mathrm{N}),
$$

where n is the number of individuals of each i species in a sample and N is the total number of individuals. Association of fish species, Pianka's similarity index (Pianka, 1973), $\mathrm{A}_{\mathrm{ij}}$ was calculated as:

$$
\mathrm{A}_{\mathrm{ij}}=\left[\Sigma \mathrm{p}_{\mathrm{ih}} \mathrm{p}_{\mathrm{jh}}\right] /\left[\Sigma^{2}{ }_{\mathrm{ih}} \Sigma \mathrm{p}^{2}{ }_{\mathrm{jh}}\right]
$$

where $A_{i j}$ is the similarity of species $j$ on species $\mathrm{i} p_{\mathrm{in}}$ is the proportion of individuals of Ai in a particular month $h \mathrm{p}_{\mathrm{jh}}$ is the proportion of individuals of $A j$ in a particular month $h$. Values for


Fig. 1. Map showing the sampling site (the black area).
the similarity index may vary between 0 , if no similarity occurs, and 1 for complete similarity. The Pianka's similarity index was subjected to an average linkage cluster analysis.
A one-way ANOVA with orthogonal design was used to analyse variations in fish abundance and environmental factors with month. The relationships between fish abundance and environmental factors were analysed using Pearson's correlation coefficient.

## Results

## Temperature and salinity

Temperature at the study site ranged from 9.6 ${ }^{\circ} \mathrm{C}$ to $26.5^{\circ} \mathrm{C}$ and varied significantly with months (one way ANOVA, $\mathrm{F}=16.8, \mathrm{p}<0.05$ ). The peak of temperature was around J uly 2005, a decline in September 2005 and a minimum during winter (J anuary 2006) (Fig. 2). Salinity ranged from $23.4 \%$ to $32.9 \%$ and did not vary significantly between months (one way ANOVA, $\mathrm{F}=4.31$, $\mathrm{p}>$ 0.05 ) with display a similar pattern except in J uly 2005 when it dropped (about 23.4\%d (Fig. 2).

## Fish species composition

A total of 2,346 fish belonging to 27 species


Fig. 2. Temporal variations of temperature and salinity in Masan Bay.
were collected from Masan Bay (Table 1). Numerically dominant fish were Konosirus punctatus(23.1\%), Mugil cephalus(18.9\%), Engraulis japonicus (15.2\%), Psenopsis anomala (7.4\%), Lateolabrax japonicus (6.2\%), Leiognathus nuchalis (6.1\%), Chel on affinis (5.4\%), and Trachurus japonicus (5.3\%), together accounting for $87.6 \%$ of the catch. The numerically dominant fish species made up $95.4 \%$ of biomass because of the presence of large M. cephalus (57.9\% of biomass) which were high in biomass.

## Temporal variation in abundance of fish

The number of fish species ( $13 \sim 22$ species) varied with months (one-way ANOVA, $\mathrm{F}=10.8$, $\mathrm{p}<0.05$ ). Fish species was abundant in September 2004 and May 2005 (Fig. 3-a). Number of individuals varied significantly with months (oneway ANOVA, $\mathrm{F}=13.4$, $\mathrm{p}<0.05$, Fig. 3-b). Fish were abundant March 2005 and J uly 2005 when K. puntatus, M. cephalus, L. japonicus and C. affinis were dominant. Fish numbers was the lowest in J anuary 2006 (Appendix 1). The fish biomass differed substantially between different seasons (oneway ANOVA, $\mathrm{F}=9.77$, $\mathrm{p}<0.05$, Fig . 3 -c). Highest biomass was in September 2005 and November 2005 when many large M. cephalus and K. punctatus were present (Appendix 1). The range of diversity index was $1.69 \sim 22.5$, and higher value was in March 2005 and May 2005 (Fig. 3-d).
Abundance of common fish species varied with months, and were distinct patterns with each fish species (Figs. 4, 5). For example, peak numbers of K. punctatus were in J uly 2005 ( 140 ind.),


Fig. 3. Temporal variations in (A) Number of species, (B) number of individuals (C) biomass, and (D) diversity index of fish species in Masan Bay.
and biomass was peaked in J uly 2005 ( $8,582.0 \mathrm{~g}$ ) and September 2005 ( $9,590.4 \mathrm{~g}$ ) due to many Iarger individuals in these periods. Number of individuals of M. cephal us was highest in J uly 2005 (79 ind.), however, biomass was peaked in September 2005, November 2005, and J anuary 2006. Peak abundances of E. japonicus were in November 2004 and September 2005. Abundances of P. anomala were peaked in September 2004 (117 ind., $8,404.3 \mathrm{~g}$ ), however, few abundances were in other periods. Number of individuals of L. japonicus and L. nuchalis were higher in May 2005 (43 ind., 42 ind.), however, higher biomass was in November 2005 ( $1,411.4 \mathrm{~g}, 293.0 \mathrm{~g}$ ).
The dendrogram shows five clusters which identify the fish species (Fig. 6). The first group was composed of K. punctatus, E. japonicus, M. cephaIus, L. japonicus, L. nuchalis, C. affinis, Repomucenus valenciennei, Hexagrammos otakii, and

Table 1. Total number of individuals and biomass of fish species in Masan Bay

| Species | Total |  |  |  |
| :--- | ---: | ---: | ---: | :---: |
|  | N | $\%$ | W | $\%$ |
| Konosirus punctatus | 542 | 23.0 | $39,387.7$ | 16.7 |
| Mugil cephalus | 443 | 18.8 | $136,696.8$ | 57.9 |
| Engraulis japonicus | 356 | 15.1 | $1,886.0$ | 0.8 |
| Psenopsis anomala | 174 | 7.4 | $10,483.7$ | 4.4 |
| Lateol abrax japonicus | 145 | 6.2 | $10,578.5$ | 4.5 |
| Leiognathus nuchalis | 144 | 6.1 | $1,187.1$ | 0.5 |
| Chelon affinis | 127 | 5.4 | $16,671.6$ | 7.1 |
| Trachurus japonicus | 125 | 5.3 | $8,161.8$ | 3.5 |
| Sardinella zunasi | 63 | 2.7 | $1,106.1$ | 0.5 |
| Ditremma temmincki | 41 | 1.7 | $1,615.4$ | 0.7 |
| Repomucenus valenciennei | 29 | 1.2 | 644.2 | 0.3 |
| Hexagrammos otakii | 28 | 1.2 | $2,427.7$ | 1.0 |
| Acanthogobius flavimanus | 25 | 1.1 | 545.2 | 0.2 |
| Limanda yokohamae | 20 | 0.8 | 970.8 | 0.4 |
| Sebastes inermis | 19 | 0.8 | 318.1 | 0.1 |
| Acanthopagrus schlegel | 14 | 0.6 | $1,703.6$ | 0.7 |
| Thryssa kammalensis | 12 | 0.5 | 295.9 | 0.1 |
| Kareius bicol oratus | 10 | 0.4 | 304.4 | 0.1 |
| Hypodytes rubrippinnis | 8 | 0.3 | 151.1 | 0.1 |
| Sebastes longispinis | 6 | 0.3 | 143.4 | 0.1 |
| Repomucenus lunatus | 5 | 0.2 | 119.5 | 0.1 |
| Platycephalus indicus | 3 | 0.1 | 314.2 | 0.1 |
| Argyrosomus argentatus | 2 | 0.1 | 154.4 | 0.1 |
| Trichiurus lepturus | 1 | 0.0 | 22.6 | 0.0 |
| Clupea pallassi | 1 | 0.0 | 19.5 | 0.0 |
| Sebastes schlegeli | 1 | 0.0 | 31.2 | 0.0 |
| Sillago jaonicus | 1 | 0.0 | 25.4 | 0.0 |
| Oplegnathus fasciatus | 1 | 0.0 | 53.1 | 0.0 |
| Total | 2,346 | 100 | $236,019.0$ | 100 |

N : Number of individuals, W : biomass (g)

Limanda yokohamae with occurring predominantly over study periods. This group can be further divided into three subgroups: subgroup A contains K. punctatus, E. japonicus, and M. cephalus with higher numbers during study periods, subgroup B composed of L. japonicus, L. nuchalis, and $C$. affinis with peak numbers in May 2005, and subgroups C was consisted of R. valenciennei, H . otakii, and L. yokohamae which were highr numbers in March 2005 and May 2005. The second group was composed of Sardinella zunasi, Acanthogobius flavimanus, Sebastes inermis, S. Iongipinnis, and Thryssa kammalensis. This group was high numbers May 2005 and $J$ uly 2005, while number of individuals were few in other periods. The third group was composed of P. anomala, Acanthopagrus schlegeli, R. Iunatus and Platycephalus indicus which were peak numbers in September 2004 and September 2005.

The fourth group was consisted of T. japonicus and Ditremma temmincki with occurrence in J uly 2005. The fifth group was composed of Kareius bicol oratus and Hypodytes rubrippinis with higher numbers in March 2005.

## Discussion

Gill net, the name of the net employed, illustrates the method used to snare target fish. They try to swim through deliberately sized mesh openings but are unable to squeeze through swimming forward. Once in this position, they are prevented from backing out due to the tendency for their gills to become caught. This effectively traps them. These methods were classified by sink gill net (e.g. bottom gill net, and trammel net) and drift gill net. In general, pelagic fishes were collected by drift gill net, whereas demersal fishes were in sink gill net.
A total of 27 fish species was recorded from Masan Bay and Konosirus punctatus, Mugil cephalus, Engraulis japonicus, Psenopsis anomala, Lateolabrax japonicus, Leiognathus nuchalis, Chelon affinis and Trachurus japonicus were numerically dominant. Most of fish species are of commercial and recreational importance. For example Hexagrammos otakii, K. punctatus, M. cephalus, L. japonicus, and C. affinis are va-lued as live fish in the Southern area, Korea, and P. anomala, and T . japonicus harvested as a food fish (Kim and Kang 1993; Y oon, 2002). Compared with studies of fish assemblages in the coastal area off Gadeok-do nearby the study area, noncommerical fishes such as Repomucenus valenciennei, Thrissa kammalensis, L. nuchalis, Zoarces gillis, Sillago japonicus were dominated by small trawl, whereas commercial fish species (e.g. Limanda yokohamae, Cynoglossus abbreviatus, Ditremma temmincki, Hexagrammos oatkii etc.) were abundant with using bottom gill net (An and Huh, 2003). These results indicated that higher numbers of commercial fish species were collected with gill nets. Such conclusions are in general agreement with other studies of fish assemblages in the coastal area, K orea. Hexagrammos otakii, H. agrammus, and S. inermis were common fish species in the coastal water off Shinsudo, Samchunpo, and Heunghae (Kim and Kang, 1991; Hwang et al., 1997), Stephanolepis cirrhifer, D. temmincki and L. yokohamae were abundant around Geoje Island (Cha, 1999).


Fig. 4. Temporal vaiations in number of individuals of common fish species in Masan Bay.

Each dominant fish species exhibited their own distinct seasonal occurrence pattern and different time of peak abundance. Peak numbers of K. punctatus in August 2004, those of E. japonicus were in November 2004 and October 2005, M. cephalus were in August 2005 and J anuary 2006, P. anomala were in September 2004, L. japonicus and L. nuchalis were in April 2005. This results indicate that peak abundance was closely related with rapid increase of larval recruitment. Peak larval recruitment and abundance of one species was separated several months from other
species, with some overlap with another species. After one fish population increased rapidly, it sustained peak abundance for several months and then decreased sharply. Subsequently, another fish population increased and reached peak abundance. Hence the coastal area in Masan Bay were partitioned temporally by dominant fish species in this ways. Other studies have shown similar patterns of variable in fish assemblages of coastal areas, K orea (Lee, 1989; Huh and Kwak, 1997, 1998a, b; Huh and Chung, 1999; An and Huh 2002, 2003).


Fig. 5. Temporal variations in biomass of common fish species in Masan Bay.

Temporal variation in both species composition and abundance of fish species appear to be considerable for fish assemblages. In Masan Bay, number of individuals was highest in May 2005, and J uly 2005 when higher temperature was in temperate area. Several other studies have demonstrated a positive correlation between temperature and fish abundances (Lee and Seok, 1984; Edgar and Shaw, 1995; Huh and Kwak, 1998b; Lazzari et al., 1999; Paperno et al., 2001; Huh and An, 2002; Oh, 2003).

Konosirus punctatus, M. cephalus, and E. japonicus, among common fish species, were comprising $57.2 \%$ of total number of individuals and $75.4 \%$ of total biomass, and this paralleled the
high dominance of L. yokohamae(53.3\% of total numbers, and $42.2 \%$ of total biomass) were collected with bottom gill net in the coastal warer off Gadeok-do(An and Huh, 2003) and also this species were predominated in the study area (Youm, 1997). These results may be explained gear selectivity by sized mesh opening of gill net and then suitable fish size were caught. An and Huh (2003) also demonstrated that size distribution of most common fish species collected by bottom gill net did not varied with months. Drift gill net fishery developed rapidly in the late 2000s off Gyeongsang Namdo including Masan Bay. In particular, Masan Bay was a semi-closed bay, and higher water velocity give it for fisher-


Fig. 6. Dendrogram illustrating the species associations of fishes in Masan Bay.
man to deploy the drift gill net and then the catch of K. punctatus, M. cephalus, and E. japonicus were increasing. These fish species were coincided with common fish species in the study area. Hence we suggested that drift gill net was a reasonable gear for studies of pelagic fishes commercially in the coastal area although the gill net were strong the gear selectivity by mesh size.

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# 마산만 해역에서 서식하는 어류군집의 종조성 및 출현량 변동 곽 석 남*. 허 성 회 ${ }^{1}$ 

(주)해양생태기술연구소 및 ${ }^{1}$ 부경대학교 해양학과

마산만 해역에서 유자망을 이용하여 어류군집의 종조성 및 출현량 변동을 조사하였다. 조사기 간 동안 총 27종이 출현하였으며, 우점종은 전어 (Konosirus punctatus), 숭어 (Mugil cephalus), 멸 치 (Engraulis japonicus), 샛돔 (Psenopsis anomala), 농어 (Lateol abrax japonicus), 주둥치 (Leiognathus nuchalis), 등줄숭어 (Chelon affinis), 전갱 이 (Trachurus japonicus)였는데, 이들 어종 은 전체 개체수의 $87.6 \%$ 와 총 생체량의 $95.4 \%$ 를 차지하였다. 어류군집의 종조성 및 출현량 변동 이 뚜렷하였는데, 출현 개체수는 2005년 3월과 7월, 생체량은 2005년 9월과 11월에 높게 나타났 다. 한편 출현개체수 및 생체량은 모두 2005년 11월에 가장 낮게 나타났다. 수온이 어류군집의 종조성 및 출현량 변동에 가장 큰 영향을 주는 요인이었다. 유자망은 어획대상 어종이나 어획 체 장에 강한 선택성을 가지고 있었지만, 상업성어종의 채집에 효율적이었다.
Appendix 1. Number of individuals and biomass of fish species in Masan Bay from 2004 to 2006

| Species | Sep.-04 |  | Nov.-04 |  | J an.-05 |  | May-05 |  | J ul.-05 |  | Sep.-05 |  | Nov.-05 |  | J an.-06 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | N | W | N | W | N | W | N | W | N | W | N | W | N | W | N | W |
| Konosirus punctatus | 81 | 7,889.4 | 63 | 5,298.3 | 22 | 539.0 | 32 | 969.6 | 140 | 8,582.0 | 96 | 9,590.4 | 94 | 5,940.8 | 14 | 578.2 |
| Mugil cephalus | 24 | 10,727.7 | 10 | 2,842.5 | 52 | 7,691.2 | 73 | 8,509.2 | 79 | 11,330.0 | 48 | 30,441.0 | 73 | 33,340.7 | 84 | 31,814.5 |
| Engraulis japonicus | 28 | 184.8 | 92 | 481.4 | 25 | 140.5 | 37 | 123.4 | 16 | 70.0 | 79 | 490.8 | 39 | 184.0 | 40 | 211.1 |
| Psenopsis anomala | 117 | 8,404.3 |  |  | 14 | 676.1 | 11 | 109.7 | 7 | 290.8 | 9 | 518.6 |  |  | 16 | 484.2 |
| Lateolabrax japonicus | 12 | 1,411.4 | 19 | 3,136.8 | 31 | 2,348.4 | 43 | 1,436.2 | 8 | 829.2 | 6 | 511.1 | 13 | 116.9 | 13 | 788.5 |
| Leiognathus nuchal is | 9 | 110.9 | 27 | 293.1 | 38 | 258.4 | 42 | 268.8 | 15 | 130.3 | 11 | 112.8 |  |  | 2 | 12.8 |
| Chelon affinis | 3 | 581.8 | 4 | 1,421.5 | 42 | 4,577.5 | 49 | 5,357.0 | 21 | 2,370.3 | 2 | 281.7 | 4 | 1,410.4 | 2 | 671.4 |
| Trachurus japonicus | 18 | 1,368.7 | 12 | 695.1 |  |  | 1 | 18.2 | 73 | 5,601.3 | 11 | 282.9 | 9 | 175.5 | 1 | 20.1 |
| Sardinella zunasi | 3 | 54.9 | 2 | 34.2 |  |  | 36 | 523.8 | 7 | 226.3 | 7 | 132.7 | 8 | 134.2 |  |  |
| Ditremma temmincki | 1 | 39.2 | 1 | 39.7 | 3 | 79.5 | 9 | 128.0 | 24 | 1,212.6 | 1 | 35.9 | 2 | 80.5 |  |  |
| Repomucenus valenciennei | 1 | 45.7 | 1 | 42.1 | 5 | 77.0 | 9 | 89.1 | 3 | 70.8 | 1 | 40.8 | 3 | 95.3 | 6 | 183.4 |
| Hexagrammos otakii | 2 | 122.5 | 3 | 324.2 | 8 | 904.6 | 6 | 486.2 | 1 | 62.3 | 3 | 201.5 | 3 | 198.2 | 2 | 128.2 |
| Acanthogobius flavimanus |  |  |  |  | 4 | 84.4 | 12 | 190.4 |  |  |  |  |  |  | 9 | 270.4 |
| Limanda yokohamae | 1 | 19.7 | 1 | 25.8 | 7 | 418.0 | 8 | 427.7 | 2 | 48.9 |  |  | 1 | 30.7 |  |  |
| Sebastes inermis |  |  |  |  | 2 | 39.4 | 15 | 230.6 | 2 | 48.1 |  |  |  |  |  |  |
| Acanthopagrus schlegel i | 9 | 1,135.1 |  |  |  |  |  |  |  |  | 5 | 568.5 |  |  |  |  |
| Thryssa kammalensis | 3 | 68.4 | 2 | 46.8 |  |  | 3 | 69.3 | 4 | 111.4 |  |  |  |  |  |  |
| Kareius bicol oratus | 1 | 17.8 | 1 | 27.5 | 5 | 162.0 | 2 | 63.2 |  |  |  |  | 1 | 33.9 |  |  |
| Hypodytes rubrippinnis |  |  |  |  | 7 | 131.4 | 1 | 19.7 |  |  |  |  |  |  |  |  |
| Sebastes longi spinis | 1 | 17.9 |  |  |  |  | 3 | 72.9 | 2 | 52.6 |  |  |  |  |  |  |
| Repomucenus lunatus |  |  |  |  |  |  | 1 | 22.3 | 1 | 22.7 | 1 | 17.7 | 2 | 56.8 |  |  |
| Platycephalus indicus | 1 | 116.5 |  |  |  |  |  |  |  |  | 2 | 197.7 |  |  |  |  |
| Argyrosomus argentatus | 2 | 154.4 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Trichiurus lepturus | 1 | 22.6 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Clupea pallassi |  |  |  |  | 1 | 19.5 |  |  |  |  |  |  |  |  |  |  |
| Sebastes schlegeli |  |  |  |  |  |  | 1 | 31.2 |  |  |  |  |  |  |  |  |
| Sillago jaonicus |  |  |  |  |  |  | 1 | 25.4 |  |  |  |  |  |  |  |  |
| Oplegnathus fasciatus |  |  |  |  |  |  |  |  |  |  | 1 | 53.1 |  |  |  |  |
| Total | 318 | 32,493.7 | 238 | 14,709.0 | 266 | 18,146.9 | 395 | 19,171.9 | 405 | 31,059.6 | 283 | 43,477.2 | 252 | 41,797.9 | 189 | 35,162.8 |

N : Number of individuals, W : biomass ( g )


[^0]:    *Corresponding author: seoknam@hotmail.com

